# MINISTRY OF ENVIRONMENT AND PARKS PROVINCE OF BRITISH COLUMBIA

# WATER QUALITY ASSESSMENT AND OBJECTIVES FRASER-DELTA AREA BOUNDARY BAY AND ITS TRIBUTARIES

TECHNICAL APPENDIX

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#### 1. INTRODUCTION

#### 1.1 BACKGROUND

The Ministry of Environment and Parks is preparing water quality assessments and objectives for priority waterbodies. This report describes the water quality of the Serpentine, Nicomekl, and Little Campbell Rivers and Boundary Bay, all located within the Fraser-Delta area (Figures 1 and 2). The main purpose of this review was to develop provisional water quality objectives for the Fraser-Delta area for use by the Greater Vancouver Regional District in preparation of a Liquid Waste Management Plan. Presented within this report are data collected to about September 1986. Two additional reports evaluating water quality within the Fraser-Delta area have been prepared. One report dealt with the Fraser River and some tributaries from Hope to Kanaka Creek, while the other dealt with the Fraser River and some of its tributaries from Kanaka Creek to Sturgeon Banks.

The Ministry of Environment and Parks is developing criteria for water quality constituents of concern in British Columbia. Until criteria for a constituent have been approved by the Ministry, working criteria will be used. The criteria mentioned in this report are working criteria, unless noted as being provincial criteria. The provincial criteria, when approved, may be different from the working criteria used in this report.

Some of the data summarized are given with mean values, but often with median values. The reason for this apparent inconsistency is that many data were collected over long time periods during which many detection limits may have existed. The authors have used whichever statistic has more meaning. Median values are always reported for coliforms and pH.

#### 1.2 ORGANIZATION

The sub-basins of the Fraser-Delta area examined in this report are Boundary Bay and its three main tributaries: the Serpentine, Nicomekl, and Little Campbell Rivers.

#### 1.2.1 LITTLE CAMPBELL RIVER

This river, approximately 30 km in length, flows in a westerly direction from Aldergrove Lake Park to Semiahmoo Bay (Figure 5). The 65 km² drainage basin includes four tributaries: Fergus Creek and Sam Hill Creek, which enter the river from the north, and Jacobson Creek and Highland Creek which enter from the south.

#### 1.2.2 SERPENTINE AND NICOMEKL RIVERS

These two rivers share a main valley, a former embayment of the sea, which extends 11 km eastward from Mud Bay to Cloverdale and varies in width from 4 km to  $5 \text{ km}^{(2)}$ .

The Nicomekl River, approximately 34 km in length, extends from Clayton and Langley southwest through the main valley to its mouth in Mud Bay. Anderson and Murray Creeks are the two major tributaries within the  $149 \text{ km}^2$  drainage basin. These creeks are approximately 15 km and 12 km in length, respectively.

The Serpentine River is approximately 35 km in length. Major tributaries of this river are Latimer, Mahood (Bear), and Hyland Creeks which drain a northern extension of the main valley bounded by Clayton and Newton Uplands. The drainage basin area of the Serpentine River is approximately  $116 \, \mathrm{km}^2$ .

#### 1.2.3 BOUNDARY BAY

Boundary Bay, including its northeasterly extension of Mud Bay and its eastern section called Semiahmoo Bay, is located on the south side of the Fraser-Delta area approximately 19 km south from the City of Vancouver. It is 15 km long and 4 km wide, covering an area of 6 090 hectares. The Bay is rectangular in shape and faces southeast onto the southern Strait of Georgia.

#### 2. LITTLE CAMPBELL RIVER

This 65 km² drainage basin, extending from the Langley Uplands westward to Semiahmoo Bay, is primarily an agricultural area with some recreational and urban development. The Little Campbell River has four major tributaries which drain the surrounding highlands (Figures 3a, 5). Fergus Creek flows southerly and is situated east from White Rock. It is 7.2 km long with a drainage area of 9 km². Sam Hill Creek, near Hazelmere, also flows southerly, has a drainage area of 3 km² and is 4 km in length. Both Jacobson Creek and Highland Creek originate in the highlands of northwest Washington State, flow northwesterly to the Little Campbell River, and have similar drainage areas (4.4 km² and 4.8 km², respectively). Jacobson Creek is 1 km longer than Highland Creek (5.5 km and 4.5 km, respectively).

The Little Campbell River watershed was logged early in the century and a lumber mill and log booming ground existed at the river mouth until 1930. The watershed is extensively farmed.

#### 2.1 HYDROLOGY

Two flow measuring sites were located within this drainage basin (Figure 3). Station 08MH059, near White Rock on the Little Campbell River, operated between 1961 and 1964. An annual discharge of 22 500 dam³ was recorded in 1962 and 36 600 dam³ in 1963 (daily discharge ranged from  $0.0~\rm m^3/s$  to  $14.2~\rm m^3/s$  with a mean flow of  $0.94~\rm m^3/s$ ). The lowest recorded flows typically occurred between June and October (mean daily discharge ranged from  $0.06~\rm m^3/s$  to  $0.32~\rm m^3/s$ ), with maximum flows occurring from November to January and in March (daily discharge ranged from  $0.23~\rm m^3/s$  to  $14.2~\rm m^3/s$ ; mean daily discharge ranged from  $1.01~\rm m^3/s$  to  $4.02~\rm m^3/s$ ).

Station 08MH0123, upstream from the confluence of Sam Hill Creek on the Little Campbell River, had mean monthly flows recorded from May to September in 1966. Values ranged from 0.1  $\rm m^3/s$  in August and September, to 0.29  $\rm m^3/s$  in May.

#### 2.2 WATER USES

The issuance of water licences was discontinued in the period from 1959 to 1969 when the river was declared as being fully subscribed (2). Forty-two licenced water withdrawals are located within the Little Campbell drainage basin (Figure 3a). The total annual permitted withdrawal from this system is 1 027 dam³ of which 66.9% is used for irrigation (686 dam³ at 26 sites), 23.6% for stock watering (242 dam³ at 7 sites), 6.8% for industrial use (70 dam³ at 4 sites), 2.4% for land improvement (24 dam³ at 3 sites), and 0.5% for domestic use (5 dam³ at 2 sites). Water withdrawals classed as industrial are used for irrigating golf courses and watering lawns. One of the stock watering licenses relates to a turkey farm. Water withdrawals for land improvements were permitted for pond creation on the Little Campbell River, a fish pond on Ellen Creek, and a fish pond and ladder on Fergus Creek. Swimming occurs in sections of the Little Campbell River near the mouth and adjacent to the Hazelmere Riverside Campground (see Section 2.3.3.2).

Steelhead and cutthroat trout utilize the Little Campbell River system, which ranks fifth or sixth in steelhead angling importance in the Lower Mainland Region.

Spawning of coho salmon in the Little Campbell basin (Figure 5) predominantly occurs between 7 km to 15 km upstream from its mouth and in the lower reaches of Sam Hill Creek, Fergus Creek, and in an unnamed creek 1.2 km upstream from the mouth of the river (42). Scattered spawning also occurs throughout the upper river reaches. Spawning of chum salmon takes place 7 km to 20 km upstream from the mouth of the Little Campbell River. Jacobson and Highland Creeks have habitats which could be suitable for spawning and rearing coho and chum salmon (Figure 5).

The mean numbers of coho and chum returning to spawn (escapement) were calculated, in four-year intervals, for the period from 1962 to 1981. Escapement values for both species of salmon increased during the period from 1962 to 1976 with the largest increase occurring between 1972 and 1976

(Table 1). Mean annual escapement values for chum and coho salmon declined significantly after 1976.

#### 2.3 WASTE DISCHARGES

Six small operations have permits to discharge effluents within this sub-basin. These can be divided into three categories: agricultural operations, gravel washing operations, and residential and recreational facilities. The locations of these waste discharges are referenced on Figure 3a by permit numbers which have been issued pursuant to the Waste Management Act. In addition, there are livestock-raising operations, extensive areas of cropland, and residential developments which contribute diffuse waste loadings to the basin.

The City of White Rock is served by municipal sewers which discharge to the Ladner trunk sewer. Within the drainage basin, there are 12 km<sup>2</sup> served by sewers, or a population of  $54\,000^{(79)}$ . It is projected that this will increase to 99 000 by  $1996^{(80)}$ .

# 2.3.1 AGRICULTURAL OPERATIONS

There are two agricultural operations within this drainage basin under permit. These include a mink food processing plant (Section 2.3.1.1) and a feedlot operation (Section 2.3.1.2).

# 2.3.1.1 S.L. Scheves, Surrey (PE 2096)

This mink-food processing plant is located near the headwaters of an unnamed creek. It was issued a permit in 1975 to discharge effluent associated with processing mink feed. The permit allowed a maximum discharge of 0.91 m $^3$ /d and the following maximum concentrations: 600 mg/L total solids, 55 mg/L suspended solids, and 580 mg/L BOD $_5$ . Odours have been noted as a problem at this site. At present, effluent from this facility is being trucked away to another location. It therefore has no impact on nearby waterbodies.

When the treatment facility consisting of a lined aerated lagoon, septic tank, and seepage pits is installed, a receiving water monitoring program to evaluate the impact of the operation on the creek should be undertaken. It should include two stations (upstream and downstream) in the creek and a third site near the confluence of the creek and the Little Campbell River (to evaluate the impact on water quality within the river). Samples should be collected on a monthly basis for the first year with measurements of flow rates, ammonia, nitrate, nitrite, orthophosphorus, BOD<sub>5</sub>, fecal coliforms, dissolved oxygen, temperature, and pH.

#### 2.3.1.2 Border Feedlot Ltd.

This operation was located adjacent to PE 4256 (a gravel washing operation), near Jacobson Creek, and closed in 1984. The proliferation of filamentous algae and tubifix worms in Jacobson Creek was believed to be attributable to runoff from the feedlot operation. Further discussions related to this operation are included in Section 2.3.2.2 of this report.

Another operation under the same company name (located just downstream from Site 0300065) has been discussed in terms of its effect on the Little Campbell River in a report prepared by Swain and Alexander (3). Data had been collected due to fish kills which were suspected as having resulted from runoff from the operation (3). The data showed a significant increase in values between sites upstream and downstream from the operation on an adjacent creek for total organic carbon, ammonia, nitrate, nitrite, organic nitrogen, BOD<sub>5</sub>, total phosphorus, orthophosphorus, and total fixed solids (3). No conclusions were drawn on what the impact was from this operation (3).

Waste discharges from feeding operations could be reduced through a system of collection ditches discharging to settling/exfiltration ponds. These ditches and settling ponds, with the capacity to contain runoff resulting from the large quantities of precipitation typical to this area, could also provide water for irrigation purposes. This could reduce the

demand on river water for irrigation, and could permit the discharge of runoff at periods when nutrients in the runoff would have little impact on receiving water quality. The practicality of such a scheme would have to be determined.

#### 2.3.1.3 Diffuse Sources

The Ministry of Agriculture and Fisheries, in cooperation with the B.C. Cattleman's Association, conducts the Beef Assurance Program. included in the program, an operator requires a minimum number of cattle Records maintained under this program provide (approximately 20 head). estimates of the average number of cows, calves, yearlings, and finishing cattle at each operation. From these estimates, nutrient loadings were calculated for each drainage basin using nutrient coefficients  $^{(78)}$  for beef (Table 28). These estimates should only be used as indicators of nutrient loadings from diffuse sources owing to the nature of the program (voluntary, minimum size stipulation), variability in quantities and types of cattle at each operation due to market demand, cattle being only a portion of the agricultural sources within the drainage basin, and the uncertainty as to the proportion of nutrients generated that would actually enter surface water. Within this watershed,  $53 \text{ km}^2$  (81.5%) of the land is used for mink, poultry, and dairy farms and cropland. This area has a population of 4 400. These types of operations are sources of nutrients to ground and surface waters and sediments to surface waters. information on which to determine the magnitude of such loadings to the receiving waters is beyond the scope of this report.

Records from the Beef Assurance Program identified three operations within the Little Campbell River watershed (one at the headwaters of Sam Hill Creek, two adjacent to the Little Campbell River). The largest operation is located on Sam Hill Creek with 300 yearlings and 50 finishing cattle. Facilities on the Little Campbell River near White Rock house 133 finishing cattle at the Border Feedlot operation on the Little Campbell River, and 29 cows and 27 calves at a site located in Hazelmere (Figure 3a).

The Border Feedlot operation near Jacobson Creek (Section 2.3.2.2) ceased operation in 1984.

Estimates of phosphorus and nitrogen loadings to the Little Campbell River are in Table 28. It was estimated that 11 060 kg of nitrogen and 860 kg of phosphorus could be generated annually by cattle in the watershed. Assuming 100% transport of these wastes to the river, which is unlikely to occur, nutrient levels in the Little Campbell River could be increased by the following based on varying river flows and arbitrarily chosen release periods.

			Increase (mg/L) With Varying River Flows and Release Periods			
River Flow (m³/s)	4 Weeks	2 Weeks	1 Week			
14.2	0.32	0.64	1.29			
4.02	1.14	2.27	4 <b>.5</b> 5			
1.01	4.53	9.05	18.1			
14.2	0.025	0.05	0.10			
4.02	0.09	0.18	0.35			
1.01	0.35	0.70	- 1.41			
	Flow (m³/s)  14.2  4.02  1.01  14.2  4.02	Flow (m³/s) 4 Weeks  14.2 0.32 4.02 1.14 1.01 4.53 14.2 0.025 4.02 0.09	Flow (m³/s) 4 Weeks 2 Weeks  14.2 0.32 0.64  4.02 1.14 2.27  1.01 4.53 9.05  14.2 0.025 0.05  4.02 0.09 0.18			

These data indicate that phosphorus and nitrogen can be increased significantly at low flows, and even at peak flows when these nutrients likely would be released. Longer release periods and smaller rates of transmission would reduce these impacts. Actual riverine conditions are discussed in Section 2.4.3.

#### 2.3.2 GRAVEL WASHING OPERATIONS

Two gravel washing operations exist in this watershed. They are operated by LaFarge Concrete and Border Sand and Gravel Ltd.

# 2.3.2.1 Lafarge Concrete Ltd. Surrey (PE 3331)

This operation was closed and dismantled in 1984.

This company operated a gravel washing operation about one km south from the Little Campbell River adjacent to Ellen Creek, which is a branch of Highland Creek (Figure 3a). It was issued a permit in 1974 for washwater from a truck wash and gravel washing operation. The quality of effluent discharged from these operations is indicated in Table 2. The gravel washwater was discharged to two settling ponds operated in series, each approximately  $30.5 \text{ m} \times 18.3 \text{ m} \times 4.3 \text{ m}$  deep. The effluent flowed by gravity to one settling pond, then through a pervious dyke to a second pond prior to being recycled. Make-up water was pumped from a deep well on the property.

The truck washwater was discharged to a settling pond approximately 9.1 m x 12.2 m x 4.3 m deep. The settled concrete residues were cleaned out regularly. Permit PE 3331 allowed the discharge of 3  $241 \text{ m}^3/\text{d}$  of truck washwater effluent to the settling pond.

The facility had on occasion discharged directly to Ellen Creek. The permit was amended in early 1980 to allow such a discharge only to occur at the discretion of the Regional Waste Manager.

Direct discharges from an operation of this nature could potentially contain large quantities of coarse and fine solids. However, the exact composition would depend upon volumes discharged and the settling pond from which the discharge took place. The result of such a discharge to Ellen Creek could be deleterious to resident fish populations in terms of scouring the creek bottom, increasing turbidity, or silting over spawning habitats.

# 2.3.2.2 Border Sand and Gravel Ltd. Langley (PE 4256)

This gravel wash and rock crushing operation is located about one km north from the International Border adjacent to Jacobson Creek. It was

issued a permit (PE 4256) in 1976. The permit restricts the maximum flow rate to 2 273  $\rm m^3/d$  and the average flow rate to 1 964  $\rm m^3/d$ .

Effluent from the gravel washing operation flows into a small basin prior to entering an open ditch which discharges into an irregularly shaped settling pond. Clarified water is recycled back to the washing plant. Natural drainage from the pit area is collected and used as make-up water. Recent dyking has reduced the direct discharge to nearby Jacobson Creek.

Border Feedlot (Section 2.3.1.2) was located adjacent to Border Sand and Gravel Ltd. Intermittent discharges from the sand and gravel operation flowed through the feedlot to the creek. Historically, effluent discharged from the pond contained as much as 7 200 mg/L dissolved solids and 2 000 mg/L suspended solids. Remedial works to collect natural runoff at the gravel pit have been installed. Runoff from this operation only occurs during periods of high precipitation or when the system malfunctions. The feedlot operation was closed in 1984.

It is recommended that a one-year monitoring program to assess the impact of the Border Sand and Gravel Ltd. feedlot on Jacobson Creek be undertaken if the feedlot should reopen. Effluent and receiving water monitoring should coincide with precipitation events. Receiving water quality sites adjacent to these operations should be located 100 m upstream Samples should be collected seven times a year (once a month during periods of low flow, and in October, November, and January). In addition, to evaluate the eventual impact on Jacobson Creek and the Little Campbell River, four samples should be collected in June, August, October, and January, at a site on Jacobson Creek and the Little Campbell River 100 m upstream from their confluence, and at Site 0301305 on the Little Campbell River downstream from the confluence. A site near the mouth of Ellen Creek would also be necessary. Field measurements should be made pH, temperature, dissolved oxygen, and flow rate. measurements should include ammonia, nitrate, nitrite, kjeldahl nitrogen, total dissolved phosphorus, total phosphorus, suspended and dissolved solids, fecal coliforms, and organic carbon.

#### 2.3.3 RECREATIONAL AND RESIDENTIAL FACILITIES

# 2.3.3.1 M. Gurtitz Trailer Park, Langley (PE 225, PE 236)

This operation is located about one km north from the Little Campbell River. Permit PE 225, issued for a single dwelling and 32-unit mobile home park in 1968, allows the discharge of 23.2  $\rm m^3/d$  of treated domestic sewage to a tile field after treatment in a 46.4  $\rm m^3$  septic tank. The tile field is in a sandy gravel soil with approximately 15 cm of organic cover. The percolation rate is 0.2 min/cm.

Permit PE 236, issued in 1970 for a 34-unit mobile home park in Langley, allowed the discharge of 23.2 m<sup>3</sup>/d. Treatment and disposal were identical to that cited for the 32-unit mobile home and single dwelling (PE 225).

Effluent from these trailer parks will likely have no measurable impact on the water quality of the Little Campbell River due to the small quantities involved, the fact that the effluents are discharged to soil adsorption systems, and the distance from the river.

# 2.3.3.2 Hazelmere Riverside Campground, Surrey (PE 1558)

This facility is located just south from the Little Campbell River and about six km upstream from White Rock. It has operated under permit since 1972. Sewage is treated in two septic tanks with 24-hour retention, prior to final disposal in two separate tile fields with a total of 1 500 m of tile located 30 m from the river. The quality of the septic tank effluent is indicated in Table 3. Values are typical of those expected from a septic tank. The permit restricts the flow of treated domestic sewage to a maximum 22.7 m<sup>3</sup>/d between May 1 and September 30 of each year.

Three fecal coliform samples were collected in the river upstream at Site 0301305 and at the downstream Site 0301306 between 1977 and 1979. These are listed below:

	Fecal Coliforms Upstream	(MPN/100 mL) Downstream
1977	310	490
1977	330	460
1979	540	240

The data were inconclusive, but may have exceeded working criteria for primary-contact recreation and irrigation of 200/100 mL as a geometric mean and 400/100 mL as a 90th percentile(11,30). Complete water treatment would be necessary prior to use for drinking or stock watering(30). These high upstream values may be attributed to upland activities. Eleven water withdrawals are licenced for this portion of the river, including one for stock watering, and ten for irrigation. Swimming occurs in the river adjacent to the campground.

Two sampling programs are proposed for the Hazelmere Riverside Campground. First, a program is proposed to evaluate the water quality of the Little Campbell River adjacent to this facility during periods of operation (May 1 to September 30). Once a week, a discrete sample from the middle of the river within the recreational (bathing) area should be analyzed for fecal coliforms. The geometric mean for five of these samples should not exceed 200 fecal coliforms/100 mL. Resampling should occur as soon as possible when any sample exceeds 400 fecal coliforms/100 mL. Second, a program to delineate any influence of the tile fields on the water quality in the Little Campbell River is proposed. Upstream and downstream river monitoring should be conducted monthly from May through September during one year. Measurements should be made for pH, dissolved oxygen, temperature, ammonia, nitrate, nitrite, orthophosphorus, and fecal coli-If substantial changes are measured from upstream to downstream, some groundwater monitoring should be undertaken, the exact program to be established at that time.

# 2.3.3.3 Sunnyside Villas Society (PE 5478)

This permit was issued to a senior citizens housing complex four km north from the City of White Rock in 1980 (Figure 3). The complex consists

of a main lodge (24 rooms), twelve cottages, and eight bachelor facilities. Sewage from this operation is treated in an aerobic batch-type secondary treatment plant which discharges to three ground disposal fields. There is a total of 625 m² of tile with a second field of similar area for back-up. A percolation test conducted in the area measured water moving at 7.4 min/cm. The interceptor ditch from these fields enters an open ditch adjacent to 152nd Street which is connected to the White Rock stormwater sewer.

A permit issued pursuant to the Waste Managment Act limits concentrations of  $BOD_5$  (45 mg/L) and total suspended solids (60 mg/L) in the effluent discharged from the treatment plant to the tile field. Single effluent data from this operation for the period 1983-1985 are presented below.

Permit 5478
SUNNYSIDE VILLAS SOCIETY

YEAR	TOTAL SUSPENDED SOLIDS	BODs
1983	8	< 10
1984	31	< 10
1985	30	< 10

These concentrations, the discharge to ground, and the distance between the discharge point and the Little Campbell River indicate that there likely would not be an impact on water quality within the Little Campbell River from this operation.

# 2.3.3.4 Del-Van Construction Ltd. (PE 4645)

This 107-unit rental mobile home park is located upstream from Jacobson Creek. It was issued a permit in 1977 for the discharge of treated domestic

sewage at a maximum rate of 123 m $^3$ /d, and maximum concentrations of 20 mg/L BOD $_5$  and 30 mg/L suspended solids to dual tile fields located 137 m from the Little Campbell River.

The sewage is treated in two batch-type package treatment plants operated in parallel, each with a capacity of  $68.2~\text{m}^3/\text{d}$ . Treated batches are discharged alternately at a rate of  $7.25~\text{m}^3/\text{h}$  to a tank which discharges to an upflow graded-media filter. The filter is cleaned by flushing, after which the wash is returned to the head of the treatment plant.

In order that the tile fields can be dosed, a constant flow from the filter is collected in a second tank for pumping through the fields. This effluent may affect the water quality (phosphorus and nitrogen levels) of the Little Campbell River since it is discharged to a soil adsorption system 137 m from the river.

#### 2.3.3.5 Diffuse Sources

Residential diffuse sources of wastewater include septic tank tile fields and urban stormwater runoff from a small area of White Rock.

In Section 2.3.1.3, it was estimated that 4 400 people live in agricultural areas within the watershed, and thus likely use septic tanks and tile fields for wastewater disposal. Assuming 4 people per household, there are approximately 1100 septic tank tile-field systems in the drainage area. The impact that each system would have on the Little Campbell River would depend upon local soil conditions, the distance to the river, and the quantity of sewage discharged. It is expected that septic tanks and tile fields contribute some loadings of nitrogen, phosphorus, and fecal coliforms, but estimation of these loadings is not possible with the available data. This would require an examination of local soils and land use, as well as groundwater analyses.

Urban stormwater inputs into the Little Campbell River from White Rock would be small in comparison to the total amount of stormwater runoff from

White Rock. Swain<sup>(24)</sup> indicated that urban runoff from residential areas could contribute loadings of fecal coliforms, lead, zinc, and PCBs.

# 2.4 WATER QUALITY AND PROPOSED PROVISIONAL WATER QUALITY OBJECTIVES

An assessment of water quality data collected until 1979 at a site on the Little Campbell River near the mouth (0300065) was conducted as part of a report by Swain and Alexander (3). This, in conjunction with an examination of data collected from 1979 to 1986 at Site 0300065 and data from a site near the headwaters (0300066), provides the basis for the following assessment of the Little Campbell drainage basin. Locations of the water quality sites are shown in Figure 3a, and the data are summarized in Tables 4 and 5.

Present water uses which should be protected in the Little Campbell basin include irrigation, livestock watering, aquatic life, wildlife, and primary-contact recreation.

Maximum values for specific conductivity (57), chloride (53), sodium (55), potassium (75) and sulphate (55) were below the criteria for all water uses, while arsenic values were below a high detection limit (0.25 mg/L) at both sites on the Little Campbell River. These constituents are not discussed further in the following sections.

# 2.4.1 pH AND BUFFERING CAPACTIY

The pH at Site 0300066 was generally more acidic than at Site 0300065 further downstream (Site 0300065 pH range: 6.9 to 8.6, median 7.55; site 0300066 pH range: 6.5 to 7.0, median 6.8). Values for pH at Sites 0300065 and 0300066 generally increased during periods of low flow when temperature and ammonia values were also high. Values for pH at both sites generally were within the criteria ranges for the protection of aquatic life  $(6.5-9.0)^{(31)}$  and for drinking water (6.5 to  $8.5)^{(7)}$ . One sample collected in 1982 at Site 0300065 had a pH value (8.6) which slightly exceeded the criteria for drinking water.

Since the pH values generally meet the more restrictive drinking water criteria, and lower pH values can prevent the formation of toxic concentrations of ammonia, the drinking water criteria are proposed as the provisional objective for pH. Specifically, the objective is that the pH should be in the range of 6.5 to 8.5 at any point in the Little Campbell River basin, except in the initial dilution zones of effluents. If upstream values are outside this range, there should be no change from upstream. These initial dilution zones are defined as occupying less than 50 percent of river width, being no further than 100 m downstream from any discharge point or series of discharge points, and extending from the bed of the river to the surface.

Mean total alkalinity values at both sites were close to 40 mg/L, indicating a moderate buffering capacity of the river to acidic inputs. Low alkalinity values (<25 mg/L) occurred during periods of low flow, high temperatures and pH, and high nitrogen concentrations at Site 0300066. The working water quality criterion of a minimum alkalinity of 20 mg/L for the protection of aquatic life<sup>(4)</sup> was generally met.

#### 2.4.2 HARDNESS AND METALS

The mean total hardness values at both Sites 0300065 and 0300066 were approximately 30 mg/L. Water with hardness values of this magnitude would be classified as soft (<75 mg/L as  $CaCO_3$ ).

#### 2.4.2.1 Metal Values in the Water Column

Metal values were usually lower than the working water quality criteria for the protection of aquatic life, and/or below detection limits. Exceptions to this were values for total aluminum, copper, iron, lead, and manganese. For these metals, higher values were associated with elevated suspended solids.

Total aluminum was measured at both sites on two occasions in 1982. Values at Site 0300066 were less than the working criterion for the protec-

tion of aquatic life of 0.10 mg/L $^{(5)}$ . However, both samples from Site 0300065 contained aluminum concentrations higher than this working criterion. No water quality objective is proposed for aluminum since background values seem high, which cannot be controlled.

Total copper values at Site 0300066 were all less than the detection limit of 0.01 mg/L; however, it cannot be determined if they were less than the working criterion of 0.002 mg/L for the protection of aquatic life<sup>(6)</sup>. Seven samples collected between 1979 and 1982 at Site 0300065 were less than or equal to the working criterion for the protection of aquatic life (three samples contained 0.002 mg/L of total copper, four samples were below the detection limit of 0.001 mg/L). Maximum copper values at Site 0300065 have been as high as 0.023 mg/L; however, the median value was below detectable levels<sup>(3)</sup>. There are no known controllable sources discharging copper directly into this drainage basin. The values at Site 0300065 likely reflect natural conditions, therefore no objective for copper has been proposed. It is recommended that more work be done to determine the sources of copper in the lower reaches of the Little Campbell River.

Seasonal variability of iron can be correlated with flow rate and suspended solids concentrations. Total iron values at Site 0300065 were inversely proportional to flow rates and directly proportional to suspended solid concentrations, while upstream, at Site 0300066, total iron values were inversely proportional to suspended solids concentrations and directly proportional to flow rates. All total iron values measured at both sites were at or above the working criteria for drinking water use (0.3 mg/L)<sup>(7)</sup> for aesthetics, and for the protection of aquatic life (0.3 mg/L to 1.0 mg/L)<sup>(72)</sup>, but less than the criteria for irrigation use (5-20 mg/L)<sup>(6)</sup>. Historical data from Site 0300065<sup>(3)</sup> indicate that total iron values also exceeded these criteria between 1972 and 1979. There is a need to measure both dissolved and total iron to determine the form present.

The mean total lead value at Site 0300065 from 1972 to 1979 was 0.001 mg/L (maximum lead value 0.004 mg/L $^{(3)}$ ). One of seven total lead values (0.011 mg/L), for the period 1979 to 1982 exceeded the working criterion of 0.004 mg/L $^{(84)}$  for the protection of aquatic life, based upon the minimum hardness of 18.3 mg/L. This same sample contained the highest recorded suspended solids (63 mg/L) value, and may have been related to stormwater runoff. Lead values at Site 0300066 were below the detection limit of 0.01 mg/L, which is too high to allow a comparison with the criterion. No objective has been proposed for lead since the one high value appears to be related to high river flow and suspended solids values.

Total manganese values exceeded the drinking water criterion for dissolved manganese of 0.05 mg/L $^{(7)}$  in five of the seven samples taken at both sites (Tables 5 and 6). Levels at Site 0300066 were directly proportional to dissolved and suspended solids concentrations and inversely proportional to flow rates. Values at Site 0300065 were relatively constant over the periods of sampling. There is a need to measure dissolved and total manganese to determine the form present. No objective has been proposed for manganese since high levels appear to reflect background.

#### 2.4.2.2 Metal Values in Sediments

One set of sediment samples collected at four sites (Sites 1, 0300065, 5, and 0300066 - Figure 3a) on the Little Campbell River were analyzed for metals. These data are reported in Table 8, and compared to values for sediments from the Fraser River (21), lakes in British Columbia (22), the Squamish Estuary (23), and from a residential stormwater catchment area in Vancouver (24).

Barium, boron, calcium, chromium, magnesium, molybdenum, selenium, strontium, tellurium, thallium, tin, titanium, and vanadium values for B.C. lakes can be high enough to exceed the values from the Little Campbell River. Sediments from the Fraser River, residential stormwater, and the Squamish Estuary were not analyzed for these metals.

The highest value of copper in sediments from the Little Campbell River (250  $\mu g/g)$  was reported at Site 0300066. This value exceeded all values recorded elsewhere in B.C. (Table 8). Similarly, the cadmium value (3  $\mu g/g)$  at this site exceeded values from four sites in the Fraser River (range 0.1  $\mu g/g$  to 0.3  $\mu g/g)$ . Aluminum, iron, and manganese were relatively high compared to sediments from other areas, except B.C. lakes (Table 8). This is likely related to higher values for these metals in the water column.

The maximum lead (100  $\mu$ g/g) and zinc (270  $\mu$ g/g) values in the Little Campbell River were for sediments from Site 0300065. The lead value for this site was exceeded by values found at the residential stormwater monitoring site and sediments from B.C. lakes (Table 8). Since upstream lead values were 24 and 30  $\mu$ g/g, and stormwater can contain fairly high levels of lead, presumably from automobile emissions, the higher values may be occurring due to stormwater entering the Little Campbell River. The values for zinc at Site 0300065 were only exceeded by values for sediments in B.C. lakes (Table 8).

#### 2.4.3 NUTRIENTS

# 2.4.3.1 Nutrients in the Water Column

The highest nitrogen and phosphorus values typically occurred at Site 0300066 during periods of high flow and low productivity. The highest total ammonia values at Sites 0300065 and 0300066 were for samples collected in February 1979 and 1981. Ammonia, pH, and temperature values at both sites and associated provincial criteria for maximum and average total ammonia for the protection of aquatic life $^{(69)}$  are in Table 6. The highest recorded total ammonia value at either site was 0.39 mg/L, while the lowest maximum criteria value was 1.54 mg/L and the lowest average criteria value was 0.30 mg/L. These data indicate that acute or chronic ammonia toxicity is unlikely to occur at any time throughout the year in the Little Campbell River.

Although there are no apparent ammonia toxicity problems, not very many data have been collected, and there are nonpoint sources of ammonia in the watershed. Therefore, provisional objectives are proposed for average and maximum total ammonia nitrogen values. These are that values should not exceed the maximum and average values outlined in Tables 29 and 30. The objectives apply to the Little Campbell River and tributaries, except in initial dilution zones of effluents, described in Section 2.4.1. The average value is to be calculated from a minimum of five samples collected in a 30-day period.

The maximum recorded nitrite values were 0.024 mg/L at Site 0300065 and at Site 0300066. These values were less than the provincial water quality criterion for nitrite of 0.06 mg/L as an average value (8) for chloride levels greater than 4 mg/L. Nitrite can be toxic to aquatic life, and can be formed when ammonia is oxidized. The following are proposed as objectives for nitrite to protect aquatic life.

Chloride Concentration (mg/L)	Maximum Nitrite Concentration (mg/L-N)	30-d Average Nitrite Concentration (mg/L-N)
<2	0.06	0.02
2-4	0.12	0.04
4-6	0.18	0.06
6-8	0.24	0.08
8-10	0.30	0.10
>10	0.60	0.20

The objectives apply to the Little Campbell River, Fergus Creek, Sam Hill Creek, Highland Creek, Jacobson Creek, and two unnamed creeks represented in Figure 5, except in initial dilution zones of effluents described in Section 2.4.1. The average value is to be calculated from a minimum of five samples collected over a 30-day period.

Total dissolved phosphorus values were similar at Site 0300066, near the headwaters, and at Site 0300065 near the mouth of the Little Campbell River (Tables 4 and 5). Values for total dissolved phosphorus increased during periods (November to March) of high flow and low productivity at Site 0300065 and, to a lesser extent, at Site 0300066.

Biologically available nitrogen concentrations (sum of nitrate/nitrite plus ammonia) were high during periods of low productivity and high flow, while lower values were associated with periods of high productivity and low flow. Both biologically available nitrogen and total dissolved phosphorus values (Table 7) were high enough to sustain a heavy algal growth provided that such physical factors as light, turbidity, turbulence, substrate, and stream velocity were suitable.

Nitrogen:phosphorus ratios (Table 7) were calculated to determine the nutrient which limits algal growth in the Little Campbell River. At Site 0300066, nitrogen appeared to be the limiting factor (N:P<5)in samples collected in September. Phosphorus was the limiting factor (N:P>15) for all samples collected at Site 0300065 and in samples collected at Site 0300066 during February, May, and November (29).

The provincial water quality criteria for nutrients and algae in streams limits chlorophyll- $\underline{a}$  levels to 50 mg/m² in recreational areas and 100 mg/m² for the protection of aquatic life<sup>(9)</sup>. No chlorophyll- $\underline{a}$  data for the Little Campbell River were available. It is recommended that chlorophyll- $\underline{a}$  data be collected on a routine basis to determine if these criteria are met in the Little Campbell River basin. However, to protect recreational use,  $\underline{a}$  provisional objective is proposed. The objective is that chlorophyll- $\underline{a}$  levels in periphyton collected on natural substrate should not exceed 50 mg/m². The objective applies along the length of the Little Campbell River, except in the initial dilution zones of effluents, described in Section 2.4.1. It is recognized that this proposed objective is likely not met.

#### 2.4.3.2 Nutrient Values in Sediments

One set of sediment samples from four stations on the Little Campbell River was analyzed for organic carbon, nitrogen, phosphorus,  $BOD_5$ , and percentage of total volatile residue (organic content) (Table 8). The location of the sites in Table 8 are shown in Figure 3a. The highest values for organic carbon (23.0 mg/g), nitrogen (2.8 mg/g),  $BOD_5$  (1.97 mg/g), and percentage of total volatile residue (organic content) (11.1%) were recorded at Site 0300066 near the headwaters. Phosphorus values increased slightly from Site 0300065 near the mouth (0.468 mg/g) upstream to Site 0300066 (0.699 mg/g).

Serpentine and Nicomekl River sediment data in Tables 11(b) and 20, respectively. Kjeldahl nitrogen values from the Little Campbell River were higher than those from the Nicomekl River, but lower than those from the Serpentine River. Total phosphorus values from the Little Campbell River were within the ranges reported for the Serpentine and Nicomekl Rivers. Organic carbon values from the Little Campbell River were similar to those reported for the Nicomekl River.

Sediment data from the Little Campbell River then were compared to data from Sturgeon Bank, adjacent to the Iona sewage treatment facility (21), B.C. lakes (22), and the Squamish Estuary (23). Organic carbon values for the Little Campbell River were lower than those from Cattermole Creek in the Squamish Estuary (23) or some lake sediments (22). Total phosphorus and Kjeldahl nitrogen values for the Little Campbell River were similar to values near the Iona sewage treatment facility, but lower than those from some lake sediments (22) (Table 8). The percentage of organic material in some lake sediments (22) was substantially greater than in sediments from the Little Campbell River (Table 8).

Nutrient values for sediments collected near the headwaters of the Little Campbell River were similar to sediments adjacent to Iona Sewage

Treatment facility. This suggests that nutrients may be sorbing onto sediments near the headwaters.

# 2.4.4 OXYGEN AND OXYGEN-CONSUMING MATERIALS

The lower reaches of the Little Campbell River are used as a salmonid migratory route between August and November. During this period at Site 0300065, dissolved oxygen values ranged from 8.1 to 14.2 mg/L (mean: 9.8 mg/L), while percent saturation values ranged from 78.7 to 139.2% (mean: 92.1%). All recorded dissolved oxygen values during this period at Site 0300065 were above 8.0 mg/L. Dissolved oxygen values in the upper reaches at Site 0300066 between August and November ranged from 2.3 to 6.5 mg/L (mean: 4.5 mg/L). Percent saturation values at Site 0300066 during this period ranged from 23 to 68.7% (mean: 44.1%). Some fish kills in the area of the Little Campbell River near Site 0300066 have been attributed to low dissolved oxygen levels (40).

Spawning can occur in the river and its tributaries from November to May. Between November and May, dissolved oxygen concentrations ranged from 4.2 to 13.1 mg/L (mean: 8.8 mg/L) at Site 0300066, and from 9.5 to 13.5 mg/L (mean: 11.4 mg/L) at Site 0300065. Criteria developed for fish hatcheries are a minimum 8.0 mg/L dissolved oxygen during the period between egg fertilization and "eyed" egg stage, and concentrations between 8.5 and 12 mg/L (temperature dependent) during the period from "eyed" egg stage to These criteria are not directly applicable to streams hatch (61). because in hatcheries eggs are held in trays in the water column, whereas in streams the eggs are in the gravels and there is a differential of  $\approx 3$  mg/L between the water column and the interstitial water in the gravels. A water quality criteria document prepared by the U.S. EPA indicated that no impairment to salmonids occurs at 8.0 mg/L for most life stages or at 11.0 mg/L when embryo larvae are present (83). Slight impairment occurs at levels of 6.0 mg/L and 9.0 mg/L, respectively (83).

At Site 0300066, 7 of 16 measurements, when paired with corresponding temperature data, were less than the criteria which are temperature dependent  $^{(61)}$ . Since percent saturation values ranged from 45.4 to 103.2% (mean: 70.1%), it is possible in many situations that dissolved oxygen concentrations should be raised to protect aquatic life. At Site 0300065, 14 of 21 samples, when paired with temperature data, met the criteria  $^{(61)}$  which vary with temperature. Percent saturation values at Site 0300065 during this period ranged from 79.5 to 112.3% (mean: 97.4%). These data indicate that there is not the same need nor opportunity to improve dissolved oxygen concentration at Site 0300065.

It is apparent that there are periods when there are real concerns with low dissolved oxygen levels in the river. Any proposed objective must recognize this situation, and encourage improvements to be made in the watershed which will be reflected in higher dissolved oxygen concentrations. Therefore, a staged provisional objective is proposed for the minimum dissolved oxygen concentration. Initially, the minimum dissolved oxygen concentration should not be less than 6.0 mg/L in the period from June through October recognizing that there will be a slight impairment to salmonids. Once this minimum is consistently met, the objective for a minimum value should be increased to 8.0 mg/L to provide for no impairment of mature salmonids. This latter objective has been met in the Lower Little Campbell River in the past. In addition, the minimum dissolved oxygen concentration should not be less than 11.0 mg/L when fish larvae and/or alevin are present. These objectives apply to in-situ measurements or discrete samples taken outside the initial dilution zones of effluents. These excluded initial dilution zones in streams are defined as extending up to 100 m downstream from a discharge (and up to 100 m upstream at times of tide reversal), and occupying no more than 50% of the stream width around a discharge point, from the bed of the stream to the surface.

There are no permitted point-source discharges within the Little Campbell drainage basin which impact dissolved oxygen values. Remedial activities would therefore relate to non-point sources which may decrease

dissolved oxygen. These can include agricultural discharges through drainage ditches.

### 2.4.5 BACTERIOLOGICAL QUALITY

Fecal coliform data (n=9) from Site 0300066 met the working criteria for recreational and irrigation use  $(200\text{--}400 \text{ MPN/}100 \text{ mL})^{(30)}$  on all but one occasion (5400 MPN/100 mL, September 1980), while four of nine samples from Site 0300065 met the criterion. Drinking water or livestock water (confined feeding) would require complete water treatment (coagulation-flocculation, sedimentation, filtration and disinfection) prior to use (50); however, the B.C. Ministry of Health does not approve of the use of these waters as a drinking water supply.

It is proposed that provisional objectives be established for the Little Campbell River and its tributaries. The proposed objectives are that the geometric mean fecal coliform value should be ≤200 MPN/100 mL from April through October. The 90th percentile value should be ≤400 MPN/100 mL from April through October. These objectives apply to samples taken anywhere in the Little Campbell River basin outside the initial dilution zones of effluents, described in Section 2.4.1. Mean values and 90th percentile values should be based upon at least five samples collected in a 30-day period. Drinking water licence holders should be advised that, at present, complete water treatment may be necessary for Little Campbell River water prior to using for drinking or livestock watering purposes. Further monitoring of fecal coliforms and fecal streptococci should be conducted to determine present levels and sources of fecal contamination within the basin.

#### 2.4.6 SOLIDS

Suspended solids data (ten samples collected from 1979 to 1983) from Site 0300066 ranged from 2 mg/L to 22 mg/L and from 4 mg/L to 63 mg/L at Site 0300065. No seasonal trends were observed at either site over the

sampling period. Values at Site 0300066 are such as to provide an excellent level of protection (<25 mg/L) to fisheries<sup>(13)</sup>, while those at Site 0300065 would provide a good to moderate level of protection (25-80 mg/L)<sup>(13)</sup>.

The provincial water quality criteria for particulate matter are that induced non-filterable residue should not exceed 10 mg/L when background non-filterable residue is  $\leq 100$  mg/L and should not be more than 10% of background when background is  $\geq 100$  mg/L  $(1^{14})$ . An objective is therefore proposed that induced non-filterable residue should not exceed 10 mg/L if background levels are  $\leq 100$  mg/L and should not be more than 10% of background if background levels are  $\geq 100$  mg/L in the Little Campbell River or its tributaries. This objective applies to discrete samples taken outside the initial dilution zones of effluents as described in Section 2.4.1.

In addition, to protect salmonid spawning areas, there should be no significant (95% confidence level) induced benthic sedimentation on the basis of accumulation by weight for particles <3 mm.

The provincial criteria (14) for turbidity to protect aquatic life and for drinking waters with some form of treatment is that induced turbidity should not exceed 5 NTU when background is less than 50 NTU. When background is greater than 50 NTU, the induced turbidity values should not be more than 10% of background. For irrigation, the criterion is that induced turbidity should not exceed 10 NTU when background is less than 50 NTU, and should not be more than 20% of background when background turbidity is greater than 50 NTU. An objective is therefore proposed that induced turbidity should not exceed 5 NTU if background levels are ≤50 NTU and should not be more than 10% of background if background levels are >50 NTU in the Little Campbell River or its tributaries. The objectives apply to discrete samples taken outside the initial dilution zones of effluents as described in Section 2.4.1.

Dissolved solids ranged from 76 mg/L to 142 mg/L at Site 0300065 and from 66 mg/L to 108 mg/L at Site 0300066. Dissolved solids levels were substantially lower than the most sensitive criterion of 500 mg/L for drinking water (52) and irrigation (4).

### 2.5 RECOMMENDED MONITORING

The quantity of data (maximum number of ten samples analyzed in an eight-year period) and the seasonal component of proposed water quality objectives predicate the need for an intensive monitoring program at Sites 0300065 and 0300066. Water samples and field measurements should be collected five times in a 30-day period between June and September, and five times in a 30-day period between November and January. Field measurements should be made for dissolved oxygen, pH, specific conductivity, and temperature. Laboratory measurements, as a minimum, should be made for ammonia, nitrate, nitrite, periphyton chlorophyll-a, dissolved orthophosphorus, total phosphorus, fecal coliforms, suspended solids, dissolved solids, turbidity, and dissolved and total copper, iron, manganese, and lead. As well, further monitoring of fecal coliforms and fecal streptococci should be conducted to determine present levels and sources of contamination within the basin.

Border Sand and Gravel Ltd. (PE 4256) has intermittently discharged to Ellen Creek and Jacobson Creek, while Hazelmere Riverside Campground, owing to its proximity to Little Campbell River, may contribute to the elevated fecal coliform values in the river. Monitoring programs for these two operations have been outlined in sections 2.3.2.2. and 2.3.3.2.

To assess the impact of upland urban development in the Little Campbell drainage basin on water quality, a one-year monitoring program is recommended. This study should identify the number of stormwater sewers from White Rock and the quantity and quality of discharge into the drainage basin. Field and laboratory measurements should be taken at at least one stormwater outfall, a site on Fergus Creek and sites on the Little Campbell

river upstream and downstream from Fergus Creek. Field measurements should include flow rates, dissolved oxygen, pH, specific conductivity, and temperature. Laboratory analyses should, as a minimum, be made for ammonia, nitrate, nitrite, total dissolved phosphorus, fecal coliforms, fecal streptococci, dissolved and suspended solids, total alkalinity, hardness, total and dissolved copper, total and dissolved lead, and total and dissolved zinc. Swain<sup>(24)</sup> has recommended that rainfall and dustfall monitoring be conducted in conjunction with this type of program. Ambient air (dustfall) sampling should be conducted monthly while water and rainfall sampling should be conducted on a precipitation event basis.

Actual monitoring undertaken will depend upon available resources and regional priorities.

### 2.6 CONCLUSIONS

The Little Campbell River is of good quality in terms of low metal and ammonia nitrogen values. It has moderate to low buffering capacity to acidic discharges and is usually classified as soft. To prevent the fouling of plumbing systems and for the protection of sensitive aquatic life a provisional water quality objective has been proposed for pH. Nitrate nitrogen, ammonia-nitrogen, and phosphorus concentrations decreased during periods of low flow and high temperature. Low flows and high temperatures also contributed to low dissolved oxygen values.

Bacteriological quality, in terms of fecal coliform concentrations, was generally poor throughout the system. An objective has therefore been proposed related to fecal coliform values during the irrigation and recreation seasons. Dissolved oxygen values were low with some seasonal variability near the headwaters. Provisional objectives have been proposed for dissolved oxygen.

### 3. SERPENTINE RIVER

This drainage basin contains three tributaries. These are Latimer Creek, Hyland Creek, and Mahood (Bear) Creek (Figures 3 and 5). Latimer Creek is 6 km long, has a drainage area of  $14.5~\rm km^2$  within the Clayton upland, and extends eastward to 2 km past the farming community of Port Kells. The largest and most industrialized drainage area is Mahood (Bear) Creek (38 km²) within the Newton upland. This upland also contains the 8.8 km long Hyland Creek (12.6 km²) which extends west from the Serpentine River to Newton.

The surficial material in these tributaries is comprised of stony marine clays, till and till-like mixtures situated at or near the surface. The relatively impervious nature of these soils results in much of the precipitation entering the system as direct runoff. The Serpentine drainage basin (132 km $^2$ ) is comprised of an upland component (Clayton and Newton Uplands) and a lowland component (22 km of ditches and dyked channels through predominantly agricultural land).

In the Serpentine fen area, there is a triangular drainage pattern (near Site 1100049). The one straight stretch is a man-made ditch.

### 3.1 HYDROLOGY

The locations of flow gauge stations within the Serpentine River drainage basin are on Figure 3. Sites 08MH018 and 08MH020 were operated on Mahood (Bear) Creek from 1926 to 1928, and from 1959 to 1984, while Site 08MH060 operated intermittently from 1961 to 1966 on the Serpentine River near Port Kells (upstream from Mahood Creek). All three stations are located in the headwaters of the Serpentine basin. Two-year and ten-year, seven-day average low flow rates were calculated to be 0.0145 and 0.0025 m³/s at Site 08MH018 and 0.073 and 0.0175 m³/s at Site 08MH020. Based on one complete year of data and two partial years, the two-year return flow on the Serpentine River would be 0.057 m³/s at Site 08MH060.

Periods of low flow were typically between June and September while maximum flows typically occurred between November and January. Maximum flows on Mahood Creek were 25 m³/s at 08MH018 and 28.3 m³/s at 08MH020. On the Serpentine River, the maximum flow at Site 08MH060 was 5.92 m³/s recorded in 1965 when infrequent measurements were taken.

Ten flow measurements taken at Site C on Latimer Creek (Figure 3) in the summer of 1977 ranged from 0.01 to 0.06 m $^3$ /s. A large percentage of this flow was attributed to groundwater contributions (28). On Hyland Creek, seven flow measurements were taken at a site three kilometres upstream from the Serpentine River, between June and September 1977. Flows ranged from 0.06 to 0.30 m $^3$ /s (mean: 0.09 m $^3$ /s; median: 0.06 m $^3$ /s).

### 3.2 WATER USES

Irrigation is the major use of water withdrawn from this drainage basin (Figure 3). Licenced withdrawals from the lowland portion of the Serpentine River account for 80% (1 674 dam³) of the total annual volume used (2 085 dam³) for irrigation. Two licences permit 1.14 m³/d of water to be withdrawn for stock watering, one licence permits 4.65 m³/d of water to be withdrawn for drinking water, and 1 233 dam³ of water is used to supplement the waterfowl refuge operated by the Wildlife Branch. Two licences permit 176 dam³ to be withdrawn for industrial use (11.36 m³/d by a greenhouse-operation and 171 dam³ by a cranberry farm).

Hyland Creek is licenced for water withdrawals of 98.7 dam<sup>3</sup> for irrigation use, and 45.5 m<sup>3</sup>/d for industrial use (brickyard). Similar uses are made of waters from Latimer Creek: 23.4 dam<sup>3</sup> being used for irrigation, 2.3 m<sup>3</sup>/d for stock watering, and 431.7 dam<sup>3</sup> for industrial use (forest nursery) from Latimer Pond. In Mahood (Bear) Creek, 75.9 dam<sup>3</sup> of water is used for irrigation, and 6.2 dam<sup>3</sup> for industrial use (golf course).

Steelhead and cutthroat trout utilize the Serpentine River system. Locations of spawning and rearing areas within the Serpentine River basin are on Figure 5. Coho salmon spawn in the mainstem of the Serpentine River

near the headwaters upstream from Port Kells, as well as in the headwaters of Hyland Creek, in Mahood (Bear) Creek, and in Latimer Creek.

The coho salmon population has fluctuated considerably over the last three decades (Table 9). The lowest recorded values for individuals spawning (75) occurred in the Serpentine River in 1969. The highest values occurred in 1971 and 1976, when 3 500 individuals spawned. Generally, values in the 1950's ranged from 400 to 1 500 (mean 1 165). This decreased through the 1960's (range 75 to 750; mean 363). In the 1970's the numbers of coho spawning in the Serpentine River increased to a level higher than those found in the two previous decades (range 750 to 3 500; mean 2 125). Numbers in the early 1980's decreased to a mean of 638 (range: 350 to 1100). Spawning of coho salmon begins in late October, reaches its peak in mid-November, and is finished by late December.

Recreation is limited to non-contact activities (e.g., boating), usually from the mouth of the Serpentine River upstream to the floodgates, although a minimal amount occurs upstream from the floodgates.

### 3.3 WASTE DISCHARGES

The location of operations issued permits to discharge effluents within the Serpentine drainage basin pursuant to the Waste Management Act are represented by their permit number in Figure 3. No permitted effluent is discharged directly to the Serpentine River; however, cooling water, stormwater, and effluent from an iron foundry are discharged to Mahood (Bear) Creek. Treated stormwater runoff from a concrete ready-mix operation is discharged to Hyland Creek. The remaining operations discharge effluent to soil disposal systems. Diffuse sources of waste include agriculture and urban stormwater.

# 3.3.1 SMITH TRAILER PARK LTD. LANGLEY (PE 209)

This operation is located 61 m north of Latimer Creek, about three km south from Barnston Island (Fraser River) and more than five km to the east

from the Serpentine River. It is allowed under permit PE 209, issued in 1967, to discharge an average of 43.19 m³/d of domestic sewage treated in a septic tank to two tile fields, each containing 610 m of tile. The system is designed to service 50 independent and 20 dependent trailer spaces (to be completed in three stages). The area has a low groundwater level and is characterized by sandy soil resulting in high percolation rates. The location of the tile field should result in little impact on water quality of the Serpentine River, although some impact may occur on Latimer Creek.

### 3.3.2 FERRO ENAMELS, SURREY (PE 2208)

Ferro Enamels is a paint manufacturing plant located near the headwaters of Mahood (Bear) Creek. Permit PE 2208, issued in 1973, allows the discharge of cooling water to Mahood (Bear) Creek. The average permitted discharge is 1.64 m³/d, while the maximum is 2.18 m³/d. There is a minimum dilution ratio of about 575:1 at the maximum permitted discharge rate and at the two-year, seven-day low flow. Since the permitted discharge rate is so small, it is well diluted whenever there is any flow in the creek. The maximum permitted temperature in the effluent is 27°C.

One analysis of the cooling water in 1978 indicated that it contained an oil and grease value of 2 mg/L, and a suspended solids value of 2.6 mg/L.

The small quantity of this discharge would likely have a negligible impact on the water quality of Mahood (Bear) Creek.

### 3.3.3 B.C. HYDRO AND POWER AUTHORITY, SURREY (PE 6026)

B.C. Hydro operates a repair shop and transformer storage area down-stream from Ferro Enamels on Mahood (Bear) Creek. This permit allows for a maximum daily stormwater discharge of 1 250 m³ containing 5 mg/L of oil and grease from this facility. Discharge is via a ditch to Mahood (Bear) Creek at 128th Street in Surrey. The electrical shop carries out routine maintenance and overhauling of powerline transformers and capacitors. Stormwater runoff which may contain oil, is treated by two gravity oil separators with

a combined capacity of 1 154  $m^3$ . Such a system under average precipitation conditions will provide a minimum of 24 hours retention to remove the oil.

Twelve samples were collected between 1981 and 1984 at the separator. The dilution ratios for the effluent in Mahood Creek ranged from 1.7:1 to 37:1, based on effluent flows from  $34~\rm m^3/d$  to  $729~\rm m^3/d$  and two-year low flows at station 08MH018. However, the flow in Mahood Creek at 08MH018 is probably twice that near the waste discharges. Thus the dilution ratios are likely closer to 0.8:1 and 18:1, respectively.

Eight samples taken between 1982 and 1984 were analyzed for PCBs. These samples were collected from the separator, at two sites within the stormwater system, and in the open ditch near the confluence with Mahood (Bear) Creek. PCB values from four samples collected at the separator ranged from below detectable levels (0.0004 mg/L) to 0.012 mg/L with a mean of 0.0036 mg/L. One sample was collected from the ditch site in 1981 with a PCB value of 0.006 mg/L when the effluent flow rate was 552 m³/d.

Using the PCB concentration of 0.006 mg/L found in the ditch and a minimum dilution ratio of 1.7:1, PCBs in Mahood Creek are calculated to be increased by 0.0035 mg/L due to this discharge, assuming complete mixing. This would exceed working water quality criteria for drinking water (3  $\mu$ g/L) (75) and the protection of aquatic life (0.001  $\mu$ g/L)(76). It is unlikely that low flows would coincide with precipitation events, and thus the drinking water criterion was probably not exceeded. However, dilutions of <400 to 12 000 would be needed to meet the aquatic life criterion which was probably exceeded. Provisional water quality objectives for PCBs are proposed in Section 3.4.7.

Further ambient monitoring for PCBs is recommended in Mahood Creek upstream from this discharge, in the ditch immediately downstream from this discharge, and 100 m downstream from the point of discharge. Samples of water, fish, and sediments should be collected to determine if objectives are also required for sediments and fish. The frequency should be based upon precipitation events and analytical results.

### 3.3.4 ASSOCIATED FOUNDRY LTD. SURREY (PE 1529)

This operation, located near the headwaters of Mahood (Bear) Creek, was described earlier by Swain<sup>(17)</sup>. Scrap iron is melted at this facility to produce plumbing pipe. Wastewater treatment consists of an oil separator and a series of catch basins. The effluent is discharged to Mahood (Bear) Creek.

Swain reported that the permit levels, equivalent to level 'C' pollution control objectives for metal finishing plants and industries discharging heavy metals, were not met for flow rate, pH values, and suspended solids (17). Suspended solids were in compliance approximately 50 percent of the time. High values for oil and grease were also reported. Data in Table 10, collected between 1978 and 1982, indicate that oil and grease, pH, and suspended solids still exceeded permitted levels in more than 50 percent of the samples taken. The flow rate met levels allowed in permit PE 1529 (pH range 6.5-8.5, flow rate average 90 m³/d). The minimum dilution ratio available for the effluent, using the two-year and ten-year, seven-day average low flow rates for Site 08MH018, would be approximately 14:1 and 2.4:1, respectively or 7:1 and 1.2:1 if appropriate flows at the point of discharge are considered.

This operation has the potential to increase temperature, oil and grease, suspended solids, and iron values to some degree in the creek. The estimated minimum dilution is not likely to provide an adequate safety factor to prevent considerable impacts being evident in the creek.

### 3.3.5 RELIANCE FOUNDRY CO. LTD. SURREY (PE 2549; PR 2550)

The foundry is located at the headwaters of Mahood (Bear) Creek. It was issued a permit (PE 2549) allowing the discharge of 115 m³/d of once-through cooling water at a temperature of 24° C. The discharge is to a ditch which carries the cooling water southeast for 200 m then into another ditch flowing northwest to Mahood (Bear) Creek. The minimum dilution avail-

able, assuming the maximum daily discharge rate and the ten-year and two-year, seven-day low flows for Site 08MH018, would be approximately 1.9:1 and 10.9:1, respectively or about 0.9:1 and 5:1 if approximate flows at the point of discharge are considered. Swain reported that effluent data indicated that the permit limits were being met (17). Samples collected upstream and downstream from the foundry in February 1980 showed that the pH of the creek decreased from 6.8 to 6.5, while the specific conductivity increased from 66 to 193 uS/cm. It is recommended that more data related to the impact of this discharge on the creek be collected.

The company also operates an industrial landfill adjacent to the foundry. Permit PR 2550, issued in 1973, allows the discharge of an average  $0.03~\text{m}^3/\text{d}$  (averaged over a five-day week) of refuse. The refuse disposed of in the landfill consists of moulding sand (85%), sand slag (10%), and broken refractory (10%).

The landfill is located 3 m from a drainage ditch. The water table has been recorded approximately 0.3 m below the bottom of the site and the typical soil profile is Newton stoney clay (to a depth of 3 m). The service life of this landfill was thought to be five years (in 1973). Owing to the nature of the refuse being discharged, no covering is required. As well, it is expected that any impact on the water quality of Mahood (Bear) Creek would be minimal.

Periodic inspections of the site by the Regional Waste Management Branch with bi-annual monitoring of the nearby ditch for iron content was proposed for this permit. No data have been recorded on the data storage system operated by the Ministry of Environment and Parks.

# 3.3.6 GOODBRAND CONSTRUCTION LTD. SURREY (PE 6050)

This ready-mix concrete truck washing operation is located on Hyland Creek approximately 1 km west from the Serpentine River. The operation was

issued permit PE 6050 in 1982, allowing the discharge of 29 m³/d of cement washwater into settling/recycle ponds (127 m³ capacity).

A maximum of 240 m³/d overflow from a settling pond (24-hour retention) handling yard drainage and some truck washdown water is permitted to be discharged to Hyland Creek. The minimum dilution available for such discharges is not known. No effluent data were available for this site. It is not known if the discharged wastewater has a high pH when it is discharged. It is recommended that samples of the effluent, as well as creek samples both upstream and downstream from where the effluent enters the Creek, be taken and analyzed for pH, dissolved and suspended solids, turbidity, and hydrocarbons.

#### 3.3.7 ALPULKA INVESTMENTS LTD. (PE 5276)

This operation is located just east from the Surrey Centre. The company was issued permit PE 5276 in 1980 for the discharge of treated domestic sewage at a maximum rate of  $14~\rm m^3/d$  from a proposed shopping center in the Township of Langley to 280 m of tile field. The permit also requires the effluent to have maximum concentrations of  $45~\rm mg/L~BOD_5$  and  $60~\rm mg/L~suspended$  solids.

The sewage is treated in a package aeration secondary-treatment plant. A sludge holding tank ( $40~\text{m}^3$ ) with aeration capacity of 70 m³/d and a hydraulic loading not to exceed 16 m³/d is incorporated to store effluent in the case of a system failure.

The small quantitites of sewage involved, the high level of treatment provided, and the fact that the treated sewage is discharged to a soil adsorption system over two km from the Serpentine River, would indicate that this operation should have no impact on the water quality of the Serpentine River.

# 3.3.8 OSWALD SEILER, SURREY (PR 5360)

This landfill is located just west from the Surrey Centre. It was issued permit PR 5360 in 1979, allowing the deposition of 523.9 m³ to the site over a one-year period. The total depth of the refuse was to be 1 m of hog fuel covered by 1 m of sand, and then the entire area was to be paved. Drainage ditches are located on three sides of the site carrying runoff water to the Serpentine River (0.75 km away). The depth to the groundwater table was estimated to be approximately 1 m.

The life of the landfill was projected to be until October 1979. In 1984, 1 acre of the 3-acre landfill was asphalted, the remaining portion was covered with 0.6 m of impervious construction material and a layer of sand and gravel. No impact on the Serpentine River is anticipated from this operation.

### 3.3.9 DIFFUSE SOURCES

The Serpentine drainage basin has 7 441 ha serviced by municipal sewers and 5 746 ha of unsewered agricultural area. Within this drainage basin there was a 1981 population (79) of 8 341 which resided in Cloverdale, 16 835 resided between Mahood (Bear) Creek and the Serpentine River, and 64 770 resided in the Newton Uplands. This latter area impacts the upper reaches of the Serpentine River and Mahood (Bear) Creek with stormwater runoff and is predicted to approximately double in population by 1996  $(124\ 200)^{(80)}$ . Therefore, stormwater management issues should be addressed to prevent substantial increases in chemical loadings and to minimize scouring of salmonid rearing and spawning channels in the upper reaches of the Serpentine River and Mahood (Bear) Creek.

Three feedlot operations, registered by the Beef Assurance Program as outlined in Section 2.3.1.2, are situated within the Serpentine drainage basin. The largest of these is located in the Newton Uplands near Mahood (Bear) Creek (maximum value of 6000 head recorded in 1982, mean five-year value of 842 head)(Figure 3). Facilities near the headwaters (26 cows, 26

calves, and 2 yearlings) of the Serpentine River, on the floodplain near Cloverdale (75 yearlings, 75 finishing cattle), and near the mouth of the river (57 cows, 58 calves, 6 finishing cattle) are more representative of the size of feedlots operating in the Lower Mainland.

Phosphorus and nitrogen loadings to the Serpentine River, assuming 100% transmission, were estimated to be about 28 000 kg and 393 900 kg, respectively. The impact of the release of these nutrients would depend upon the period of their release, and the flow in the river. Typical flow rates for the Serpentine River are not available. However, if the maximum recorded flows for Sites 08MH020 and 08MH060 were added, a flow in the Serpentine of about  $35~\text{m}^3/\text{s}$  would result. These loadings do not include losses from other agricultural activities such as the application of chemical fertilizers.

Using this flow, the following increases could be possible in the river, based upon the following release periods and assuming that the nutrients accumulate during a one-year period.

Period of Release	Potential Increase (mg/L)	
	Phosphorus	Nitrogen
4 weeks 2 weeks 1 week	27.8 55.5 111	32.6 65.1 130

Obviously, such large increases do not occur due to the transmission coefficient being less than 100%, accumulations not occurring for as long as one year, and larger flows being available for dilution. Actual riverine conditions are discussed in Section 3.4.3.

# 3.4 WATER QUALITY AND PROPOSED PROVISIONAL WATER QUALITY OBJECTIVES

Water quality data have been collected at ten sites along the Serpentine River, as well as on Hyland, Mahood (Bear), and Latimer Creeks from 1974 to 1982. These data have been reported by  $Arber^{(28)}$ , Bourque and  $Hebert^{(25)}$ , and Hirst et al.  $^{(33)}$  for the period 1974 to 1977. This report includes these data for seasonal and long-term trends in water quality and sediment chemistry data, and adds data collected between 1979 and 1982. The data are in Tables 11a and 11b, 12, 13, 14, and 15. The locations of the sites are shown on Figure 3.

Proposed designated water uses for the Serpentine River, and Hyland, Latimer, and Mahood (Bear) Creeks are the protection of aquatic life, wildlife, livestock, and irrigation.

### 3.4.1 pH AND BUFFERING CAPACITY

Total alkalinity on the Serpentine River (Table 11a) ranged from <0.5 to 113 mg/L at Site 1100049, with mean values ranging from 4.8 at Site 1100037 to 65.8 mg/L at Site 1100034. These data indicate that the Serpentine River has a moderate to low buffering capacity to acidic discharges. Generally, low alkalinity values occurred from January to March, a period with high precipitation. Under these conditions, streamflows are dominated by surface runoff which has had little contact with the surficial geological materials that impart buffering agents to the water.

In Mahood Creek (Table 13), alkalinity values ranged from 16.1 mg/L to 68.1 mg/L at Site 1100056. Mahood Creek, with a mean alkalinity value of 43.3 mg/L, had more buffering than the Serpentine River. Alkalinity data were not collected for either Hyland or Latimer Creeks (Tables 12 and 14); however, similar buffering capacities to Mahood Creek would be anticipated based upon data for pH.

The highest values for pH in the Serpentine River were recorded at Site 1100037 near the headwaters (range: 6.7-9.3; median 7.5), while the lowest

value was recorded at Site 1100049 at the Serpentine fen (range: 3.8-7.7; median 5.1). The pH range for the protection of aquatic life is 6.5 to  $9.0^{(31)}$ , while for drinking water it is 6.5 to  $8.5^{(7)}$ . Values less than 6.5 were found in four of sixty-six samples at Site 1100034, five of forty-one samples at Site 1100035, four of sixty-one samples at Site 0300057, nineteen of thirty-two samples from Site 1100049, three of fifty-eight samples from Site 0300059, and greater than 8.5 for five of forty-two samples from Site 1100037, one of 66 samples at Site 1100034, and two of sixty-one samples at Site 0300057.

In Mahood Creek, the minimum pH was 6.8 at Site 0300056, while the maximum was 8.1 at Site 1100056 (Table 13). Values in Hyland Creek were from 7.4 to 8.0 (Table 12) and from 6.7 to 8.1 in Latimer Creek (Table 14). No samples collected on any of the creeks were outside the working criteria pH ranges.

It is recognized that at higher pH values, ammonia toxicity increases. However, high (and low) pH values exist naturally in the Serpentine River. Such is not the case in Mahood, Hyland, or Latimer Creeks. To protect sensitive aquatic life, a provisional objective is proposed for Mahood (Bear) Creek, Hyland Creek, Latimer Creek and the Serpentine River. The pH should be in the range 6.5 to 8.5. If values upstream are greater than 8.5, then the change in pH going from upstream to downstream from a discharge or series of discharges should not exceed 0.2. The objectives are applicable along the lengths of these creeks and the river except in initial dilution zones of effluents, described in Section 2.4.1.

#### 3.4.2 HARDNESS AND METALS

Values for total hardness at Sites 1100036, 0300059, and 1100037 in the Serpentine River above the confluence of Mahood Creek did not appear to be affected by salt water intrusions. Mean values were 45.4, 51.8, and 36.9 mg/L, respectively (Table 11a). In Mahood Creek, mean values were

45.4 mg/L at Site 0300056 and 48.9 mg/L at Site 1100056 (Table 13). These data indicate that the water is soft, having a hardness value less than 75 mg/L of  $CaCO_3$ . No data for hardness were available for Hyland or Latimer Creeks.

Hardness values of Site 1100034, 3 km upstream from the mouth of the Serpentine River, were as high as 4 380 mg/L. These values reflect the presence of salt water (specific conductivity of 38 700  $\mu$ S/cm).

### 3.4.2.1 Metal Values in the Water Column

Values for total aluminum were collected at Site 0300057 and at Site 0300059 on the Serpentine River (Table 11a) and at Site 0300056 on Mahood Creek (Table 13). The average values at all sites exceeded the criterion of 0.1 mg/L for the protection of aquatic life $^{(6)}$ . However, these concentrations appear to be high background rather than the result of anthropogenic activities.

Average total copper values recorded at most sites exceeded the average water quality criterion of 0.002 mg/L for the protection of sensitive aquatic life at hardness below 50 mg/L $^{(66)}$ . The maximum recorded copper values seemed to increase in the Serpentine River going downstream, from a maximum recorded value of 0.007 mg/L at Site 1100037 near the headwaters, to a maximum value of 0.03 mg/L at Site 1100034 near the river mouth (Table 11a), although the maximum value at Site 1100035 was 0.05 mg/L. A maximum value of 0.02 mg/L was recorded at Site 0300056 in Mahood Creek (Table 13). One sample in each of Latimer (Table 14) and Hyland Creeks (Table 12) had 0.002 mg/L. The high values in the Serpentine River and Mahood Creek are presumably natural background.

Total lead values throughout the Serpentine River ranged from <0.001 mg/L at several sites to 0.044 mg/L at Site 1100037 (Table 11a). This maximum is substantially higher than found elsewhere in the drainage basin. Mean values at several sites equalled or exceeded the criterion for the protection of sensitive aquatic life (0.003 mg/L) for a minimum recorded

hardness of 9.2 mg/L $^{(84)}$ . Seven of 22 samples at Site 1100037 exceeded this criterion. A lead value of 0.054 mg/L was recorded at Site 1100056 in Mahood Creek (Table 13), while a value of 0.006 mg/L was found in Hyland Creek (Table 12), and a value of 0.003 mg/L in Latimer Creek (Table 14). These high lead values may originate from stormwater runoff; however, more work is required to confirm this.

The maximum total zinc value collected at Site 1100049 (0.24 mg/L) on the Serpentine Fen was higher than those from any other location within this drainage basin. The maximum zinc value in Mahood Creek was 0.12 mg/L at Site 1100056 (Table 13), 0.005 mg/L in Hyland Creek (Table 12), and <0.005 mg/L in Latimer Creek (Table 14). The mean concentration at the remaining sites (except Site 1100049) were lower than the most sensitive criterion level (0.05 mg/L) for the protection of aquatic life<sup>(6)</sup>. One value exceeded the criterion level at Sites 1100034, 1100035, 1100037, 1100056, and 0300057, while it was exceeded in seven values collected at Site 1100049. Seasonal variability was more pronounced at Site 1100037, in the uplands, where total zinc concentrations increased between October and February. Stormwater runoff may be a contributing factor to higher zinc values during this period (24).

Total iron values collected from most locations in Latimer, Mahood, and Hyland Creeks and in the Serpentine River exceeded the working criteria for drinking water  $(0.3~\text{mg/L})^{(7)}$ , and for the protection of aquatic life (0.3~mg/L to  $1.0~\text{mg/L})^{(72)}$ . The criterion of  $5.0~\text{mg/L}^{(76)}$  for continuous irrigation on all soils was exceeded in some samples at Sites 1100049, 1100035, and 0300059 on the Serpentine River and at both sites on Mahood Creek. These high iron values are thought to be naturally occurring. The drinking water criterion of  $0.3~\text{mg/L}^{(7)}$  is related to aesthetics, not health concerns.

All total manganese values collected at Site 0300057 in the Serpentive River exceeded the working criterion for drinking water  $(0.05 \text{ mg/L})^{(7)}$ . The working criteria for the protection of aquatic life are 0.10 to 1.0 mg/L $^{(71)}$ . The upper criterion was not exceeded in the Serpentine

River or Mahood Creek, although the lower criterion was. No data were collected in Hyland or Latimer Creeks. The drinking water criterion is related to aesthetics, not health.

More data are required for metals along the length of Mahood Creek to determine if the discharges from the two foundries are significantly affecting water quality.

### 3.4.2.2 Metal Values in Sediments

Metal values in sediments were examined at three sites in 1979 and three sites in 1982 (Table 11b). Sites 0300057 and 1100034 were approximately at the same location, near the mouth of the Serpentine River. Sites 0300059 and 1100036, upstream from the confluence of Mahood (Bear) Creek, were also close together (Figure 3).

Data for eight constituents from the six sites were compared to data from the Main Arm and North Arm of the Fraser River downstream from Pattullo Bridge (21), sediment cores collected from lakes in B.C.(22), and sediments accumulated in urban residential stormwater (24) (Table 11(c)).

Values for copper, iron, lead, manganese, nickel, and zinc in sediments from the Serpentine River were generally of the same order of magnitude or higher than values in sediments from the Main Arm of the Fraser River; cadmium values were lower in the Serpentine River (Table 11(c)).

Metal values in Serpentine River sediments were within the range of values in samples from B.C. lakes (22); cadmium, copper, cobalt, iron, and lead values in the Serpentine River were within the lower part of the range for B.C. lakes. Lead and zinc values were higher in stormwater sediments (24) than in sediments from the Serpentine River, while iron values from the river were higher than those from the stormwater. Manganese, copper and cobalt values from the Serpentine River were within the range of values from stormwater samples.

### 3.4.3 NUTRIENTS

#### 3.4.3.1 Nutrient Values in the Water Column

Maximum total ammonia nitrogen values throughout the Serpentine River ranged from 0.007 mg/L at Site 1100032 to 0.735 mg/L at Site 1100035. Criteria for the protection of aquatic life are in Table 29 for maximum values and Table 30 for average values. A comparison of actual high total ammonia values with their corresponding pH and temperature values indicated that the criteria for maximum and average total ammonia were met.

The maximum total ammonia nitrogen value in Mahood Creek was 0.38 mg/L at Site 0300056 (Table 13), at a pH of 6.5 and a temperature estimated at 5.0°C. The maximum ammonia level on Hyland Creek was 0.058 mg/L (Table 12). In Latimer Creek, the maximum value was 0.64 mg/L (Table 14) at a pH of 7.7 and a temperature of 18.2°C. None of these concentrations are of concern with respect to ammonia toxicity.

A concern exists for ammonia generated from animal wastes. Therefore, provisional water quality objectives are proposed for the total ammonia nitrogen concentration in the Serpentine River, Mahood Creek, Latimer Creek, and Hyland Creek. Total and average values should not exceed values listed in Tables 29 and 30. The objectives apply in all areas, except initial dilution zones of effluents described in Section 2.4.1. The average value is to be calculated from a minimum of five samples collected in a 30-day period.

Nitrite nitrogen values were as high as 0.246 mg/L at Site 0300057 in the Serpentine River (Table 11a), 0.029 mg/L at Site 0300056 in Mahood Creek (Table 13), 0.011 mg/L in Hyland Creek (Table 12), and 0.049 mg/L in Latimer Creek (Table 14). At chloride concentrations less than 2 mg/L, the B.C. criteria for nitrite are 0.02 mg/L as a 30-day average value and 0.06 mg/L as a maximum value  $^{(69)}$ . Thus the assumed criteria for maximum values are met in the three tributary creeks. The minimum chloride concentration at Site 0300057 in the Serpentine River was 10.8 mg/L (Table 11a) indicating

saltwater intrusion. The corresponding B.C. criteria are a maximum concentration of 0.6 mg/L and a 30-day average value of 0.2 mg/L. Thus nitrite levels is the watershed are below criteria to protect aquatic life.

Since this watershed is used by salmonids and because nitrite can be formed when ammonia is oxidized, provisional objectives are proposed for nitrite nitrogen in the Serpentine River, Mahood Creek, Latimer Creek, and Hyland Creek, as follows:

Chloride Concentration (mg/L)	Maximum Nitrite-N (mg/L) Concentration	30-day Average Nitrite-N (mg/L) Concentration
<2	0.06	0.02
2-4	0.12	0.04
4-6	0.18	0.06
6-8	0.24	0.08
8-10	0.30	0.10
>10	0.60	0.20

These objectives apply in all areas except the initial dilution zones of effluents, described in Section 2.4.1. The average value is to be calculated from a minimum of five samples collected in a 30-day period.

The highest total phosphorus values in the Serpentine River were 1.29 mg/L at Site 1100033 and 1.56 mg/L at Site 1100035 (Table 11a), both in June 1975. Values from Mahood Creek were as high as 0.273 mg/L (Table 13). The maximum value in Hyland Creek was 0.071 mg/L (Table 12), while in Latimer Creek it was 0.089 mg/L (Table 14). The maximum dissolved orthophosphorus was 1.01 mg/L at Site 1100035 (Table 11a) in the Serpentine River, 0.36 mg/L at Site 1100056 (Table 13) in Mahood Creek, 0.034 mg/L in Hyland Creek (Table 12), and 0.167 mg/L in Latimer Creek (Table 14).

These quantities of phosphorus are sufficient to cause excessive algal production, assuming phosphorus is the limiting nutrient and other factors

which would have a bearing on algal production such as availability of substrate, light availability, and river velocity are not limiting.

The water quality criteria for nutrients and algae in streams limits the periphyton chlorophyll-a levels to 100 mg/m² for protection of aquatic life<sup>(9)</sup>. Chlorophyll-a data were not available for the Serpentine River watershed. Since phosphorus and nitrogen can enter the watershed from agricultural practices, a provisional water quality objective of 100 mg/m² chlorophyll-a is proposed, applicable to naturally growing periphytic algae. The objective is to be evaluated from the average of at least five replicates. The objective, in the absence of monitoring data, will have to be a long-term objective. It applies to all parts of the Serpentine River, Mahood Creek, Latimer Creek, and Hyland Creek, except in the initial dilution zones of effluents, described in Section 2.4.1.

### 3.4.3.2 Nutrient Values in Sediments

Nutrient values in sediments were examined at three sites in 1979 and three sites in 1982. A description of the sites is in Section 3.4.2.2.

Stancil<sup>(21)</sup> reported data (mean values for transects) for samples collected by B.C. Research (Figures 11a and 12a in reference 21) for total phosphorus and Kjeldahl nitrogen adjacent to the Iona STP outfall. A comparison between these two locales is shown below.

Constituent	Serpentine River	Adjacent to Iona STP <sup>(21)</sup>
Total Phosphorus (mg/L Kjeldahl Nitrogen (mg/L		0.588 - 0.910 0.221 - 0.728

This indicates that the sediments in the Serpentine River generally have more phosphorus and nitrogen than sediments adjacent to the Iona STP outfall, an area considered to be heavily polluted by municipal effluent. Elevated phosphorus and nitrogen values may be attributed to the agricultural practices conducted within this drainage basin.

### 3.4.4 OXYGEN AND OXYGEN-CONSUMING MATERIALS

Dissolved oxygen values in the Serpentine River ranged from a minimum 1.1 mg/L at Site 0300059 to a maximum of 17.6 mg/L at Site 1100049 (Table 11a). Such a large range of values was typical, although not necessarily to the same degree, of dissolved oxygen levels found throughout the Serpentine River. Low dissolved oxygen values occurred typically between May and November.

Percent saturation ranges were also large in the Serpentine River. Values ranged from 9.7% at Site 0300059 to 278.7% at Site 1100033 (Table 11a). These extreme values also were typical, although not necessarily to the same degree, of percent saturation values throughout the River. Such wide ranging values are typical in a stream with heavy algal production. These extremes can cause stress to aquatic life.

A slight impairment to salmonids occurs at dissolved oxygen levels of 6.0 mg/L $^{(83)}$ . In comparing dissolved oxygen levels in the Serpentine River to this criterion, it is apparent that there are times when there are real concerns about oxygen levels in the river.

Spawning occurs in the Serpentine River (Figure 5) near Site 1100037, where dissolved oxygen levels were good, ranging from 8.0 to 14.7 mg/L. These levels will not impair salmonids (83).

Dissolved oxygen levels in Mahood Creek were also good, ranging from 7.5 mg/L to 13.4 mg/L at Site 1100056 (Table 13). Percent saturation values ranged from 76.3% to 135.2% at Site 1100056, indicating some stress to aquatic life could result, but certainly not in any degree close to that in the Serpentine River.

Dissolved oxygen values in Hyland Creek ranged from 8.5 to 10.8 mg/L (Table 12). In Latimer Creek, these ranged from 5.6 to 9.6 mg/L (Table 14). It was not possible to calculate percent saturation values in these two creeks. However, it is apparent that salmonids in Mahood, Latimer, and

Hyland Creeks have only slight to no impairment (i.e.,  $\geq 6.0$  mg/L)<sup>(83)</sup> at all times.

No impairment to salmonids occurs at dissolved oxygen concentrations  $\ge 8.0 \text{ mg/L}^{(83)}$ . This appears to have been achieved in Hyland Creek at all times, and in Mahood and Latimer Creeks most of the time.

Staged provisional objectives for minimum dissolved oxygen values are proposed for the Serpentine River. The first objective is for the water quality to meet 6 mg/L dissolved oxygen from June to October. The second objective (long-term) would be implemented when dissolved oxygen levels consistently met the first level. The second stage would increase the objective value for dissolved oxygen to 8.0 mg/L from June to October. In Latimer, Hyland, and Mahood Creeks, the proposed provisional objective is that the minimum dissolved oxygen value should be 8.0 mg/L at all times. This objective may not always be met in Mahood or Latimer Creeks. In addition, dissolved oxygen levels in the Serpentine River and Latimer, Hyland and Mahood Creeks should be, as a minimum, 11.0 mg/L when larvae and/or alevin are present. This objective will likely be met in areas of the river system where spawning occurs.

These objectives apply to discrete samples or in-situ measurements taken outside the initial dilution zones of effluents, described in Section 2.4.1.

Remedial activities to increase dissolved oxygen levels in the lower reaches of the Serpentine River have been initiated. These activities take into consideration fisheries, agriculture, and flood control concerns. At present, the use of local diffusers to introduce air into the river is being evaluated.

### 3.4.5 BACTERIOLOGICAL QUALITY

Fecal coliform levels ranged from <20 to >24 000 MPN/100 mL at Site 0300057 in the Serpentine River (Table 11a), while median values were as

high as 1 300 MPN/100 mL at Site 0300059 and 790 MPN/100 mL at Site 1100032. Values were also high in Mahood Creek, ranging from 330 to >2 400 MPN/100 mL at Site 0300056 (Table 13). Similar high levels were also found in Hyland Creek (Table 12) and Latimer Creek (Table 14). These high levels possibly originate from agricultural wastes.

The working criteria for livestock watering limit the concentration of fecal coliforms to 10 MPN/100 mL (90th percentile) for water used after disinfection only (7,30). For use of Serpentine River water for drinking and livestock watering, complete water treatment plus disinfection would be required; however, the B.C. Ministry of Health does not approve of the use of these waters for a drinking water supply. The working criteria for irrigation use are a geometric mean value of 200 MPN/100 mL, and a 90th percentile of 400 MPN/ 100 mL $^{(30)}$ . These recent working criteria are not met, and are more restrictive than the approved water quality objectives for the Fraser River  $^{(82)}$ .

Therefore, staged provisional (long-term) objectives for fecal coliforms in the Serpentine River, Hyland Creek, Latimer Creek, and Mahood Creek are proposed. The first objective (long-term) is that the 30-day geometric mean should be less than 200 MPN/100 mL year-round to protect irrigation and livestock watering. In the short-term, the geometric mean should be \geq1000 MPN/100 mL from April to October and the maximum \geq4000 MPN/100 mL. These objectives apply to discrete samples, as the mean or 90th percentile of at least five weekly samples taken in a 30-day period, outside the initial dilution zones of effluents, described in Section 2.4.1

Domestic water licence holders should be advised that, at present, complete water treatment may be necessary for Serpentine River water prior to use and that the B.C. Minsitry of Health does not approve of its use for drinking water. Further monitoring of fecal coliform and fecal streptococci should be conducted to determine the present levels and sources of fecal contamination in the basin. This will indicate the feasibility of attaining the long term objectives.

#### 3.4.6 SOLIDS

Dissolved solids values from near the mouth of the Serpentine River were significantly higher than those observed further upstream, due to saltwater intrusion. Saltwater influences, as indicated by high dissolved solids levels and specific conductivity measurements, have been evidenced as far upstream as Hyland Creek. In Mahood Creek (Table 13), dissolved solids were at a maximum of 338 mg/L at Site 0300056. In Hyland Creek and Latimer Creek, values were about one-half this number (Table 12 and 14). The most restrictive dissolved solids criteria of 500 mg/L are to protect drinking water (7) and irrigation (76).

Suspended solids values were relatively uniform throughout the system, with a maximum value of 85 mg/L at Site 1100041. Maximum suspended solids values in the creeks were 129 mg/L at Site 0300056 in Mahood Creek (Table 13), 18 mg/L in Hyland Creek (Table 12), and 14 mg/L in Latimer Creek (Table 14).

Bourque and Hebert (1982)<sup>(25)</sup> examined suspended solids data between 1974 and 1975, finding values ranging from 5 mg/L to 73 mg/L.

The B. C. water quality criteria for particulate matter (14) limit the quantity of induced suspended solids to not exceed 10 mg/L when background levels are less than 100 mg/L, nor to be more than 10% of background when background is greater than 100 mg/L, for the protection of aquatic life. In the Serpentine River, Mahood (Bear) Creek, Hyland Creek, and Latimer Creek, the provisional objectives proposed are that induced levels of suspended solids should not exceed 10 mg/L when background levels are less than 100 mg/L, nor should these levels be more than 10% of background when background is greater than 100 mg/L. These objectives apply to discrete samples collected outside the initial dilution zones of effluents described in Section 2.4.1.

In addition, to protect salmonid spawning areas, there should be no significant (95% confidence level) induced benthic sedimentation on the basis of accumulation by weight for particles <3 mm.

High turbidity values were associated with the periods September to March, and July. During five sampling periods (July to September 1974, July 1975, March 1976, September 1976 to January 1977, and November 1978) high turbidity values were recorded throughout the Serpentine River system. This underlines the contribution of stormwater runoff to water quality within the river. The maximum turbidity (192 NTU) was at Site 1100049 (Table 11a). It should be noted that turbidity frequently exceeded the maximum acceptable level of 5 NTU for drinking water, indicating that filtration or its equivalent would be needed to remove suspended solids prior to drinking, during much of the year.

The maximum turbidity level in Mahood Creek was 104 NTU at Site 1100056 (Table 13). In Latimer Creek the maximum level was 7.3 NTU (Table 14) while the only measurement in Hyland Creek was 3.8 NTU (Table 12).

The B.C. criteria for turbidity<sup>(14)</sup> are that induced levels should not be more than 5 NTU when background levels are less than 50 NTU, nor be more than 10% of background when background levels exceed 50 NTU. A provisional water quality objective is proposed for turbidity in the Serpentine River, Mahood Creek, Latimer Creek, and Hyland Creek. The objective is that the induced turbidity levels should not be more than 5 NTU at a background level less than 50 NTU, nor be more than 10% of background if background exceeds 50 NTU.

The objective applies to discrete samples taken outside the initial dilution zones of effluents, described in Section 2.4.1.

# 3.4.7 POLYCHLORINATED BIPHENYLS (PCBs)

In Section 3.3.3, it was calculated that PCBs could be increased in Mahood Creek from a discharge from B.C. Hydro. Concerns were expressed that

ambient levels could exceed the drinking water criterion of 3  $\mu g/L^{(75)}$  and the aquatic life criterion of 0.001  $\mu g/L^{(76)}$ .

Of equal concern is the possibility that PCBs could be deposited with bottom sediments, thereby being available to enter the food web, eventually contaminating fish. Working water quality criteria for PCBs are 0.05  $\mu g/g$  (dry weight) in sediments  $^{(75)}$ , 2  $\mu g/g$  (wet weight) in edible fish tissues to protect human consumers  $^{(75)}$  and 0.1  $\mu g/g$  (wet weight) in whole fish to protect fish eating mammals in water  $^{(72)}$ . For the Fraser River, Swain and Holms proposed 0.03  $\mu g/g$  (dry weight) for bottom sediments and 0.5  $\mu g/g$  (wet weight) in fish muscle, to protect aquatic organisms  $^{(82)}$ . The former had been derived arbitarily to reflect background Fraser River concentrations while the latter was based on 1972 criteria  $^{(26)}$  which were greater than the detection limit of 0.3  $\mu g/g$  (wet weight)  $^{(82)}$ . In the absence of site-specific data for PCBs, the use of more restrictive criteria is warranted as a conservative approach to establishing provisional objectives.

For these reasons, provisional water quality objectives are proposed for PCBs in the water column, sediments, and in fish muscle and in whole fish in Mahood Creek and in the Serpentine River. The objectives are that the maximum PCB concentrations should not exceed 0.001  $\mu$ g/L in the water column, 0.03  $\mu$ g/g (dry-weight) in the bottom sediments, 0.5  $\mu$ g/g (wet-weight) in muscle of large fish consumed by man, and 0.1  $\mu$ g/g (wet weight) in whole small fish subject to predation.

The objectives apply to water or sediment samples collected outside the initial dilution zones of effluents, described in Section 2.4.1. The objective for fish applies to fish of any species caught in any part of Mahood Creek or the Serpentine River, including the initial dilution zones of effluents. The term PCBs applies to the sum of Aroclor 1242, 1254, and 1260, which may be present in water, sediment, or fish. The achievement of the objective for bottom surface sediments should be determined from the average of at least three replicate sediment samples taken from the same site.

It is recognized that some of the proposed objectives are lower than analytical detection limits. It is recommended that these detection limits be lowered. In the interim, values below detection will be considered as meeting the objectives.

#### 3.4.8 CHLOROPHENOLS

In March 1984, a fish kill (>5000) occurred in Hyland Creek as the result of a spill of 45 000 litres of pentachlorophenol and tetrachlorophenol. Chlorophenols are not discharged normally into the Serpentine River or its tributaries. The 1984 spill was an isolated incident according to investigations carried out by Waste Management staff. Therefore, no objectives will be proposed for chlorophenols.

### 3.5 RECOMMENDED MONITORING

Water quality objectives have been proposed on Hyland, Latimer, and Mahood Creeks and on the Serpentine River. It is recommended that as a minimum, objectives should be checked at Site 0300056 on Mahood Creek, at sites near the mouths of Latimer and Hyland Creeks, and at Sites 0300059 and 0300057 on the Serpentine River. Additional sampling sites should be used when the objectives are exceeded to determine the extent of the problem. A minimum sampling frequency of five weekly samples is recommended. Samples should be analyzed for pH, temperature, dissolved oxygen, specific conductivity, ammonia, nitrite, nitrate, fecal coliforms, fecal streptococci, turbidity, suspended and dissolved solids, periphyton chlorophyll-a, dissolved orthophosphorus, total phosphorus, and total and dissolved copper, iron, manganese, and lead. However, resources and other regional priorities will determine the exact program.

Two special studies have been recommended for the Nicomekl River (Section 4.5): one to determine effects of agriculture and a second to examine the impact of urban stormwater runoff. The findings of these special studies should be applicable to the Serpentine River watershed for estimating levels and sources of fecal contamination.

### 3.6 CONCLUSIONS

There are four permitted point source discharges to Mahood Creek, one to Hyland Creek, and none to Latimer Creek or the Serpentine River. The discharge into Mahood Creek could increase suspended solids, oil and grease, iron, and PCBs. The discharge into Hyland Creek might increase suspended solids.

The greatest concern in this watershed is the contribution from agricultural sources, which increases the availability of nutrients. These in turn can increase algal production, and subsequently cause dissolved oxygen elevations and depressions.

The pH in the Serpentine River can vary as much as 3.0 pH units from minimum to maximum. Alkalinity values can be from moderate to low, while the water would be classed as soft. Some high metal values for total aluminum, copper, lead, iron, manganese, and zinc were thought to be naturally occurring. Phosphorus values were high enough to cause algal blooms, if phosphorus is the limiting factor. Ammonia and nitrite toxicity were not thought to be a concern. Large variations occur in dissolved oxygen and percent saturation values, likely due to algal photosynthesis, respiration, and decomposition. Fecal coliform values were too high to would permit drinking without complete treatment. Dissolved solids reflected the influence of saltwater intrusion near the mouth, but were not so high in the upper reaches as to restrict any water uses.

### 4. NICOMEKL RIVER

The Nicomekl River has a drainage area of 175 km². It receives runoff from the southeast slope of Clayton Upland and the Langley Upland (Figure 5). Three of the major tributaries within the drainage area are Anderson Creek with a drainage area of 24.7 km², Murray Creek with a drainage area of 27.1 km², and the old logging ditch with a drainage area of 7.7 km² (Figure 5). The area is underlain by stoney marine clays, except in the uplands around Anderson Creek. This portion of the drainage basin is underlain by 27 m of sand and gravel which is an unconfined aquifer recharged directly by precipitation in the area.

### 4.1 HYDROLOGY

The locations of flow gauging sites within the Nicomekl drainage basin are on Figure 3. These sites are situated on Anderson Creek (08MH104), Murray Creek (08MH107, 08MH129), and the Nicomekl River below the confluence of Anderson Creek (08MH050) and Murray Creek (08MH105).

Historical average daily flows have ranged from 0.062 m³/s to 28.3 m³/s at Site 08MH105 and, for a later period of record, from 0.125 m³/s to  $35.4 \text{ m}^3/\text{s}$  at Site 08MH050, both on the Nicomekl River. For Murray Creek, average daily flows ranged from 0.011 m³/s to 14.5 m³/s at Site 08MH107 and for a later period of record, from 0.001 m³/s to 19.7 m³/s at Site 08MH129. In Anderson Creek, average daily flows have ranged from 0.108 m³/s to  $17.2 \text{ m}^3/\text{s}$ .

Ten-year and two-year, seven-day average low flow values were calculated to be 0.13 m $^3$ /s and 0.24 m $^3$ /s at Site 08MH050 on the Nicomekl River below Anderson Creek; 0.089 m $^3$ /s and 0.117 m $^3$ /s at Site 08MH105 on the Nicomekl River below Murray Creek; 0.13 m $^3$ /s and 0.149 m $^3$ /s at Site 08MH104 on Anderson Creek; and 0.001 m $^3$ /s and 0.006 m $^3$ /s at Site 08MH129 on Murray

Creek. At Site 08MH107, based on data from  $^4$  years, the two-year return flow was 0.025 m $^3$ /s.

### 4.2 WATER USES

Sixty water licences located within the Nicomekl drainage area (Figures 3 and 3b) have been issued for an annual withdrawal of 2 272 dam<sup>3</sup>. Fortyfour licences were issued for the withdrawal of 2 239 dam<sup>3</sup> for irrigation (98.5% of the annual withdrawal), four licences were issued for the withdrawal of 18.4 dam<sup>3</sup> (0.8% of the annual withdrawal volume) for land improvement (construction of ponds), seven licences were issued for the withdrawal of 9.8 dam<sup>3</sup> (0.4% of the annual withdrawal volume) for stock watering, three licences were issued for the withdrawal of 4.15 dam<sup>3</sup> (0.2% of the annual withdrawal volume) for drinking water, and two licences were issued for the withdrawal of 1.23 dam<sup>3</sup> (0.1% of the annual withdrawal volume) for industrial use (for a greenhouse and a fish culture operation).

Steelhead and cutthroat trout utilize the Nicomekl River system. Coho salmon spawn in the Nicomekl River (Figure 5) between 21 km to 23 km and 26 km to 30 km upstream from Mud Bay. Spawning occurs between 0.5 km to 10.6 km on Anderson Creek and from 1 km to 8.6 km on Murray Creek, upstream from their confluences with the Nicomekl River.

Coho in the Nicomekl River have demonstrated similar patterns of population fluctuations (Table 15) as those described for the Serpentine River (Section 3.2). Spawning in this system begins in early November, a half month later than the Serpentine River, with the peak of spawning occurring in late November and ending in late December.

There is no primary-contact recreation in the Nicomekl River.

### 4.3 WASTE DISCHARGES

The locations of operations discharging effluent within the Nicomekl drainage basin are represented by their permit numbers, issued pursuant to the Waste Management Act, in Figures 3 and 3b.

### 4.3.1 LANDFILLS

# 4.3.1.1 Rocla Concrete Pipes, Langley (PR 4965)

This company operated a 4.9 ha refuse site located about one km north from the Nicomekl River in Langley (Figure 3b). The company was issued a permit in 1978 for the disposal of waste concrete (80%) and broken concrete pipe (20%) at a rate of 2 m³/d. The location of the site is just south from the Fraser River dyke highway in North Langley. The site was estimated to have a service life of three years. The covering requirements for this operation were that wastes were to be covered every 40 days. No drainage problems have been reported in connection with this operation. The highest recorded water table was 3 m from the surface.

The landfill, operated by Flexlox Ltd. for Rocla Concrete Pipes, was closed in 1984. The nature of the landfilled material should not result in a negative impact on the Nicomekl River.

# 4.3.1.2 Township of Langley (PR 1898)

The Township of Langley operated a municipal landfill until 1978. It was located about one km south from the Nicomekl River, equidistant from Anderson and Murray Creeks (Figure 3b). The landfill was located in an excavated pit and received refuse from a population equivalent of 50 000 persons.

Leachate generated at the landfill discharged into a minor tributary (formerly a salmon spawning stream) of the Nicomekl River. Elevated levels of ammonia, COD, total organic carbon, pH, and specific conductivity were reported in samples collected from the tributary downstream from this operation. Fungal blooms (Leptomitaceas) have been observed in the tributary immediately adjacent to the landfill operation. It has been proposed that the tributary be diverted around the landfill and a leachate collection system be installed. The impact on the Nicomekl is not known but

is probably of concern since fungal growths have been observed at least one kilometre downstream.

#### 4.3.2 COOLING WATER

### 4.3.2.1 Mansonville Plastics Ltd. (PE 4413)

This polystyrene foam manufacturing plant is located just north from the Nicomekl River and downstream from its confluence with Anderson Creek (Figure 3b). The company was issued a permit in 1976 to discharge cooling water to the Nicomekl River and boiler blowdown to a steel tank in a gravel pit. Permit PE 4413 allows the discharges at a maximum rate of 109 m³/d and 0.09 m³/d, respectively. The minimum dilution available to the cooling water after complete mixing, based upon the 10-year, seven-day low flow of 0.13 m³/s and the maximum discharge rate, would be 103:1, or 190:1 with the two-year low flow. The cooling water must be discharged at a temperature less than 32° C. Values for the flow of the cooling water collected from 1977 to 1979 ranged from 16.4 to 33.7 m³/d (mean, 25.1 m³/d). No cooling water discharges have occurred since 1979.

The cooling water was treated by discharging it to a wide ditch containing a flotation trap, screens, and two dams 3 m apart. Screens and flotation traps were used to capture any polystyrene chips before flows entered the Nicomekl River.

It is not known what effect this effluent had on the Nicomekl River since no data on the actual temperature of the discharged cooling water have been recorded. However, the available dilution presumably ensured that there were no appreciable effects on the water quality of the Nicomekl River.

Boiler blowdown is drained daily. The discharge enters an exfiltration pond via a 1.5 m  $\times$  0.6 m steel tank with perforations on the lower portion of the tank to allow the discharge to enter the pond. No further informa-

tion about the boiler blowdown water was available but, given the very small quantity and its discharge to ground, no effects on surface water quality are probable.

#### 4.3.3 SOIL DISPOSAL SYSTEMS

# 4.3.3.1 Ambassador Industries Ltd., Langley (PE 5269)

This company is located about two km south from the Nicomekl River and equidistant between Anderson and Murray Creeks. It discharges domestic sewage from a commercial complex at an average flow rate of 19 m³/d after treatment in an aeration treatment plant with a capacity of 18.6 m³. The treated sewage is discharged to two disposal fields (minimum of 153 m of tile each). Permit PE 5269 requires that the effluent contain no more than 45 mg/L BODs and 60 mg/L suspended solids. Stormwater from the adjacent paved parking lot is disposed of in rock pits.

Data collected for the treated effluent from this operation are in Table 16. These limited data indicate that the limits were not met with respect to BOD<sub>5</sub> removal. Effluent from this operation should not negatively impact water quality in the Nicomekl River, Anderson, or Murray Creek owing to the distance from the point of discharge to these water bodies.

### 4.3.3.2 Willowbrook Meats Ltd., Langley (PE 2072)

This company, located near the headwaters of the Nicomekl River, stopped operating in 1985. Permit PE 2072 was first issued in 1976 for the discharge of uncontaminated cooling water, which originated from a refrigeration compressor. The cooling water entered a perimeter drainage and stormwater runoff ditch, at an average weekly discharge rate of 45.5 m³. The ditch then discharged to the Nicomekl River. The process effluent was trucked to a municipal treatment facility in Fort Langley (PE 4339) as surface irrigation or sub-surface disposal was not a viable alternative, due to the nature of the soil.

### 4.3.3.3 Big A Tank Services Ltd. (PE 5968)

This company operates throughout the Fraser Valley, injecting effluent from a small-scale abattoir (Westcoast Reductions) and confined livestock facilities into the ground at different locations to a depth of about 30.5 cm. This occurs over the growing season from April to October. The wastes are hauled to a treatment facility in Langley during the rest of the year.

The permittee is required to submit a monthly schedule to the Regional Waste Manager outlining the date and location of discharge, nature and quantity of waste applied and area to which waste was applied. Permit PE 5968 limits the maximum flow rate to 180 m<sup>3</sup>/d.

This operation has stopped using effluent from the abattoir, but continues to haul the washwater from the abattoir to the Langley sewage treatment facility. The manure from the confined livestock facility is stored in pits and applied to the fields throughout the Lower Mainland. This operation would therefore have no direct impact on the water quality of the Nicomekl River.

#### 4.3.3.4 Western Paving Ltd., (PE 2367)

This conventional gravel washing plant, located one km to the west of Anderson Creek and about four km upstream from its confluence with the Nicomekl River (Figure 3b), ceased operations in 1985. Washwater flowed at a permitted rate of 2 050 m³/d to a settling pond where the supernatant exfiltrated to other ponds and the surroundings. Water was recycled to the plant from an adjacent pond into which no gravel washwater was added. The plant held Permit PE 2367 since 1973.

Data in Table 17 indicate that the discharge rate allowed by the permit was not exceeded between 1974 and 1978. This settling/recycle system, with no positive discharge, would not have affected the water quality of Anderson Creek.

# 4.3.3.5 Noel Booth Elementary School, Langley (PE 6323)

This elementary school is located west from Anderson Creek, and about four km upstream from its confluence with the Nicomekl River. It discharges domestic sewage from a population projected to range from 360 to 480 students. The wastewater is discharged to a 30 m³ septic tank and then to two fields with 549 m of tile each, located approximately 97 m from Anderson Creek. At this location, Anderson Creek dries up during the summer; however, there would be little or no discharge to the tile system during this time.

The soil in the area consists of loamy sand and gravel. Percolation rates were found to average between 11 cm/minute and 4.4 cm/minute, depending on soil depth. No hardpan or groundwater was found to a depth of 3.7 m. Permit PE 6323 restricts the maximum discharge of treated wastewater to  $24.5 \, \text{m}^3/\text{d}$ .

This system should not affect Anderson Creek due to its distance from the creek, duration of operation (when flows exist in the creek), and the small volume of wastewater discharged to ground.

# 4.3.3.6 School District No. 35 Brookswood Junior Secondary School, Langley (PE 3148)

This school is located about one km east from Anderson Creek, and about four km upstream from its confluence with the Nicomekl River. Permit PE 3148 was issued in 1974 to allow a maximum discharge of 54.6  $\rm m^3/d$  of septic tank effluent from the school, which has a student population of 1 000. The domestic sewage is treated in a septic tank (capacity:  $45.6~\rm m^3/d$ ) prior to being discharged to two tile fields (840 m each).

This operation should not affect water quality in Anderson Creek due to its distance from the creek, duration of operation, and the small volume of wastewater discharged to ground.

#### 4.3.3.7 Walter H. Arber and George A. Smith (PE 218)

This 37-unit mobile home park is located just west from Anderson Creek and 3 km upstream from its confluence with the Nicomekl River. Domestic sewage is treated in a 63.6 m³ septic tank prior to disposal in a 885 m tile field. Permit PE 218, issued in 1968, allows the discharge of 29.55 m³/d. This effluent should not affect the water quality of Anderson Creek or McPherson Creek (a tributary) as it is discharged to a soil absorption system, and the volume of wastewater discharged is small.

#### 4.3.3.8 Old Yale Investments (1982) Limited (PE 4690)

This poultry processing plant is located near the headwaters of Anderson Creek. It operates from April through mid-December, with peak production in mid-November reaching 13 600 kg/d. The permit, issued in late 1977, was ammended in June 1983 to allow the following maximum values: flow, 114 m $^3$ /d; suspended solids, 75 mg/L; total solids, 500 mg/L; BOD $_5$ , 200 mg/L; oil and grease, 40 mg/L; and pH, 5-7. The amendment also allowed the company to install a blood recovery tank, coarse and fine screens, chemical coagulation, dissolved air flotation, a facultative lagoon, and sprinklers for spray irrigation of effluent on hay crops on adjacent land.

No data are available to determine if the new treatment works will ensure no adverse effects on water quality in Anderson Creek. Monitoring upstream and downstream from areas of spray irrigation is recommended when spray irrigation is taking place.

#### 4.3.4 SURREY COOPERATIVE ASSOCIATION, Surrey (PE 4262)

The Surrey Cooperative Association bulk petroleum storage plant is located about four km downstream from Anderson Creek on the Nicomekl River (Figure 3). Permit PE 4262 allows the discharge of an average 19.3  $m^3/d$  and a maximum of 430  $m^3/d$  of stormwater to an unnamed ditch which flows to the Nicomekl River. At the maximum discharge rate and at the 10-year, seven-day

low flow of 0.13 m³/s in the Nicomekl River below Anderson Creek, a minimum dilution of 26:1 is available, or 48:1 with the two-year low flow. The actual minimum dilution is probably much greater because it is very unlikely that the maximum stormwater discharge would occur during low streamflow periods. An average stormwater discharge would be diluted about 600:1 during the 10-year low flow. The permit allows concentrations of 20 mg/L suspended solids and 5 mg/L oil and grease, with a pH range from 6.5 to 8.5.

The stormwater is treated in an oil separator with a capacity of 10.8 m<sup>3</sup>. It provides a retention time of 13 hours, based on an average storm, and 0.6 hours based on the maximum storm.

Limited data for the operation are included in Table 18. The mean oil and grease value of 37 mg/L for all values in Table 18 indicates that the oil separator was not reducing values adequately in the discharge. However, the large dilution available in the Nicomekl River may have precluded any measurable or noticeable effect on the river.

#### 4.3.5 DIFFUSE SOURCES

Feedlots within the Nicomekl River watershed registered with the Ministry of Agriculture's Beef Assurance Program are on Figures 3 and 3b. The largest of these operations is located at the headwaters of Anderson Creek (267 yearlings). One facility was located on the Nicomekl River near Langley (105 cattle: 52 cows, 48 calves, 4 yearlings, 1 finishing cattle - Figure 3b), and one was located within the floodplain near the mouth of the river between the Serpentine and Nicomekl Rivers (121 cattle: 57 cows, 64 calves - Figure 3).

An estimate of phosphorus and nitrogen loadings to the Nicomekl River is in Table 28. It was estimated that 10 700 kg of nitrogen and 1 000 kg of phosphorus could be generated and potentially transported to the river. Assuming 100% transmission to the river, which is unlikely, the nutrient levels in the Nicomekl would increase as indicated below for different flows and release periods:

Increase	(mg/L)	With	Varying	River	Flows
	and R	elease	Periods	3	

Constituent	River Flow (m³/s)	4 weeks	2 weeks	1 week
Nitrogen	35.4	0.12	0.25	0.5
	28.3	0.16	0.31	0.62
	0.24	18.4	36.9	73.7
	0.13	34	68	136
Phosphorus	35.4	0.012	0.023	0.047
	28.3	0.015	0.029	0.058
	0.24	1.72	3.44	6.89
	0.13	3.18	6.36	12.7

These data indicate that phosphorus and nitrogen could be increased significantly at low flows, and even at peak flows when these nutrients likely would be released. Longer release periods and smaller rates of transmission would reduce these impacts. Actual riverine conditions are discussed in Section 5.4.3.

Stormwater runoff from Langley could also be a source of contaminants such as lead, zinc, fecal coliforms, and PCBs. The impact of these contaminants would depend upon the volume of stormwater discharged. No attempt will be made to estimate these loadings for the purpose of this assessment.

In addition, septic tanks discharging to tile fields are potential sources of nutrients and fecal coliforms. The loading from these sources will depend upon the number of systems in the watershed, the distance each is from the river, the type of soil each is located in, and the volume treated in each system. Estimating these loadings is beyond the scope of this assessment.

# 4.4 WATER QUALITY AND PROPOSED PROVISIONAL WATER QUALITY OBJECTIVES

Several studies have reported on water quality within the Nicomekl River drainage basin (3,25,28,33). This report incorporates the findings of those studies and assesses water chemistry data collected at 13 sites on the Nicomekl River and two sites on both Anderson Creek and Murray Creek (see Figures 3, 3b) over an eleven-year period (1972-1983). No data have been collected since 1982. From these data, water quality objectives for the Nicomekl drainage basin are proposed. Data collected from 1979 to 1983 for water chemistry are summarized in Tables 19, 21, 22, those for sediment chemistry are in Tables 20 and 23.

Present water uses which should be protected in the Nicomekl River include: irrigation, stock watering, and the protection of aquatic life and wildlife.

# 4.4.1 pH AND BUFFERING CAPACITY

Median pH values ranged from 7.2 at Sites 0300060 to 7.75 at Site 1100003 on the Nicomekl River, from 7.6 to 7.8 in Anderson Creek, and from 7.45 to 7.5 in Murray Creek. A general seasonal trend was observed at all sites throughout this system. Median pH values between January and April (6.9 to 7.7) were lower than those collected between May and August (7.5 to 8.26). As was observed in data from the Little Campbell River, low pH values occurred during periods of low flow, high temperatures, and high ammonia values. The lower values in the January to April period reflect the presence of large quantities of runoff which are poorly buffered due to the short residence time on the soil.

Geographically, samples collected from sites (1100002, 0300060, 1100003, 1100004, 1100005) between Cloverdale and the mouth of the river (Figure 3) had lower median pH values than from other sites on the Nicomekl River, Anderson Creek, or Murray Creek. The lowest values for pH ranged from 6.4 to 6.8 at these five sites compared to values of 6.9 to 7.2 in

upper reaches of the Nicomekl River, 7.2 in Murray Creek, and 7.0 in Anderson Creek. Most values were above the minimum working criterion for pH of 6.5 to protect aquatic life and drinking water (31,34).

Values greater than the pH range for the protection of aquatic life  $(6.5 to 9.0)^{(31)}$  occurred in three of thirty-five samples collected at Site 0300060 (value range: 9.0 to 9.3) and in three of forty-one samples collected at Site 1100003 (value range: 9.0 to 9.2). These values occurred in May 1977, July and August, 1982 at Site 0300060 and in September 1974, July 1975 and August 1979 at 1100003. Both of these sites are located near the mouth of the Nicomekl River. Since the upper limit of the working criteria for drinking water is  $8.5^{(34)}$ , a larger number of samples exceeded this limit at the two sites over a period of years. The number of excursions were six at Site 0300060 and seventeen at Site 1100003. addition, one excursion (of 16 samples) was recorded at Site 0300064 on Murray Creek and seven (of 42) at Site 1100002 near the mouth of the Nicomekl River. There are no licenced drinking water withdrawals in the lower reaches of the Nicomekl River. The pH seems to be naturally high in some areas of the basin. A provisional water quality objective is proposed to protect drinking water supplies and aquatic life. The provisional objective proposed is that the pH should be from 6.5 to 8.5 in the Nicomekl River and its tributaries. If upstream values are greater than 8.5, values should not change downstream from a discharge or series of discharges by more than 0.2 units. These objectives apply to all discrete samples collected in the Nicomekl River, and do not apply in initial dilution zones of effluent described in Section 2.4.1.

Mean alkalinity values at sites near the mouth of the river to Clover-dale (1100002, 1100003, 1100004) ranged from 54.3 mg/L to 76.6 mg/L, indicating a higher buffering capacity than at sites further upstream (0300060, 1100005, 1100006, 0300061, 0300062, 1100007; mean value range: 46.7 mg/L-62.5 mg/L), on Anderson Creek (49.8 mg/L - 40.3 mg/L) and Murray Creek (37.7 mg/L - 41.4 mg/L). Alkalinity values less than 40 mg/L were associated with the period between January and April. The highest alkalinity values usually occurred between June and November, during periods of low

flow and increased biological activity. Anderson Creek and Murray Creek also displayed similar seasonal tendencies. However, in the period between September and November, alkalinity values from sites on Murray Creek decreased to less than 40 mg/L.

#### 4.4.2 HARDNESS AND METALS

Water in the Nicomekl River upstream from about the Township of Langley in Anderson Creek and in Murray Creek would be classified as soft (having a hardness value less than 75 mg/L of  $CaCO_3$ ). Higher concentrations were found at sites near the mouth of the Nicomekl River which indicate the presence of salt water entering the system from Boundary Bay and Mud Bay.

Seasonal variability in hardness data was more pronounced near the mouth of the river than at sites near the Township of Langley and further upstream. Typically, values increased to a maximum between July and September followed by a decrease in samples collected between October and December. A factor contributing to these fluctuations is higher flows in the river which prevent saltwater intrusion in the lower reaches of the Nicomekl River.

#### 4.4.2.1 Metal Values in the Water Column

A total of 20 total aluminum measurements were made on the Nicomekl River in 1982 at Sites 0300060 near the mouth, 0300061 downstream from Langley, and 0300062 upstream from Langley. The values ranged from 0.03 to 0.77 mg/L, with 19 of 20 values exceeding the working criterion of 0.10 mg/L for the protection of aquatic life $^{(5)}$ . However, a water quality objective is not proposed for aluminum since it is believed to be naturally occurring and not the result of anthropogenic activity.

Dissolved boron values collected near the mouth of the Nicomekl River at Sites 1100002, 1100003, and 1100005, during the period between May and August ranged from <0.1 mg/L to 1.3 mg/L. Swain and Alexander (3) had

reported a median value of 0.1 mg/L (n=19) at Site 1100003. Higher boron values near the mouth of the river can be attributed to the intrusion of saltwater (seawater contains  $\approx$  4.5 mg/L boron) from Mud Bay. The working water quality criterion to protect sensitive crops is 0.5 mg/L<sup>(76)</sup>, and 5 mg/L for stock watering. Thus there may be times when the river near the mouth should not be used to irrigate some crops. These criteria, and that for drinking water (5.0 mg/L)<sup>(7)</sup>, were not exceeded at any sites on Anderson Creek or Murray Creek.

The highest concentrations of total chromium were at Site 0300060 near the mouth of the Nicomekl River (0.016 mg/L). Most values were below the detection limit of 0.01 mg/L. The working water quality criteria are 0.02 mg/L for fish, 0.002 mg/L for phytoplankton and zooplankton  $^{(76)}$ , 0.05 mg/L for drinking water  $^{(7)}$ , 1 mg/L for livestock, and 0.1 mg/L for irrigation  $^{(76)}$ . Thus there may have been times when phytoplankton and zooplankton were affected. Detection limits were generally too high to determine if this concern exists along the river. A similar concern exists in Murray Creek (maximum of 0.013 mg/L at Site 0300064) and Anderson Creek (maximum value of 0.007 mg/L at Site 0300063). Additional monitoring should be carried out to determine if these levels are natural and if not, their source.

Median and mean total copper values for the Nicomekl River decreased from sites near the mouth (Site 0300060: 0.004 mg/L and 0.006 mg/L; Site 1100003: 0.005 mg/L and 0.006 mg/L) to sites upstream (Site 0300062: 0.002 mg/L and 0.0018 mg/L; Site 1100007: 0.0025 and 0.005). Median and mean total copper varied little in Anderson Creek (Site 1100017: <0.001 mg/L and 0.002 mg/L) and in Murray Creek (Site 1100022 mean and median values: 0.003 mg/L, Site 0300064 mean and median value: 0.003 mg/L). There were no seasonal trends observed in copper values obtained at any site within the Nicomekl River, Anderson Creek, or Murray Creek.

Provisional B.C. criteria for total copper concentrations  $^{(66)}$  for stock watering (0.3 mg/L) and irrigation (0.2 mg/L) were not approached in

any of the samples collected on the Nicomekl River, Anderson Creek or Murray Creek. However, the criteria for the protection of aquatic life (30-day mean: 0.002 mg/L if hardness <50 mg/L; maximum: 0.007 mg/L, 0.011 mg/L and 0.021 mg/L when hardness, as CaCO, is 50 mg/L, 100 mg/L, and 200 mg/L, respectively) were exceeded in September 1975 at Sites 0300062 (0.02 mg/L) and 0300061 (0.02 mg/L) on the Nicomekl River, at Site 0300063 (0.02 mg/L) on Anderson Creek, and Site 0300064 (0.02 mg/L) on Murray Creek; in July 1979 the criteria were exceeded at Sites 1100006 (0.018 mg/L), 1100007 (0.016 mg/L), and 1100003 (0.016 mg/L) on the Nicomekl River, and at Site 11000022 (0.016 mg/L) on Murray Creek.

Although high values have been measured at several sites, values less than the criterion for the maximum value have been measured frequently. The high copper values were found at most sites on Murray and Anderson Creeks, and the Nicomekl River at the same time period. This implies that high copper values are natural throughout the watershed at the same time.

Swain and Alexander (3) had reported lead values at Site 1100003 to range from <0.001 mg/L to 0.012 mg/L (0.002 mg/L median). This is typical of values found at most of the other Nicomekl River sites (Table 19). The highest value for Anderson Creek and Murray Creek was 0.005 mg/L at Site 1100017 (Table 21) and Site 0300064 (Table 22), respectively. Water quality criteria are 0.05 mg/L for raw drinking water (6,84), 0.004 mg/L for aquatic life with minimum hardness of 40 mg/L $^{(84)}$ , 0.2 mg/L for irrigation (76,84), and 0.1 mg/L for wildlife and livestock water $ing^{(76,84)}$ . This indicates that lead values in only the Nicomekl River exceeded the most sensitive criteria for the protection of aquatic life on certain occasions. Downstream from the Township of Langley, at Sites 0300061 and 1100006, total lead values over a period of years exceeded the aquatic life criterion on one of thirty occasions and two of seven occasions, respectively. The criterion was exceeded in two of seventeen samples collected at Site 1100005, two of thirty-seven at Site 1100003, three of thirty-nine at Site 0300062 and was exceeded in one of thirty-five samples collected at Site 0300060.

Stormwater could be entering the Nicomekl River from Langley. Therefore, a provisional objective is proposed for total lead to protect aquatic life. The objective is that the average value should not exceed 0.005 mg/L and the maximum should not exceed 0.010 mg/L. The objective (infrequently exceeded in the past) applies to the Nicomekl River, to samples collected outside the initial dilution zones of effluents described in Section 2.4.1.

All mean total iron values sampled at sites on the Nicomekl River, Anderson and Murray Creeks exceeded the criterion of 0.3 mg/L for drinking water use  $^{(7)}$  and for the protection of aquatic life  $^{(76)}$ , but were less than the criterion for irrigation use  $(5 \text{ mg/L to } 20 \text{ mg/L})^{(76)}$ . The criterion for drinking water is based on aesthetic considerations, not health. The maximum values of 5.5 mg/L at Site 0300062 on the Nicomekl River and of 8.6 mg/L at Site 1100022 on Murray Creek exceeded the criterion of 5 mg/L for continuous irrigation use on all soils.

Since these high values are frequent and occur throughout the watershed, we believe that they are naturally occurring. An example of the frequency of not meeting the drinking water and aquatic life criteria is given for Site 0300062 where 18 of 37 samples collected over a period of years exceeded the criteria. All samples from Murray Creek (Table 22) were at or exceeded the criteria level. At Site 0300063 on Anderson Creek (Table 21), 8 of 16 samples exceeded the criteria. Therefore, no objective for iron is proposed.

Swain and Alexander<sup>(3)</sup> reported total manganese values from 0.03 to 0.12 mg/L at Site 1100003 on the Nicomekl River. Values at Site 0300061 were as high as 0.27 mg/L (Table 19); however, most values at most sites have not exceeded 0.15 mg/L. The maximum value in Anderson Creek was 0.07 mg/L at Site 0300063. On Murray Creek, the maximum value was 0.05 mg/L at Site 0300064.

Working water quality criteria are 0.05 mg/L for drinking water (due to aesthetic considerations) $^{(7)}$ , 0.1 mg/L to 1.0 mg/L for the protection

of aquatic life $^{(52)}$ , and 0.2 mg/L to 10 mg/L for irrigation $^{(76)}$ . Thus, the drinking water criterion was equalled or slightly exceeded by maximum values in Anderson Creek and the Nicomekl River. The criteria to protect aquatic life were also exceeded in the Nicomekl River. The most restrictive irrigation criterion was exceeded at one site on the Nicomekl River.

The drinking water criterion was exceeded in nine of eighteen samples collected at Site 0300063 on Anderson Creek and on three of eighteen occasions at Site 0300064. The criterion for the protection of aquatic life was exceeded at Sites 0300061 (four of thirty-five occasions) and 1100003 (three of thirty-five occasions) on the Nicomekl River.

The highest total zinc values were recorded at Site 1100006, where a value of 0.13 mg/L was recorded (Table 19). Values as high as 0.11 mg/L were recorded at Sites 1100003 and 1100007. The highest value in Murray Creek was 0.12 mg/L at Site 1100022, while the highest value in Anderson Creek was 0.11 mg/L at Site 1100017. The fact that similar maximum zinc values were found throughout the watershed indicates that the zinc occurs naturally.

Working water quality criteria are 0.03 mg/L to protect aquatic life  $^{(76)}$ , 1.0 mg/L for irrigation water  $^{(76)}$ , 50 mg/L for livestock  $^{(76)}$  and 5 mg/L for drinking water  $^{(7)}$ . Thus the criterion for aquatic life can be exceeded at times in all three water bodies. However, levels were well below criteria to protect other uses.

# 4.4.2.2 Metal Values in Sediments

Sediment samples were collected from three downstream sites (1100003, 0300060, and 1100005) and three upstream sites (1100006, 0300061, and 1100007) on the Nicomekl River. The data are summarized in Table 20. These values were compared to values for the Serpentine River, the Fraser River (21), B.C. Lake sediments (22), and urban stormwater sediments (24) in Table 23.

Concentrations of all of the metals in Nicomekl sediments were similar to concentrations found at the other locations (Table 20). Total cadmium, copper, lead, and zinc values in sediments from the Nicomekl River were lower than those from a stormwater catchment. This may reflect the lack of urban development within this drainage basin. Total cadmium, lead, and manganese values were similar in sediments from both the Serpentine and Nicomekl River. Total cobalt values in sediments from the Nicomekl River were marginally higher than those from the Serpentine River. These higher cobalt values may be attributed to unconsolidated deposits from a cobalt-bearing local granite source (2) within the drainage basin.

#### 4.4.3 NUTRIENTS

# 4.4.3.1 Nutrient Values in the Water Column

Tables 19, 21 and 22 contain nitrogen values analyzed at eleven sites on the Nicomekl River and two sites on both Anderson Creek and Murray Creek. The majority of the samples were collected quarterly (February, May, September, November) over a period of years.

The highest ammonia value in the Nicomekl River was 0.525 mg/L at Site 0300062 (Table 19). Corresponding to this was a pH of 7.4 and a temperature of  $5.8^{\circ}\text{C}$ . This ammonia concentration was well below the provincial water quality criteria for maximum and average ammonia concentrations to protect aquatic life<sup>(69)</sup>. Similarly, maximum values at other sites on the Nicomekl River, and in Anderson Creek and Murray Creeks were below the criteria.

A concern can exist for ammonia generated from animal wastes. Therefore provisional water quality objectives are proposed for total ammonia in the Nicomekl River, Murray and Anderson Creeks. Total and average values should not exceed values listed in Tables 29 and 30 which are the provincial criteria for ammonia (69). The objectives apply in all areas, except initial dilution zones of effluents described in Section 2.4.1. The average

value is to be calculated from a minimum of five samples collected in a 30-day period.

The maximum nitrite value in the Nicomekl River was 0.127 mg/L in October 1982 at Site 0300060. A value of 0.11 mg/L was recorded also in the same month. At chloride concentrations of 3 mg/L (minimum value in Nicomekl River), B.C. criteria for nitrite are 0.04 mg/L as a 30-day average value and 0.12 mg/L as a maximum value  $^{(69)}$ . Except for the high levels, recorded nitrite values did not exceed the 30-day average value. Maximum nitrite values in Anderson and Murray Creeks were less than both criteria.

Owing to the seasonal use of the Nicomekl River by salmonid populations, the fact nitrite can be formed when ammonia is oxidized, and the timing of values exceeding some criteria for the protection of aquatic life, provisional objectives for nitrite are proposed which are the same as the provincial nitrite criteria (69). The objectives are as follows:

Chloride Concentration mg/L	Maximum Nitrate Concentration mg/L	30-day Average Nitrite Concentration mg/L
<2	0.06	0.02
2-4	0.12	0.04
4-6	0.18	0.06
6-8	0.24	0.08
8-10	0.30	0.10
>10	0.60	0.20

The objectives apply to the Nicomekl River, Anderson and Murray Creeks, except in initial dilution zones described in Section 2.4.1. The average value is to be calculated from a minimum of five samples collected over a 30-day period.

All nitrate values in Anderson and Murray Creeks and in the Nicomekl River were below the B.C. criterion of 10 mg/L for raw drinking water supplies (69). This is the most sensitive use with respect to nitrate.

Phosphorus concentrations were highest at Sites 1100002, 1100003 and 0300060 near the mouth of the Nicomekl River and decreased at sites further upstream and on Anderson and Murray Creeks (Tables 19,21,22).

The highest dissolved ortho and total dissolved phosphorus values were 0.394 mg/L and 0.437 mg/L, respectively, at Site 1100005 in the Nicomekl River. These quantities of phosphorus are sufficient to cause heavy algal production, assuming that phosphorus is the limiting nutrient. Other factors which would have a bearing on algal growth would be availability of suitable substrate, adequate light, and the river velocity. High phosphorus values were also found in Murray Creek, but not at the same magnitude in Anderson Creek. Nevertheless, phosphorus concentrations were at high enough levels to cause algal growth, given the previous assumptions.

The water quality criterion for nutrients and algae in streams limits the chlorophyll-a levels to 100 mg/m² chlorophyll-a for the protection of aquatic life<sup>(9)</sup>. No chlorophyll-a data were available for the Nicomekl River, Anderson Creek, or Murray Creek. Since phosphorus and nitrogen can enter the water from agricultural practices, a provisional objective of 100 mg/m² chlorophyll-a is proposed, applicable to naturally-growing periphytic algae. This objective is to be evaluated from the average of at least five replicates collected on the same date. The objective applies to all parts of Anderson and Murray Creeks and the Nicomekl River, except initial dilution zones of effluents described in Section 2.4.1. This objective, in the absence of data, may have to be a long-term objective. It is recommended that chlorophyll-a monitoring of natural substrates be conducted on a routine basis to determine if this proposed objective is appropriate.

# 4.4.3.2 Nutrient Values in Sediments

Sediment samples were analyzed from three sites (1100003, 1100005, 1100006) on two occasions (June, July) in 1979 and from three sites (0300060, 0300061, 1100007) in 1983 (Table 20). The highest Kjeldahl nitrogen (3 mg/g), total phosphorus (1.4 mg/g), organic carbon (26 mg/g) and inorganic carbon (10 mg/g) values were found at Site 0300060, towards the mouth. This indicates that a portion of the nutrients which are entering the river may be settling out.

Among the Serpentine, Nicomekl, and Little Campbell Rivers, nutrients in sediments were similar (see Section 2.4.3.2).

#### 4.4.4 OXYGEN AND OXYGEN-CONSUMING MATERIALS

The lowest recorded dissolved oxygen value on the Nicomekl River was 0.8~mg/L at Site 1100003 (Table 19), near the mouth. Other low minimum values were near the mouth: 1.3~mg/L at Site 0300060, 2.5~mg/L at Site 1100002, and 5.4~mg/L at Site 1100005. Minimum values were higher further upstream near the headwaters. Criteria for the protection of salmonids (83) are a minimum dissolved oxygen value of 8.0~mg/L to provide for no impairment and 6~mg/L which could cause slight impairment.

Percent saturation values became more variable near the mouth. At Site 1100003, values ranged from 8.8% to 263.8%. Such wide ranges in percent saturation will cause extreme stress in aquatic life. They are likely associated with high nutrient and algal growth levels which have arisen due to agricultural inputs.

Near the mouth of the Nicomekl River, between May and August of 1979, values ranged from 2.5 mg/L to 6.7 mg/L at Site 1100002, while 5.2 mg/L was recorded at Site 1100003 and 5.4 mg/L at Site 1100005. Between September and November, values ranged from 4.0 mg/L to 4.3 mg/L at Site 1100002, and from 0.8 mg/L to 1.9 mg/L at Site 1100003. A value of 1.3 mg/L was recorded

in November 1982 at Site 0300060. Salmonid migration takes place from August to November, with spawning from November to May. Thus, low oxygen levels existed at critical times of the year.

Fish spawning (Figure 5) occurs in the upper reaches of the Nicomekl River near Sites 0300062 and 1100007. At these sites, dissolved oxygen levels have ranged from 7.8 to 14.1 mg/L and 8.7 to 13.0 mg/L, respectively. These levels provide virtually no impairment to salmonids. Since spawning occurs after about November 1, there is likely sufficient dissolved oxygen to meet spawning requirements.

The lowest dissolved oxygen level in Murray Creek was 6.0 mg/L at Site 1100022 (Table 22). In Anderson Creek, the lowest value was 8.25 mg/L at Site 1100017 (Table 21). Percent saturation values ranged from 62.8% at Site 1100022 to 123.5% at Site 0300064 in Murray Creek. In Anderson Creek, values ranged from 76.4% to 159.3% at Site 1100017. Both the dissolved oxygen and percent saturation values are not as variable in Anderson and Murray Creeks as in the Nicomkel River.

Salmonids in Murray Creek were subjected to from slight to no impairment at all times ( $\ge 6.0~\text{mg/L}$ )(83), while those in Anderson Creek were not affected ( $\ge 8.0~\text{mg/L}$ )(83). As well, they were not subjected to the same degree of stress as was aquatic life in the Nicomekl River where percent saturation values fluctuated more widely. This is likely due to the fact that most inputs of nutrients and oxygen-demanding materials from all the water bodies are concentrated nearer the mouth of the Nicomekl River, and thus this is where the greatest impact from lack of oxygen occurs.

There are therefore periods of real concern with dissolved oxygen levels in the Nicomekl River. Any proposed objective must recognize this, and allow for improvements to be made in the watershed which will be reflected in higher dissolved oxygen concentrations in the future.

Therefore, staged provisional objectives are proposed for the minimum dissolved oxygen concentration in the Nicomekl River. To protect salmonids initially, the minimum dissolved oxygen concentration should not be less than 6.0 mg/L. Once this minimum is consistently met as it appears to be in Anderson and Murray Creeks, the objective for a minimum value should be increased to 8.0 mg/L which is also the proposed objective for Anderson and Murray Creeks. In addition, between November and May in spawning areas of the Nicomekl River, Murray and Anderson Creeks, the minimum dissolved oxygen concentration should not be less than 11.0 mg/L when larvae or alevin are present. These objectives apply to insitu measurements or discrete samples taken outside the initial dilution zones of effluents, as described in Section 2.4.4.

There are no permitted point-source discharges within this drainage basin which impact dissolved oxygen values. Remedial activities would therefore relate to non-point sources. Remedial action to maintain oxygen concentrations within a suitable range for the protection of aquatic life is not presently being undertaken (40). However, results of aeration studies on the Serpentine River may also be applicable to the Nicomekl River system.

#### 4.4.5 BACTERIOLOGICAL QUALITY

Fecal coliform data within the Nicomekl River system were collected from four sites (1100006, 1100007, 1100017, 1100022) in April 1979, five different sites (0300060, 0300061, 0300062, 0300063, 0300064) between 1974 and 1982, and at one site (1100001) in March 1976. Values collected in March 1976 ranged from 130 to 490 MPN/100 mL (mean: 305 MPN/100 mL, median: 330 MPN/ 100 mL) at Site 1100001 near the mouth of the river.

Fecal coliform data were also analyzed for Sites 0300060, 0300061, 0300062, 0300063, and 0300064 between 1974 and 1982. These data indicate variable values at Site 0300062 near the headwaters, ranging from 207 to >24 000 MPN/100 mL. Values then decreased downstream to Site 0300060, near

the mouth of the Nicomekl River where the range was <2 to 9200 MPN/100 mL. The main water uses on the Nicomekl River, Anderson Creek, and Murray Creek are irrigation, and livestock watering. The B.C. Ministry of Health not approve of the use of these waters as a drinking water supply. criteria for confined livestock limit the 90th percentile of fecal coliforms to 10 MPN/100 mL for water used after disinfection only (7,30)90th percentile of fecal coliform data, collected over a ten-year period at Sites 0300060 (2 400 MPN/100 mL), 0300061 (2 400 MPN/ 100 mL), 0300062 (7 680 MPN/ 100 mL) on the Nicomekl River and for a four-year period at Sites 0300063 (880 MPN/100 mL) and 0300064 (1 540 MPN/ 100 mL) on Anderson Creek and Murray Creek, respectively, exceeded this criterion level and complete water treatment plus disinfection would be needed before use. A long-term objective of reducing fecal coliform concentrations to a geometric mean below 200 MPN/100 mL to protect livestock watering is proposed. Geometric mean values are to be calculated from a minimum of five weekly samples collected over a 30-day period. Drinking water license holders should be advised that, at present, complete water treatment may be necessary for Nicomekl River, Anderson Creek, and Murray Creek water prior to use and that the B.C. Ministry of Health does not approve of the use of these waters for human consumption. Further monitoring for fecal coliforms and streptococci should be conducted to determine the present levels and sources of fecal contamination in the basin. This will indicate the feasibility of attaining the long-term objective. This objective applies to areas outside initial dilution zones of effluent as outlined in Section 2.4.1.

The B.C. criterion for general irrigation (geometric mean value of 1000 MPN/ 100 mL)  $^{(30)}$  was exceeded at Site 0300062 between May and August (1730 MPN/100 mL), and between September and December (1347 MPN/100 mL); and at Site 0300061, downstream from the Township of Langley, between September and December (1056 MPN/100 mL). This criterion is similar to the approved water quality objectives for the Fraser River  $^{(82)}$ . There is a more restrictive B.C. criterion (geometric mean of  $\leq$ 200 MPN/100 mL) for irrigation of crops eaten raw. There are twelve

permitted irrigation withdrawals in the upper reach of the river, from the headwaters of the Nicomekl River to Site 0300061. Therefore, it is proposed that a provisional objective should be established for fecal coliforms in the Nicomekl River basin. In the long-term, the geometric mean fecal coliform values should remain below 200 MPN/ 100 mL at all times. In the short-term, the geometric mean should be ≤1000 MPN/100 mL and the maximum not greater than 4000 MPN/ 100 mL in the April to October period. These objectives apply to discrete samples or the mean of discrete samples taken anywhere in the Nicomekl River basin outside the initial dilution zones of effluents, described in Section 2.4.1.

#### 4.4.6 SOLIDS

The maximum recorded suspended solids concentration in the Nicomekl River was 87 mg/L at Site 0300062 (Table 19). In Anderson Creek the maximum value was 42 mg/L at Site 0300063 (Table 21), while it was 55 mg/L at Site 0300064 in Murray Creek (Table 22). Such values would provide a moderate level of protection to aquatic life  $^{(26)}$ . The most restrictive criteria to protect aquatic life are that the induced suspended solids concentration should not exceed 10 mg/L when background levels are less than 100 mg/L, and should not be more than 10% of background when background is greater than 100 mg/L $^{(14)}$ .

These criteria are proposed as the provisional water quality objective for suspended solids in the Nicomekl River, and Anderson and Murray Creeks. The objective is that induced suspended solids concentrations should not exceed 10 mg/L when background levels are less than 100 mg/L, nor should be more than 10% of background when background is greater than 100 mg/L. These objectives apply to discrete samples collected outside the initial dilution zones of effluents as described in Section 2.4.1.

In addition, to protect salmonid spawning areas, there should be no significant (95% confidence level) induced benthic sedimentation on the basis of accumulation by weight for particles <3 mm.

Maximum turbidity values were measured at Site 0300062 in the Nicomekl River (56 NTU), at Site 1100022 in Murray Creek (88 NTU), and at Site 0300063 in Anderson Creek (11 NTU). Since arithmetic mean values at all sites in the Nicomekl River and Murray Creek exceeded the water quality criterion of 5 NTU (maximum) for drinking water supplies, filtration or equivalent would be required in most cases prior to drinking. In Anderson Creek, the values met this criterion in most of the samples (17 of 19).

Higher turbidity values were recorded between November and May in the Nicomekl River. These increases are likely due to increased stormwater runoff during this time of the year.

B. C. criteria exist for induced turbidity<sup>(14)</sup>. These will be used as the proposed water quality objectives for turbidity. The objective is that induced levels should not exceed 5 NTU when background is less than 50 NTU nor should they increase by more than 10% when background is greater than 50 NTU. The objective applies to the Nicomekl River, Anderson Creek, and Murray Creek.

Dissolved solids, as indicated by their direct measurement or by measuring specific conductivity, can reach high values due to saltwater intrusion as far upstream in the Nicomekl River as Site 1100006. At this site, and further upstream, all values have been less than 500 mg/L (Table 19). Such was also the case for Anderson Creek (Table 21) and Murray Creek (Table 22). Working water quality criteria are 500 mg/L for drinking water supplies  $^{(7)}$ , 1000 mg/L for livestock watering  $^{(76)}$ , and from 500 to 3500 mg/L for irrigation, depending on the soil and  $\operatorname{crop}^{(76)}$ . Thus there are times in the lower reaches of the Nicomekl when dissolved solids exceed these criteria and the water may not be suitable for these uses.

# 4.5 RECOMMENDED MONITORING

Several calculations in this report have predicted the maximum potential impact of agriculture. A site-specific study should be designed for sites near the headwaters and near the mouth of the Nicomekl River to

determine the local impact of agricultural operations. The exact study design should be considered in the future. As well, a site-specific study should be undertaken near Langley to determine stormwater impacts. The exact study design should also be considered in the future.

Several water quality objectives have been proposed. To determine whether these are being achieved, it is recommended that a minimum of two sites on the Nicomekl River, one upstream from Langley and one near the mouth, as well as a site on each of Anderson Creek and Murray Creek, be selected. Additional sampling sites should be used when the objectives are exceeded. A minimum sampling frequency of five samples in a 30-day period is recommended. Samples should be analyzed for pH, dissolved oxygen, temperature, specific conductivity, ammonia, nitrite, nitrate, dissolved orthophosphorus, total phosphorus, dissolved and suspended solids, fecal coliforms, fecal streptococci, turbidity, periphyton chlorophyll-a, and dissolved and total copper, iron, manganese, and lead. However, resources and other regional priorities will determine the exact program.

#### 4.6 CONCLUSIONS

Operations discharging effluents within the Nicomekl drainage basin include two landfills, one cooling water discharge, eight soil disposal systems, and a bulk petroleum storage plant. Remedial activities (installation of a leachate collection system) to improve water quality in a tributary to the Nicomekl River have been initiated at the landfill operated by the Township of Langley (PR 1898). The leachate collection system is expected to discharge to a sewer system in the future. Mansonville Plastics Ltd. may discharge uncontaminated cooling water to the Nicomekl River. It has been determined that the dilution rate is sufficient within the Nicomekl River for effluent discharged at the permitted temperature (<32°C) not to impact the water quality. However, it is not known what actual effect this effluent has on the Nicomekl River as no data on actual temperatures of the discharged cooling water have been recorded. Several soil disposal systems and one landfill operation either discharged to areas at a sufficient dis-

tance from the river or have the effluent trucked to a local municipal sewage treatment facility in Langley. The largest impact on the Nicomekl River system is likely from agricultural operations.

The pH of water in the basin was naturally high, and it was well-buffered to acidic inputs. Some higher metal values were recorded, but these were likely naturally occurring. High lead values may originate from stormwater discharges. Ammonia and nitrite values were below criteria to protect aquatic life. High levels of phosphorus and the consequent heavy algal growth, photosynthesis, respiration, and decay are believed to be responsible for low dissolved oxygen levels and wide ranges of percent saturation values. These are not as pronounced in the Upper Nicomekl River and in Murray or Anderson Creeks as they are in the lower Nicomekl River. High fecal coliform levels require a high level of treatment of water prior to human and livestock consumption. High dissolved solids near the mouth of the Nicomekl River reflect saltwater intrusion.

#### 5. BOUNDARY BAY

#### 5.1 OCEANOGRAPHY

Boundary Bay, including its northeasterly extension of Mud Bay and the southeast section called Semiahmoo Bay, is located on the south side of the Fraser-Delta area approximately 19 km south from the City of Vancouver. It is 15 km long and 4 km wide, covering an area of 6087 hectares. The Bay is rectangular in shape and faces southeast onto the southern Strait of Georgia (Figure 4).

Weir (1963)<sup>(44)</sup> used six automatic tide gauges to determine vertical movement within Boundary Bay. It was observed that the tides enter from the south and are more concentrated on the eastern side during flood tide, while being more concentrated on the western side during ebb tide.

Water flow from the Blaine area of Washington State into Boundary Bay occurred on a daily basis from January to June and in October and on a semidaily basis for the remainder of the year (43,44). This indicates that water quality within Boundary Bay may be influenced by Puget Sound water, entering during flood tides from the Blaine area. It can also be influenced by outflows from the Serpentine River, Nicomekl River, and Little Campbell Rivers. The frequency at which water flowed north from Boundary Bay to the Fraser River delta was equally distributed between daily and semi-daily occurrences during 1985

#### 5.2 WATER USES

Boundary Bay and Mud Bay sustain a crab fishery (Table 25), a herring herring fishery (Table 26), and a salmonid sports fishery. The oyster fishery has been closed to commercial and recreational harvesting since 1963 due to fecal contamination and the presence of an oyster parasite (Japanese Oyster drill Oceanebea japonica). Clams are still taken illegally.

Recreational usage in this area consists of swimming, boating, bird watching, and horseback riding along the dyke  $^{(49)}$ . Swimming occurs at Boundary Bay Beach, Centennial Beach, Crescent Beach, and near White Rock (Figure 4).

# 5.3 WASTE DISCHARGES

Swain and Alexander (3) described the sewage discharge facilities operating in Boundary Bay. Sewage pump stations in White Rock were connected to the Greater Vancouver Sewage and Drainage District (GVS & DD) in 1977. The Semiahmoo Indian Band reserve was unsewered in 1977 with possible septic tank seepage occurring at three residences; however, it is doubtful if the seepage could exert a significant impact on the water quality at Semiahmoo Bay (3). This area has not been sewered since that time.

There are no permits to discharge wastes, pursuant to the Waste Management Act, into Boundary Bay or Mud Bay.

Within the Fraser River Estuary, holding tanks with a capacity of less than 22.7 m³/d have been approved as a method of sewage treatment for houseboat communities and marinas. There are no houseboat communities in Mud Bay or Boundary Bay. Sewage from a marina near Crescent Beach is sent by way of a sanitary sewer to the Annacis Island Sewage Treatment Facility. A marina at the mouth of the Nicomekl River also discharges sewage to the Annacis Island Sewage Treatment Facility.

In summary, point source waste discharges do not enter Mud Bay or Boundary Bay, and therefore do not have a negative impact on water quality. However, non-point sources containing contaminants continue to be discharged into Boundary Bay through land-drainage pump stations. The locations of these are indicated by P1 to P5 on Figure 4. The pump stations are required to lift drainage water, which accumulates in ditches, over the dykes and into the Bay. The following discussion has been extracted from Swain and Alexander (3).

The five land-drainage pump stations were sampled during November and December of 1977. These pump stations drain predominantly agricultural-type land. Bacteriological results indicate that the lowest total and fecal coliform values occurred at P-2 (median values 270 MPN/100 mL and 33 MPN/100 mL, respectively) and the highest at P-3 (median values 5 400 MPN/100 mL and 140 MPN/100 mL, respectively). These results suggest that the pump discharges were not significant sources of fecal contamination, but this would depend on the volumes pumped. Local contamination can possibly be expected.

The following median nutrient values were reported in the drainage water: orthophosphorus (P), 0.17 mg/L; total phosphorus, 0.25 mg/L; nitrite-N, 0.018 mg/L; nitrate-N, 1.4 mg/L; and ammonia-N, 0.44 mg/L $^{(3)}$ .

Samples of land-drainage water were also collected at the same time to determine 96-hour LC50 values from static bioassays, using rainbow trout as the test species. Seven tests were performed on the drainage water from the five pump stations. These tests indicated that the waters were not acutely toxic.

Sediments adjacent to the outfalls from the drainage pump stations were analyzed at the same time for hexachlorobenzene and the polychlorinated biphenyl, Arochlor 1 260. All samples analyzed had less than one part per billion (ppb) hexachlorobenzene. However, significant concentrations of Aroclor 1 260 were found at P-3 and P-4. The concentrations at a distance of 4.6 metres west of the outfall were 1 200 ppb at P-3 and 3 800 ppb at P-4, compared to 31 ppb at background stations. Subsequent sampling by the Environmental Protection Service at these locations did not reveal high levels of polychlorinated biphenyls.

# 5.4 WATER QUALITY AND PROPOSED PROVISIONAL WATER QUALITY OBJECTIVES

Swain and Alexander (3) examined water quality data at Sites 0300070 and 0300071 within Boundary Bay for a seven-year period between 1972

and 1979. Three samples, two in 1980 and one in 1982, were collected and analyzed after this period. These data are in Table 24. The data from Swain and Alexander (3) and Table 24 are discussed in the following sections.

#### 5.4.1 pH AND BUFFERING CAPACITY

Values for pH, collected between 1973 and 1982, were relatively uniform in Boundary Bay. At Sites 0300070 and 0300071 these values ranged from 7.6 to 8.5 with a median of 8.0. All pH values reported for Sites 0300070 and 0300071 were within the criterion range for the protection of marine aquatic life  $(6.5 - 8.5)^{(31)}$ .

Median alkalinity values were about 115 mg/L at Site 0300071 and 100 mg/L at Site 0300070.

#### 5.4.2 METALS

Total cadmium values were collected at both sites in July 1982. These values were similar (0.001 mg/L at Site 0300070 and 0.0015 mg/L at Site 0300071), but outside the historical ranges reported by Swain and Alexander  $^{(3)}$  (0.0001 mg/L to 0.0005 mg/L at both sites). The working criteria for the protection of marine aquatic life (maximum value: 0.043 mg/L, 4-day mean value: 0.009 mg/L)  $^{(63)}$  were not exceeded by total cadmium values at either site in Boundary Bay.

Total manganese values between 1974 and 1980 at Sites 0300070 and 0300071 ranged from 0.002 mg/L to 0.016 mg/L and between 0.003 mg/L to 0.015 mg/L, respectively. Dissolved manganese values for the period 1973 to 1977, ranged from 0.002 mg/L to 0.006 mg/L at Site 0300070 and from 0.002 mg/L to 0.008 mg/L at Site 0300071. All values were less than the criterion level for the protection of consumers of marine mollusks  $(0.1 \text{ mg/l})^{(81)}$ .

Total chromium values between 1972 and 1982 ranged from <0.005 mg/L to 0.022 mg/L at Site 0300070 and from <0.005 mg/L to 0.02 mg/L at Site 0300071. All values were less than the working criteria (4-day average value: 0.05 mg/L, maximum value: 1.1 mg/L) for the protection of marine aguatic life  $^{(47)}$ 

Total copper values ranged from <0.001 mg/L to 0.04 mg/L at Site 0300070 (period of record: 1972-1982) and from <0.001 mg/L to 0.007 mg/L (period of record: 1973-1982). B.C. criteria for total copper, in an estuarine or marine environment, for the protection of aquatic life are: 30-day average value 0.002 mg/L, maximum value 0.003 mg/L $^{(66)}$ . Eleven of fifty-seven samples collected at Site 0300070 exceeded the maximum criterion. At Site 0300071, the maximum criterion was exceeded in nine of fifty samples collected over the period of record. It is not known if these copper levels are natural.

Total lead values ranged from <0.001 to 0.032 mg/L at Site 0300070 from 1972 to 1982 and from <0.001 mg/L to 0.007 mg/L at Site 0300071 from 1973 to 1982. The B.C. criterion for the protection of aquatic life in a marine or estuarine environment limits total lead values to a 30-day average value of 0.002 mg/L and a maximum value of 0.14 mg/L $^{(84)}$ . The 30-day average criterion was not exceeded by the median value of all samples collected and the maximum criterion was never reached.

Total mercury values were measured between 1978 and 1982 at Sites 0300070 and 0300071 (values ranged from < 0.00005 mg/L to 0.00068 mg/L and from <0.00005 mg/L to 0.00083 mg/L, respectively). Dissolved mercury values, collected between 1973 and 1978, ranged from <0.00005 mg/L to 0.00025 mg/L at Site 0300070 and from <0.00005 mg/L to 0.00033 mg/L at Site 0300071. The working criteria for the protection of consumers of fish  $(0.0001 \text{ mg/L})^{(12)}$  and the protection of aquatic life  $(0.0002 \text{ mg/L})^{(12)}$  were exceeded in one (0.00083 mg/L) of twelve samples at Site 0300071 and in six (value range: 0.0001 mg/L - 0.00068 mg/L) of

sixteen samples from Site 0300070. It is recommended that monitoring be conducted to determine the cause of these elevated values. Such monitoring would examine anthropogenic sources which may contribute mercury to this area and determine the frequency at which values greater than the criteria occur. It should include sampling of resident aquatic species to see if criteria for human consumption are achieved.

Nickel values at both sites were at or below the detection level of 0.01 mg/L.

Total zinc values at Site 0300070, for the period 1972 to 1982, ranged from <0.005 mg/L to 0.13 mg/L (twenty-six of fifty-seven samples were greater than the 0.005 mg/L detection level). At Site 0300071, for the period 1973 to 1982, total zinc values ranged from <0.005 mg/L to 0.05 mg/L (seventeen of forty-five values were greater than the 0.005 mg/L detection level). The working criteria for the protection of marine aquatic life limit total zinc values to a 30-day average value of 0.058 mg/L and a maximum value of 0.17 mg/L $^{(68)}$ . The average criterion was exceeded in one (0.13 mg/L) of fifty-seven samples collected at Site 0300070, while all values at Site 0300071 were less than the average criterion.

High metal values occurred more frequently at Site 0300070 than at Site 0300071. This may be attributed to the proximity of the site to the City of White Rock, Blaine, and Drayton Harbour. Urban discharges, including stormwater runoff, may contribute to higher metal concentrations at this site.

#### 5.4.3 NUTRIENTS

Total ammonia nitrogen values ranged from <0.01 mg/L to 0.09 mg/L at Site 0300070 between 1972 and 1980. At Site 0300071, between 1973 to 1980, total ammonia values ranged from <0.01 mg/L to 0.07 mg/L. All values were less than the B.C. criteria for marine aquatic life of 2.5 mg/L as a maximum value and <1.0 mg/L as an average value  $^{(69)}$ .

Nitrogen concentrations (nitrate-nitrogen, nitrite-nitrogen, and organic nitrogen) remained generally constant over the period of record at Sites 0300070 and 0300071 in Boundary Bay. Nitrate-nitrogen values at Site 0300071 were slightly higher than at Site 0300070 (value range at Site 0300070: <0.02 mg/L to 0.41 mg/L, mean: 0.15 mg/L; at Site 0300071: <0.02 mg/L to 0.57 mg/L, mean: 0.24 mg/L). One nitrite value of 0.079 mg/L was reported for Site 0300071 (Table 24). Although there are no B.C. criteria for nitrite to protect marine life, a B.C. criterion for freshwater of a maximum of 0.6 mg/L nitrite is to be used for water with a chloride concentration >10 mg/L. Since the waters of Boundary Bay have far in excess of 10 mg/L chloride, the single high nitrite value is not of concern.

Orthophosphorus values ranged from <0.003 mg/L to 0.077 mg/L (mean value: 0.04 mg/L) at Site 0300070 and from <0.003 mg/L to 0.088 mg/L (mean value: 0.04 mg/L) at Site 0300071 for the period 1973 to 1982. Total dissolved phosphorus values ranged from 0.014 mg/L to 0.076 mg/L (mean value: 0.05 mg/L) at Site 0300070 and from 0.013 mg/L to 0.075 mg/L (mean value: 0.06 mg/L) at Site 0300071. Seasonally, values for orthophosphorus and total dissolved phosphorus decreased between May and September and increased between October and April.

### 5.4.4 OXYGEN AND OXYGEN-CONSUMING MATERIALS

Dissolved oxygen values ranged at Site 0300070 from 5.8 mg/L to 12.1 mg/L (mean: 9.0 mg/L) and at Site 0300071 from 6.3 mg/L to 14 mg/L (mean: 8.8 mg/L). These values became less variable in more recent times at both sites (value range for the period 1972-1975 at Site 0300070: 5.8 mg/L to 12.1 mg/L, for the period 1976-1982: 6.6 mg/L to 10.8 mg/L; value range for the period 1973-1975 at Site 0300071: 6.3 mg/L to 14 mg/L, for the period 1976-1980: 6.6 mg/L to 10.5 mg/L). The working marine criteria to provide moderate to high protection for non-anadromous fish species and anadromous species with salmonids are that dissolved oxygen levels should be from 6.75 mg/L to 8.75 mg/L and from 6.5 mg/L to 9.0 mg/L, respectively<sup>(10)</sup>. At Site 0300070, the level fell below the minimum of

6.5 mg/L for a moderate level of protection only in one of fifty-five samples collected over a ten-year period. At Site 0300071, the level fell below the minimum value of 6.5 mg/L only in one of forty-five samples collected over a seven-year period. However, the mean dissolved oxygen levels in Boundary Bay were about the same as the minimum criteria for a high level of protection, suggesting that a high level of protection was achieved only about 50% of the time.

Since oxygen-consuming materials are entering the Bay from non-point discharges and through the tributaries, a provisional objective is proposed for dissolved oxygen. The objective is that the minimum dissolved oxygen level should not be less than 6.5 mg/L. A second long-term objective of a minimum of 9.0 mg/L is also proposed. The objectives apply to in-situ measurements or discrete samples in all areas of the Bay, except in initial dilution zones of effluents. These excluded initial dilution zones in the Bay are described as extending 100 m horizontally in all directions, but not to exceed 25% of the width of the waterbody.

#### 5.4.5 SOLIDS

Suspended solids values at both sites were low, with mean values of about 6 mg/L and a maximum of 29 mg/L at Site 0300070 for the period 1972-1982. At Site 0300071 for the period 1973 to 1982, the maximum value was 31 mg/L. Dissolved solids values ranged from 24 500 mg/L to 33 200 mg/L (mean value: 30 561 mg/L) at Site 0300070 for the period 1974 to 1978 and from 25 200 mg/L to 33 600 mg/L (mean value: 30 836 mg/L) at Site 0300071 for the period 1974 to 1978. These values indicate that suspended solids values and dissolved solids values were relatively uniform throughout Boundary Bay.

Turbidity values ranged at Site 0300070 from 0.4 NTU to 6.2 NTU (mean value: 1.45 NTU) and from 0.4 NTU to 8.4 NTU (mean value: 1.65 NTU) at Site 0300071.

Water quality objectives for suspended solids and turbidity are proposed for the protection of aquatic life in Boundary Bay. Induced turbidity values are to be less than 5 NTU while induced non-filterable residue values are to be less than 10 mg/L when background levels are less than 100 mg/L suspended solids or 50 NTU turbidity. If background levels exceed 100 mg/L suspended solids or 50 NTU turbidity, induced levels should not increase by more than 10% of background. These objectives apply in the Bay, except in initial dilution zones of effluents, described in Section 5.4.4.

#### 5.4.6 BACTERIOLOGICAL QUALITY

Fecal coliform values, collected between 1973 and 1982, ranged from <2 MPN/100 mL to 790 MPN/100 mL at Site 0300071 and from <2 MPN/100 mL to 130 MPN/100 mL at Site 0300070. At Site 0300070, five of twenty-eight samples collected were greater than the detection levels (<2 MPN/100 mL, <20 MPN/100 mL). The detectable values ranged from 2 MPN/100 mL to 130 MPN/100 mL with a median value of 5 MPN/100 mL and 90th percentile value of 54 MPN/100 mL. Sixteen of twenty-six samples collected at Site 0300071 were less than the detection levels of 2 MPN/100 mL and 20 MPN/100 mL. In the remaining ten samples fecal coliform values ranged from 2 MPN/100 mL to 790 MPN/100 with a median value of 5 MPN/100 mL and 90th percentile value of 23 MPN/100 mL. Only one of fifty-four values exceeded the recreation criterion of 200 MPN/100 mL(70).

Kay<sup>(49)</sup> conducted a sanitary survey of Boundary Bay, Mud Bay, and Crescent Beach in 1976 (Figure 4). This survey measured fecal coliform levels from thirty-three marine stations during the three-month period from March to May," a time when hydrographic and pollution conditions were not expected to be the worst<sup>(49)</sup>. Twenty of "33 marine stations met the shellfish growing water standard. Of the remaining 13 sampling stations which were classified as unacceptable, 7 exceeded the standard at the median level and 6 exceeded the standard at the 90th percentile level<sup>(49)</sup>. Values at individual stations had ranged from <2 to 540 MPN/100 mL, with median values from <2 to 205 MPN/100 mL and 90th percentile from <4 to

350 MPN/100 mL. Thus, it was recommended that the Bay should stay closed to shellfish harvesting. However, it may be feasible to use Boundary Bay for growing shellfish commercially if they are removed to "clean" water prior to harvesting. Exact timing of such activities would have to be carefully investigated.

Fecal coliform data collected by the Ministry of Health at eleven beach sites (Figure 4) for the period of June to August between 1979 and 1985 are reported in Table 27.

The criteria for recreational beaches limit fecal coliform values to a geometric mean of less than 200 MPN/100 mL and a 90th percentile of less than 400 MPN/100 mL $^{(70)}$ . The 90th percentile for fecal coliforms at Sites 1 to 4 adjacent to the City of White Rock and Sites 10 and 11 adjacent to the City of Boundary Bay exceeded the 90th percentile criterion (values ranged from 460 MPN/100 mL to >2400 MPN/100 mL), while the geometric mean criterion was met except at Site 1 (if median values reflect geometric means).

Slightly higher values near the City of White Rock may be attributed to the discharge of water from the Little Campbell River, tidal movement (water from Drayton Harbour and Puget Sound on a daily or twice daily basis), or stormwater runoff from the City of White Rock. It is proposed that a provisional objective should be established for all beaches in Boundary Bay. The geometric mean fecal coliform value should be less than or equal to 200 MPN/100 mL and the 90th percentile should be less than or equal to 400 MPN/100 mL (April through October). In addition, a long-term objective is proposed year-round for fecal coliforms in Boundary Bay where shellfish harvesting potentially can occur. The objective is that the median fecal coliform value should not exceed 14 MPN/100 mL and the 90th percentile value should not exceed 43 MPN/100 mL. For both proposed objectives, a minimum of five weekly samples should be collected within a period of 30 days. objectives apply to areas outside the initial dilution zones of effluents, described in Section 5.4.4.

Further monitoring of fecal coliforms and fecal streptococci should be conducted to determine the present levels and sources of fecal contamination in the basin. This would indicate the feasibility of attaining these objectives.

#### 5.4.7 POLYCHLORINATED BIPHENYLS (PCBs)

PCBs can enter Boundary Bay through the Serpentine River, from Mahood Creek. Within the confines of the Bay, the immediate exposure of fish or most other aquatic life would be minimized. However, sediments carried to the Bay may settle, carrying PCBs. This can possibly be a concern to the invertebrate populations in the Bay. For this reason, a provisional objective is proposed for PCBs in sediments in Boundary Bay. The objective is that the maximum concentration in bottom surface sediments should not exceed 0.03 µg/g (dry-weight). The term PCBs applies to the sum of Arochlor 1242, 1254, and 1260. The average of at least three replicate samples taken from the same site should be used to check the objective, which does not apply in the initial dilution zones of effluents, described in Section 2.4.1.

#### 5.4.8 CHLOROPHENOLS

Chlorophenols entered Boundary Bay through the Serpentine River, from Hyland Creek in March 1984. Maximum total chlorophenol concentrations were 22.9  $\mu g/L$  four days after the spill, but these were reduced to <0.1  $\mu g/L$  six days later. Within the confines of the Bay, the immediate exposure of fish or most other aquatic life was minimal. Sediments carried into the Bay had no measurable levels (<10  $\mu g/L$ ). Clam tissue concentrations dropped from a maximum of 108  $\mu g/g$  to <10  $\mu g/g$  two months after the spill. Since there is no risk of a future discharge of chlorophenols to Boundary Bay, an objective will not be proposed.

#### 5.5 RECOMMENDED MONITORING

No additional routine monitoring is recommended for Boundary Bay or the Mud Bay area. However, bacteriological sampling should be expanded to include fecal streptococci. This would provide information to determine the levels and sources of fecal contamination within Boundary Bay and Mud Bay.

A monitoring program to assess the elevated fecal coliform values adjacent to the City of White Rock is recommended to determine coliform loadings from stormwater discharging to the beach area. The aim of such a program would be to recommend remedial procedures to reduce fecal coliform levels within the White Rock area.

A second monitoring program is recommended to determine which factors contribute to the elevated copper and mercury levels and the frequency at which these exceed the criteria for the protection of aquatic life. The program should include anthropogenic sources, including stormwater, within the Boundary Bay and Mud Bay basin and sources which are outside the basin, but impact upon the water quality of the basin. As well, resident species of aquatic life should be examined for mercury levels.

Sites and frequency of sampling for both programs should be determined in an independent study design.

#### 5.6 CONCLUSIONS

There are no permits issued pursuant to the Waste Management Act for the discharge of effluent into Boundary Bay or Mud Bay.

Boundary Bay and Mud Bay are used as a migratory corridor by salmonids and sea birds, sustain a commercial crab industry and herring industry, and are used for recreational purposes (boating, bathing). Prior to 1962, shellfish harvesting took place. Clams are still taken illegally.

Water quality in Boundary Bay was generally consistent for the period of record.

No routine monitoring is recommended in addition to existing programs conducted by the Ministry of Health and the Ministry of Environment and Parks. Two special programs are recommended to address elevated fecal coliform values adjacent to White Rock and a few elevated metal values in Boundary Bay and Mud Bay.

Elevated metal values occurred more frequently at Site 0300070 than at Site 0300071. This may be attributed to the proximity of the site to the city of White Rock, Blaine and Drayton Harbours. Urban wastes, including stormwater runoff, may contribute to higher metal concentrations at this site.

Routine bacteriological monitoring should be expanded to include fecal streptococci. A program should also be initiated to determine sources of fecal contamination within the drainage area. This would include determining fecal loadings from the Serpentine, Nicomekl, and Little Campbell River drainage areas, possibly from Puget Sound and Drayton Harbour, and anthropogenic sources resulting in point and non-point discharges to Boundary Bay and Mud Bay.

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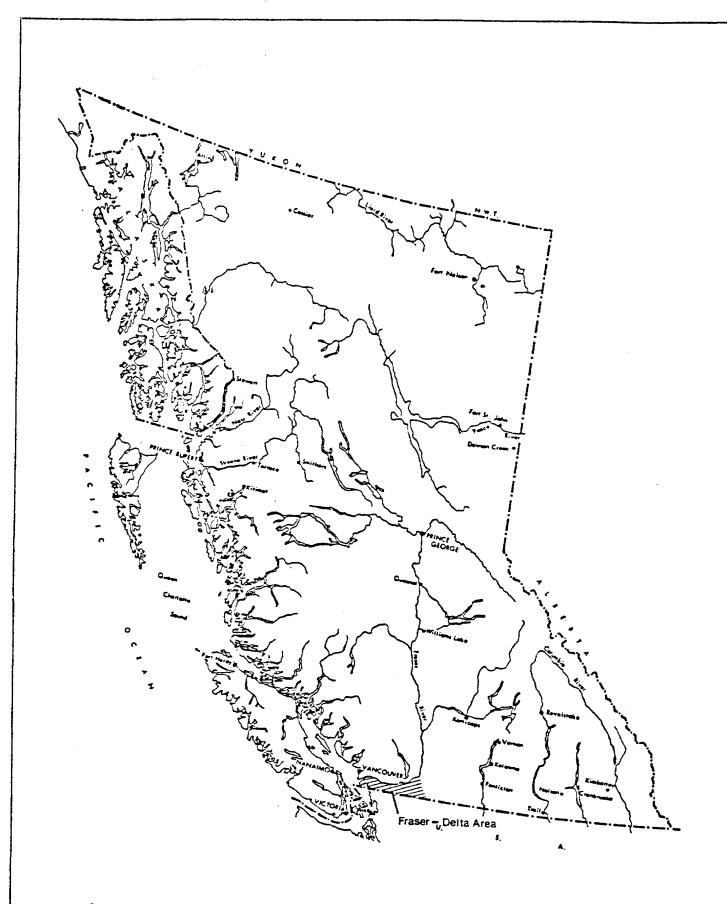


Figure 1. Location Map: FRASER-DELTA AREA

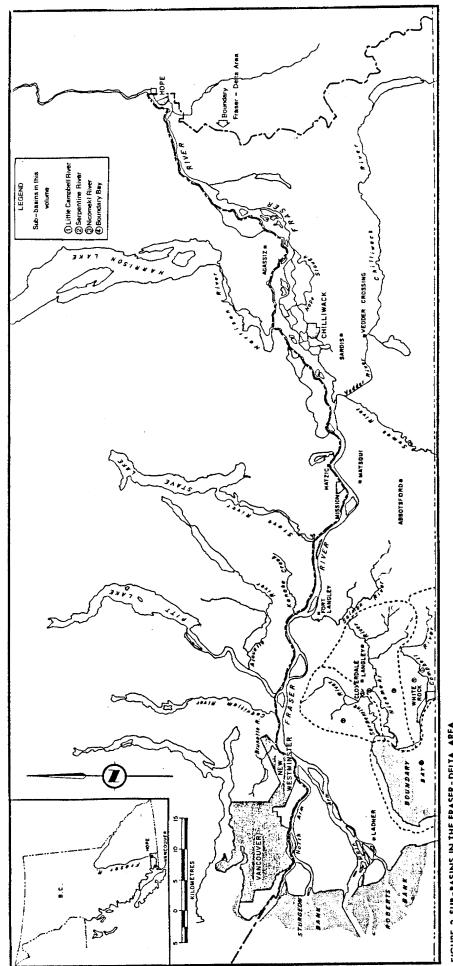


FIGURE 2 SUB-BASINS IN THE FRASER-DELTA AREA

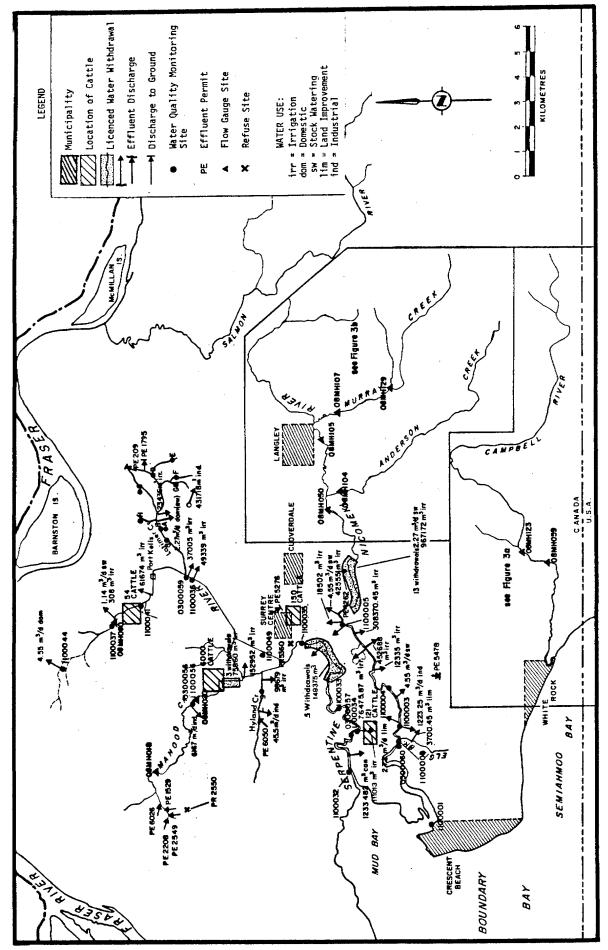


Figure 3: Location of Waste Discharges, Water Quality Monitoring Sites, Flow Gauging Sites and Licenced Water Withdrawals, Boundary Bay and its Tributaries.

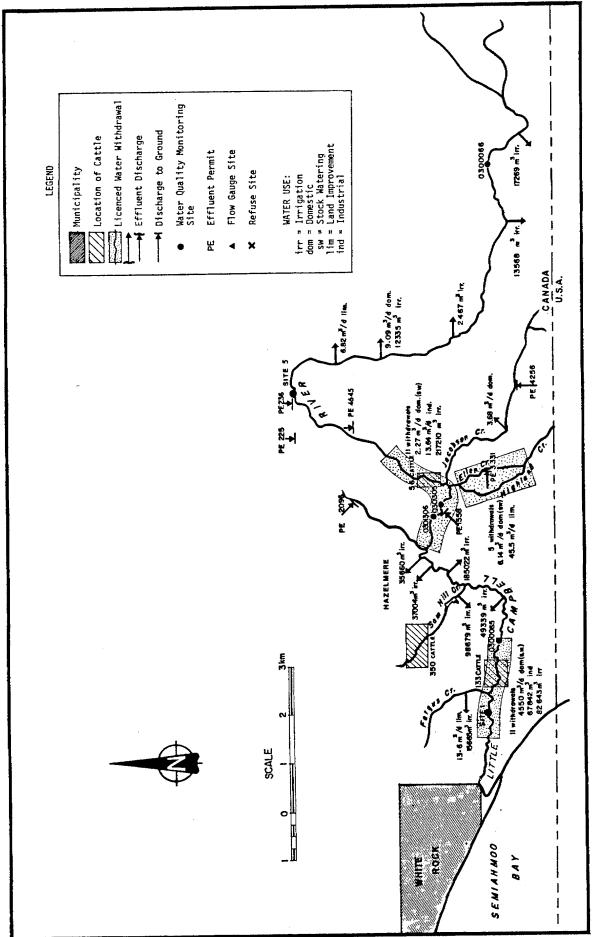


Figure 3a: Locations of Waste Discharges, Water Quality Monitoring Sites, and Licenced Water Withdrawals of Little Campbell River Watershed.

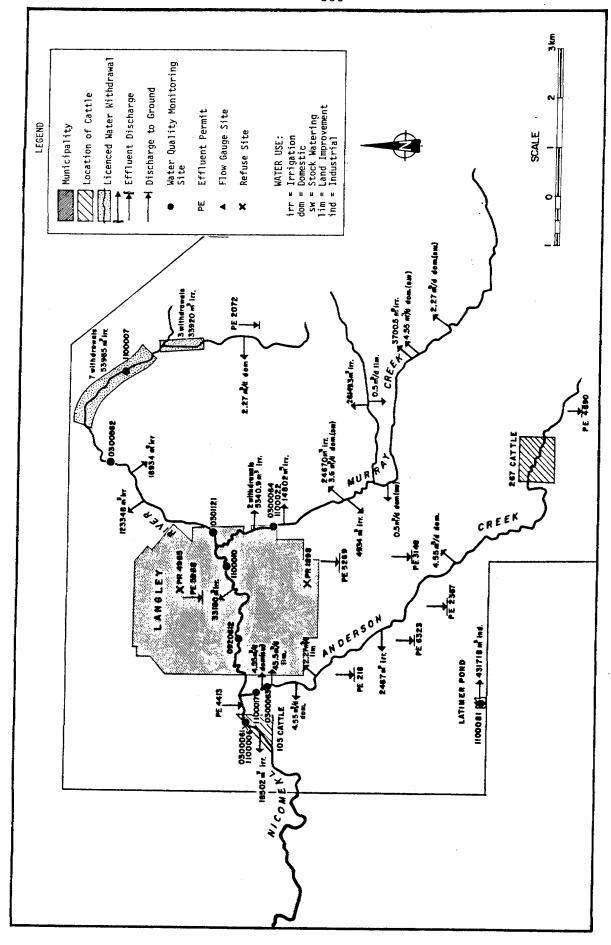
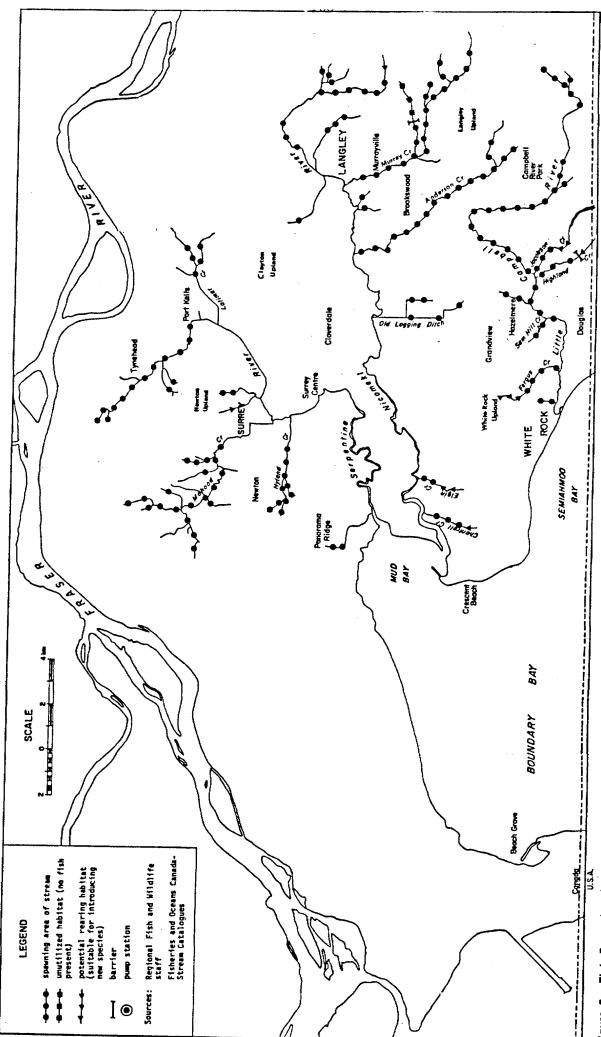


Figure 3b: Location of Waste Discharges, Water Quality Sites, and Licenced Water Withdrawals of Upper Nicomek! Watershed.

O Bacteriological Monitoring Sites-Boundary Health Unit Bacteriological Monitoring Sites EPS-KAY 1976 Water Quality Monitoring Sites LEGEND HARBOUR DRAYTON SEMIAHMOO BAY Serpentine 1100034 ● 0300070 OCEAN 300 Single Singl BAY 20 DZ N CO Ε, EAST ●0300071 BAY 0~ BOUNDARY WEST POINT ROBERTS CENTENNIAL # POINT ROBERTS COHILUKTHAN SLOUGH GE ORGIA 9 STRAIT ISLAND WESTHAM

LOCATION OF WATER QUALITY AND BACTERIOLOGICAL MONITORING SITES IN BOUNDARY BAY FIGURE 4



igure 5 Fish Spawning Areas in the Little Campbell, Nicomekl, Serpentine Drainage Basins.

TABLE 1
ESCAPEMENT RECORD FOR THE LITTLE CAMPBELL RIVER

1951	YEAR 1947 1948	CHINOOK	<u>СОНО</u> 1500 750	<u>CHUM</u> 200 200	STEELHEAD
1957       400       75         1958       400       75         1959       750       25         1960       400       75         1961       400       75         1962       200       25         1963       400       25         1964       750       75         1965       200       25         1966       400       75         1968       200       0         1969       200       25         1970       750       200         1971       750       200         1972       7500       400         1973       3500       200         1974       3500       250         1975       2200       400         1976       3500       200         1977       1500       100         1978       50       2800       50	1950 1951 1952 1953 1954 1955		750 3500 1500 1500 750 400	750 750 1500 1500 0 200	75 400 400 400
1969       200       25         1970       750       200         1971       750       200         1972       7500       400         1973       3500       200         1974       3500       250         1975       2200       400         1976       3500       200         1977       1500       100         1978       50       2800       50	1957 1958 1959 1960 1961 1962 1963 1964 1965 1966		400 400 750 400 400 200 400 750 200 400 200	75 75 25 75 25 25 75 25 75 0	200
1980 30 2500 200 1981 10 1900 149	1969 1970 1971 1972 1973 1974 1975 1976 1977 1978 1980 1981 1982	25 30 10 <b>7</b> 5	200 750 750 7500 3500 3500 2200 3500 1500 2800 1500 2500 1900 3087	200 200 400 250 400 200 100 50 75 200 149 205	10

<sup>\*</sup> Presence of one sockeye noted.

Timing of Spawning:

Arrive October October
Start Late October Early November
Peak Middle November Middle November
End Middle December Late November

Mean salmon escapements to Little Campbell River, 1962 to 1978.

YEAR	COHO	CHUM
19 <del>62-1</del> 966	390	45
1967-1971	420	90
1972-1976	4 040	290
1977-1981	2 075	105

Data From: M.J. Hancock, Marshall, D.E. 1985. Catalogue of Salmon Streams and Spawning Escapements of Statistical Area 29. Canadian Data Report of Fisheries and Aquatic Science No. 495. Department of Fisheries and Oceans.

TABLE 2

EFFLUENT DATA SUMMARY
LAFARGE CONCRETE LTD. (PE 3331)

# TRUCK WASHING OPERATION

Constituent	Period of Record	No. of Values	Maximum	Values* Minimum	Mean
Coliform:fecal	1973	1	<2	_	_
Flow	1973-1978	4	243	1.2	68.5
Oxygen:BODs	1973	1	<5	-	-
pH Solids:	1973	1	6.35	-	-
Suspended	1973	1	27.2	-	
Dissolved	1973	1	68	-	-
Total	1973	1	95	-	-

# GRAVEL WASHING PLANT

Constituent	Period of Record	No. of Values	Maximum	Values* Minimum	Mean
Coliform:fecal	1973	1	<2	-	-
Oxygen: BOD <sub>5</sub>	1973	1	<5	-	-
Flow	1973-1977	241	3222.5	196.4	414.1
Solids:	-			-	
Suspended	1973	1	15.2	-	
Dissolved	1973	1	399	-	_
Total	1973	1	414	4999	-

<sup>\*</sup>Values are as mg/L except:

- 1) Coliform-fecal as MPN/100 mL
- 2) Flow as  $m^3/d$
- 3) pH

TABLE 3

EFFLUENT DATA SUMMARY
HAZELMERE RIVERSIDE CAMPGROUND, SURREY (PE 1558)

Constituent	Period of Record	No. of Values	Maximum	Values* Minimum	Mean
Chloride	1975	1	166	-	- 166 000†
Coliform:fecal Oxygen:BOD <sub>5</sub>	1975-1977 1975-1978	4 10	>240 000 462	>24 000 32	166 000 <sup>+</sup> 225
pH Specific	1975-1978	4	7.6	7.1	7.4+
Conductivity	1975-1977	4	1 770	45	1 350
Solids: Total	1975-1978	11	446	5	106

<sup>\*</sup>All values are as mg/L except:

<sup>(1)</sup> Coliform as MPN/100 mL

<sup>(2)</sup> pH

<sup>(3)</sup> Specific Conductivity as μS/cm

<sup>&</sup>lt;sup>+</sup>Median Value

TABLE 4
WATER QUALITY DATA SUMMARY

# LITTLE CAMPBELL RIVER SITE 0300066

	Period of	No. of		Values*	<del></del>
Constituent	Record	Values	Maximum	Minimum	Mean
Alkalinity:Total	1979-1983	10	71.6	14.7	37.6
Arsenic:Total	1982	2	<0.25	<0.25	<0.25
Chloride	1979-1982	8	9.3	4.3	6.3
Coliform:			,,,	. • 3	0.5
fecal	1979-1983	9	5400	8	90+
Hardness:		-		Ŭ	, ,
total	1979-1981	5	43.3	18.3	29.4
calcium	1979-1981	5	10.9	4.7	7.3
magnesium	1979-1981	5	3.9	1.6	2.8
Metals:Total		,	J• J	1.0	2.0
Aluminum	1982	2	0.08	<0.02	_
Cadmium	1982	2	<0.01	<0.01	<0.01
Chromium	1982	2	<0.01	<0.01	<0.01
Copper	1982	2	<0.01	<0.01	<0.01
Iron	1979-1982	6	3.29	0.3	1.49
Lead	1981-1982	3	<0.1	<0.01	-
Manganese	1979-1982	2 6 3 7 2 2	1.29	<0.01	
Molybdenum	1982	2	<0.01		0.35
Nickel	1982	2	<0.05	<0.01	<0.01
Vanadium	1982	2	<0.01	<0.05	<0.05
Zinc	1979-1982	7	0.007	<0.01	<0.01
Nitrogen:	1919 1902	1	0.001	<0.005	0.007
Ammonia	1979-1983	9	0.389	0.009	0 111
Nitrate	1979-1983	10	2.14		0.111
Nitrite	1979-1983	10	0.024	<0.02	0.98
Organic	1979-1983	10	1.64	<0.005	0.01
Nitrate/Nitrite		10	2.16	0.39	0.84
Kjeldahl	1979-1983	10	1.91	<0.02	0.99
Oxygen:	1919 1905	10	1.71	0.44	0.91
dissolved	1979-1983	9	10.2	2.7	<i>C</i> =
%saturation	1979-1982	9	76.1	2.7	6.5
pH	1979-1983	10	7.0	30.1	56.7
Phosphorus	1212 1203	10	<i>i</i> • 0	6.5	6.8+
ortho diss.	1979-1982	8	0.101	0.010	0.007
total diss.	1979-1982	8	0.141	0.013	0.037
Potassium: Dissolve		8	3.8	0.026	0.062
Sodium: Dissolved	1979-1982	8	9.6	1.4	2.4
Solids:	1919 1902	J	9.0	4.0	5.6
Dissolved	1979-1983	10	108	66	92 -
Suspended	1979-1983	10	22	66	83.5
Total	1979-1983	10	130	2 70	7
	171,7 1703	, 0	130	70	90

TABLE 4 (Continued)

Constituent	Period of Record	No. of Values	Maximum	Values* Minimum	Mean
Specific Conductivity Sulphate Temperature	1979-1983	10	174	74	111
	1979-1982	8	12.8	2.7	5.7
	1979-1983	9	17.5	0.3	10.3

- \*All values are as mg/L except:
  - (1) Coliform as MPN/100 mL

  - (2) pH
    (3) Temperature as °C
  - (4) Specific conductivity as  $\mu$ S/cm
  - (5) % saturation as %

+Median value

TABLE 5 WATER QUALITY DATA SUMMARY LITTLE CAMPBELL RIVER SITE 0300065

Constituent	Period of Record	No. of Values	Maximum	Values* Minimum	Mean
Alkalinity:Total Arsenic:Total Chloride Coliform: fecal Hardness:	1979-1983 1982 1979-1982 1979-1983	10 2 8 9	66.1 <0.25 13.4 9200	18.2 <0.25 6.0 130	44.4 <0.25 9.3 625
total calcium magnesium Metals:Total	1979-1981 1979-1981 1979-1981	5 5 5	57.8 14.4 5.3	23.6 6.0 2.1	37.8 9.3 3.5
Aluminum Cadmium Chromium Copper Iron Lead Manganese Molybdenum Nickel Vanadium Zinc Nitrogen:	1982 1982 1982 1979-1982 1979-1982 1979-1982 1982 1982 1982 1982 1982	2 2 7 7 7 7 2 2 2	0.34 <0.01 <0.01 0.002 2.2 0.011 0.08 0.01 <0.05 <0.01 0.02	0.11 <0.01 <0.01 <0.001 0.20 <0.001 0.05 <0.01 <0.05 <0.01	0.22 <0.01 <0.01 0.001 1.05 0.007 0.06 -<0.05 <0.01
Ammonia Nitrate Nitrite Organic nitrate/nitrite Kjeldahl Oxygen:	1979-1983 1979-1983 1979-1983 1979-1983 1979-1983 1979-1982	9 10 10 10 10	0.320 2.39 0.024 2.10 2.40 2.34	0.016 0.93 0.007 0.27 0.75 0.29	0.107 1.43 0.014 0.66 1.45 0.73
dissolved % saturation pH Phosphorus:	1979-1983 1979-1983 1979-1983	9 9 10	14.2 139.2 8.6	8.4 79.5 6.9	8.8 100.0 7.6+
ortho - dissolved total - dissolved Potassium:Dissolved Sodium:Dissolved Solids:	1979-1983 1979-1983 1979-1982 1979-1982	10 10 8 8	0.090 0.122 3.3 13.3	0.015 0.021 1.4 5.9	0.048 0.064 2.2 9.7
Dissolved Suspended Total Specific	1979-1983 1979-1983 1979-1983	10 10 10	142 63 186	76 4 98	105.3 16.7 122.0
Conductivity Sulphate Temperature	1979-1983 1979-1982 1979-1983	10 8 9	218 14.6 16.8	92 5.1 0.5	149.2 8.0 9.1

<sup>\*</sup>All values are as mg/L except:
(1) pH
(2) Temperature as °C
(3) % saturation as %
(4) Specific conductivity as µS/cm
+Median value
Data Source: Ministry of Environment and Parks

TABLE 6

TOTAL AMMONIA VALUES: LITTLE CAMPBELL RIVER

	Temperature °C	Hd	Total Ammonia-N mg/L	Maximum Concentration of Total Ammonia-N for Protection of Aquatic Life(8)	Average 30-day Concentration of Total Ammonia-N for Prgtection of Aquatic Life mg/L
Site 0300065 February, 1979 August, 1980 February, 1981 May, 1981 May, 1982 September, 1982 November, 1982 July, 1983 August, 1983	3.7.6 16.8 3.2 11 13 0.5 14.8	6.7 7.7 8.3 7.7 7.7 7.7 7.7	0.320 0.045 0.241 0.065 0.097 0.023 0.016 0.118	23.5 14.7 9.2 19.0 3.0 1.5 16.1	1.99 1.89 1.56 1.98 0.57 0.30 2.06 1.62
Site 0300066 February, 1979 August, 1980 February, 1981 May, 1981 November, 1982 November, 1982 July, 1983 August, 1983	2 7.5 17.5 2.9 11 0.3 19.2	6.9 6.9 6.9 6.9 6.9 6.9	0.386 0.048 0.275 0.114 0.049 0.038 0.057 0.061	27.9 - 22.3 20.7 24.8 18.5 27.9 20.5	2.02 - 1.89 1.48 1.99 1.83 2.07 1.30 1.34

Data Source: Ministry of Environment and Parks

TABLE 7

BIOLOGICALLY AVAILABLE NITROGEN TOTAL DISSOLVED PHOSPHORUS RATIOS:
LITTLE CAMPBELL RIVER

1	51	te 0300066		Sit	e 0300065	
Date	Nitrate + Ammonia (mg/L)	Total Dissolved Phosphorus (mg/L)	N:P Ratio	Biologically Available Nitrogen* (mg/L)	Total Dissolved Phosphorus (mg/L)	N:P Ratio
Feb. 1979 Sept. 1980 Feb. 1981 May 1981 Nov. 1981 May 1982 Sept. 1982 Nov. 1982 July 1983	2.529 0.058 2.155 0.975 1.919 1.018 0.075 1.917	0.082 0.031 0.054 0.080 0.067 0.038 0.056 0.043 0.141	30.8 1.9 39.9 12.2 28.6 16.8 1.3 44.6	2.71 1.71 1.88 1.03 2.11 1.01 0.95 1.87	0.088 0.036 0.122 0.059 0.066 0.021 0.050 0.067	30.8 47.5 15.4 17.5 32.0 48.1 19.0 21.9

TABLE 8

SEDIMENT CHEMISTRY ANALYSES
LITTLE CAMPBELL RIVER AND OTHER B.C. SITES

					הזווה כאחר פניי	billus Camf Beel niven and Olnen B.C. Siles	0.00	**************************************			1
		anvo a tet	davie madoury area.			FRASER RIVER SEDIMENTS	SEDIMENTS				
THEFT	3	LILLE CAMI	ספרף עווענע		N WESTMINSTER!	OAK ST BRIDGE!	WESTMINSTER ONK OF RRIDGE! N WESTMINSTER ONK ST BRIDGE	OAK ST BRIDGE?	SENTMENTS	BESTDENTIAL SOMBATSH	SOHAMISH
	SITE 1 NEAR HWY. 99	0300065	SITE 5 NEAR 200 STREET	9900080	SOUTH OF MIDSTREAM	SOUTH OF MIDSTREAM	SOUTH OF MIDSTREAM	SOUTH OF MIDSTREAM			ESTUARY 5
METALS											
+	8.23	7.65	60.6	10.4					6	4.60-8.02	
		14.7	53	51					15 -1860		
Boron*		41.0	رب م	9	-				144 - 1		
Cadm1 um*		41.0	<1.0	m	0.1-0.2	0.1	0.1-0.2	0.1-0.3	= -	<1.0-19	
		2.27	2.96	2.87					0.3 - 361		
*_		28	37	34					1 - 120		
Copper*		6	6	250	9.6-14	6.0-0.9	19.0-23.0		ŧ	22-29	80
		11.7	19.8	17	3.1-3.3	4.0-4.3	5.3-6.1	5.0-7.8	0.3 - 128	11-13.4	
		100	24	30	2.1-2.4	3.5-3.7	2.5-3.0	5.3-7.6	10 - 560	160-209	
sium+		4.2	5.16	5.06					0.57- 273		
*		297	843	366	150	120-130	170-190	150-190	11 -9660		
		7	=	13					1 - 183		
		19	23	27	12-13	13.0-14.0	17.0-20.0	17.0-26.0	ı		
Selenium*		<10.0	13	<10.0					10 - 85		
*		13	<del></del>	15					2 -1790		
		38	26	20					70 - 604		
		<20.0	<u></u>	30					20 - 122		
Tin*		<5.0	<5.0	<5.0					5 - 42		
Titanium*		387	619	469					1 -3610		
Vanadium*		22	<u>Q</u>	₩.					1 - 203	į	
		270	6†	h6	9-10	17.0-19.0	13.0-15.0	31.0-42.0	8 - 402	101-183	
			-		THE RESERVE THE PROPERTY OF THE PERSON NAMED IN THE PERSON NAMED I						-

+Unit - mg/g dry weight
\*Unit - ug/g dry weight
\*Unit - ug/g dry weight
'Analyses on sediments with particle size >180 µm, <850 µm; by Stancil; Ministry of Environment (1980)

Analyses on sediments with particle size <180 µm; by Stancil; Ministry of Environment (1980)

Squamish Estuary, Air and Water Quality Work Group. 1981.

\*Squamish Estuary, Air and Water Quality Work Group. 1981.

\*\*L. McKean; Ministry of Environment and Parks (unpublished)

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TABLE 8 (Continued)

	LI	TTLE CAM	PBELL RIVER				
Constituent <sup>+</sup>	SITE 1 NEAR HWY. 99	0300065	SITE 5 NEAR 200 STREET	I .	ADJACENT TO IONA STP <sup>21</sup>	FROM B.C. LAKES <sup>22</sup>	SQUAMISH ESTUARY <sup>5</sup>
Carbon Organic Nitrogen Kjeldahl Residue Total Vol. Oxygen B.O.D. Phosphorus Total	2.0 0.2 1.5 0.36 0.468	5.8 0.6 2.4 0.76 0.424	<0.3 0.4 1.9 0.67 0.515	23.0 2.8 11.1 1.97 0.699	0.221-0.728 0.588-0.910	1.0-382.0 0.30-41.50 2.70-84.8	0.5-67.0

<sup>+</sup> Unit: mg/g dry weight

<sup>&</sup>lt;sup>5</sup> Squamish Estuary, Air and Water Quality Work Group. 1981.

<sup>&</sup>lt;sup>22</sup> C. McKean; Ministry of Environment (Unpublished)

D. Stancil; Fraser River Estuary Study. Aquatic Biota and Sediments 1980.

TABLE 9
ESCAPEMENT VALUES FOR THE SERPENTINE RIVER

<u>,</u>			
YEAR	СОНО	STEELHEAD	
1947	750	0	
1948	1500		
1949	400		
1950	1500		
1951	1500	<b>7</b> 5	
1952	1500		
1953	1500	•	
1954	750		
1955	750		
1956	1500		
1957	400		
1958	750		
1959	1500		
1960	400		
1961	400		
1962	400		
1963	400		
1964	750		
1965	400		
1966	400		
1967	200		
1968	200		
1969	75		
1970	750		
1971	3500		
1972	1500		
1973	1500		
1974	2000		
1975	2800		
1976	3500		
1977	2400		
1978	1800	•	
199	1500		
1980	500		
1981*	1100 600		
1982** 1983	350		
1303	390		

\* 2 chinook

\*\* 4 chinook

Timing of Spawning:

Arrive

October

Start Peak Late October Middle November

End

Late December

Data From: M.J. Hancock, Marshall, D.E. 1985. Catalogue of Salmon Streams and Spawning Escapements of Statistical Area 29. Canadian Data Report of Fisheries and Aquatic Science No. 495. Department of Fisheries and Oceans.

TABLE 10
EFFLUENT DATA SUMMARY
ASSOCIATED FOUNDRY LTD. SURREY (PE 1529)

CONSTITUENT	PERIOD OF RECORD	NO. OF VALUES	MAXIMUM	VALUES* MINIMUM	MEAN	LEVEL "C" OBJECTIVES
pН	1978-1982	36	11.2	5.4	7.03+	6.5-8.5
Oil/grease	1978-1982	37	60	2	19.9	15
Solids				:		-
Suspended	1978-1982	36	246	4	77.2	100
Flow	1978-1982	12	63	46	53.4	
Arsenic:Total	1980	1	<0.05	-	_	1
Hardness:calcium	1980	1	2.61	-	-	-
magnesium	1980	1	0.34		_	-
METALS:Total				:		
Cadmium	1980	1	<0.01	_	-	0.10
Chromium	1980	1	<0.01	-	-	0.60
Copper	1980	1	<0.01	-	_	1
Iron - dissolved	1980	1	1.9	_	-	1
- total	1980	1	2.43	-	-	-
Manganese	1980	1	0.05	<del>-</del>	-	1
Molybdenum	1980	1	<0.01	-	_	10
Nickel	1980	1	<0.05	_	-	2
Zinc	1980	1	0.06		-	1
Oxygen:BOD <sub>5</sub>	1980	1	<10	-	-	- [
Oil & Grease	1980	18	60	3 6	19.7	15
рH	1979-1980	17	7	6	6.69+	6.5-8.5
Phosphorus: total	1980	1	<0.5	-	-	5
Solids: suspended	1979-1980	17	246	4	90	100
Specific Conductivity		5 3	27	21	23	
Temperature	1979-1980	3	34	28	31.3	-

<sup>\*</sup> All values are mg/L except:

- 1) pH
- 2) Specific Conductivity as uS/cm
- 3) Temperature as °C
- 4) Flow as  $m^3/d$

### + Median Value

Data Source: Ministry of Environment and Parks

NOTE: Level "C" objectives are for industries which discharge metals.

# TABLE 11(a) WATER QUALITY DATA SUMMARY SERPENTINE RIVER

	Site 1100032 Serpentine River at Freeway					
Constituent	Period of Record	No. of Values		Values*		
	лесога	varues	Maximum	Minimum	Mean	
Colour: True Carbon: Organic	1981 1981 1981 1981 1981 1981 1981 1981	22252 222222222222222222222222222222222	15 3 13 2200 6.4 0.007 0.44 0.006 0.45 < 10 13 2 7.8 0.020 0.028 0.028 0.03 1.6 6.9 100 2 99 1	15 2 13 230 5.8 0.005 0.44 0.006 0.45 < 10 < 10 1.2 7.8 0.016 0.021 0.03 1.4 6.4 98 1	15 2.5 13 790+ 6.1 0.006 0.44 0.006 0.45 <10 <10 1.6 7.8 0.018 0.024 0.03 1.5 6.65 99 1.5 98	
Specific Conductivity Silica: Reactive Sulphate Tannin & Lignin	1981 1981 1981 1981	2 2 2	139 18.6 6.3 0.4	127 17 5.4 0.4	133 17.8 5.85 0.4	

<sup>\*</sup> All values are as mg/L except:

<sup>(1)</sup> pH
(2) Specific conductivity as μS/cm
(3) Colour as true colour units
+ Median value

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TABLE 11(a) Continued

	Site 0300057 Serpentine R. King George Hwy				
Constituent	Period of Record	No. of Values	Values*		
	necor d	varues	Maximum	Minimum	Mean
Alkalinity: total	1972,1983	43	102.	8.5	46.7
Boron: dissolved	1972	3	0.7	< 0.1	0.2+
Carbon: Organic	1981-1983	27	18	6	11.7
Inorganic	1981-1983	20	21	9	15.1
Chloride Coliforms: fecal	1972-1983 1973-1983	44 51	13 800 >24 000	10.8	1229 110 +
1	1973-1963	19	>24 000	140	>1609+
total Colour:True	1972-1977	20	160	1	
Hardness:total	1972-1982	30	4580	30 40.5	79.5 588.
calcium	1972-1982	30	292	9.8	47.4
magnesium	1972-1982	30	935 -	3.9	115
Metals: (Total)	1912 1902	٥٥	300 -	3.9	117
Aluminum	1982-1983	6	1.32	0.51	0.83
Cadmium	1974-1982	8	< 0.01	< 0.0005	-
Chromium	1972-1983 1972-1983	23 36	0.016	< 0.005	0.006+
Copper Iron	1973-1983	36	0.03 3.6	0.001 0.7	0.006
Lead	1972-1982	39	< 0.1	< 0.001	0.005+
Manganese	1972-1983	36	0.46	0.06	0.19
Molybdenum	1975-1983	7	0.05	0.002	< 0.01+
Zinc	1972-1983	39	0.08	< 0.005	0.019
Nitrogen: Ammonia	1974-1983	52	0.532	< 0.005	0.142
Nitrate	1972-1983	46	2.83	< 0.02	0.63
Nitrite	1972-1983	61	0.246	< 0.005	0.022
nitrate/nitrite	1975-1983	47	2.86	< 0.02	0.66
Kjeldahl	1972-1983	51	3.62	0.43	1.17
Organic Oxygen: BOD <sub>5</sub>	1973-1983 1981-1982	46 8	3.16 < 10.	0.39 < 10	1.05
dissolved	1972-1983	44	15.6	2.7	9.4
COD	1981-1982	31	65.	13	40.9
% Saturation	1972-1983	44	176.5	24.7	90.5
Oil and Grease	1981	5	4.	1.2	2.2
pH Phenol	1972-1983 1982	61	8.9	6.2	]
Phosphorus:ortho-diss.	1973-1983	57	0.006 0.122	< 0.002 < 0.003	0.003 0.032
total	1972-1983	37	0.262	0.003	0.139
total dissolved	1976-1983	43	0.153	0.016	0.139
Potassium: diss.	1972-1983	41	28.6	2.2	24.9
Silica	1972-1982	11	19.1	11.7	15.0
Sodium:dissolved	1972-1982	40	7680	9.5	610
Solids:Suspended	1974-1983	49	57	9	24
Dissolved	1972-1983	22	26 500	118	4297
total	1972-1983	40	9790	142	1388
Specific Conductivity	1972-1982	62	36 000	156	3168
Sulphate	1972-1981	42	1804	19.6	182
Tannin/Lignin	1981	14	2.8	1.4	2.1
Temperature	1972-1983	44	27.5	1.8	12.3
Turbidity	1972-1983	8	36	7.2	15.9

<sup>\*</sup> All values are as mg/L except

<sup>(1)</sup> Coliforms as MPN/100 mL (2) Colour as colour units (3) pH

<sup>(4)</sup> Temperature as °C (5) Turbidity as N.T.U. (6) Specific Conductivity as  $\mu S/cm$ 

<sup>(7) %</sup> Saturation or %

<sup>+</sup> Median Value

TABLE 11(a) Continued

	Site 1100034 Serpentine River at Hwy 99					
Constituent	Period of No. of		Values*			
	Record	Values	Maximum	Minimum	Mean	
Alkalinity: total Arsenic:Total Boron dissolved Chloride Coliforms: fecal total Colour:True Hardness:total Calcium Magnesium Metals:(total) Cadmium Copper Iron Lead	1979-1982 1981 1979 1979-1982 1979-1981 1979-1982 1979-1982 1979-1982 1979-1982 1979-1982 1979-1982 1979-1982	6114217666 1556	98.6 < 0.05 0.1 13900 1100 >2400 150 4380 350 880 < 0.01 0.05 3.4 0.009	27.7 - 123 43 - 15 78.9 15.6 9.7 - 0.002 0.4 0.002	65.8 - 6432 - 60 1423 104 283 0.015 1.66 0.005	
Zinc Nitrogen:  Ammonia Nitrate Nitrite Nitrate/Nitrite Kjeldahl Organic Total Oxygen: COD Dissolved pH Phenol Phosphorus:ortho-diss. total Potassium dissolved Sodium dissolved Solids:Suspended dissolved Solids:Suspended Total Specific Conductivity Sulphate Dissolved Tannin/Lignin Temperature	1979-1982 1981 1981 1981 1979-1982 1979-1982 1979-1982 1979-1981 1979-1982 1979-1982 1979-1982 1979-1982 1979-1982 1979-1981 1979-1981 1979-1982 1979-1982 1979-1982 1979-1982 1979-1982 1979-1982 1979-1982 1979-1982	6 31176461381574477774156	0.11  0.25 0.40 0.01 1.82 2 2 3 16.5 8.5 0.006 0.101 0.202 280 7750 44 27399 27700 38700 1811 1.4 21 25	0.005  < 0.01 - < 0.02 0.44 0.49 - 2.5 6.7 - 0.011 0.127 5 83.5 5 121 126 163 27 - 7 11	0.032 0.097 - 0.544 1.12 1.17 1.65 - 5.1 7.8+ - 0.074 0.173 130 3603 30 7420 7488 106 839 - 16.7 16.5	

<sup>\*</sup> All values are as mg/L except:

<sup>(1)</sup> Coliform as MPN/100 mL (2) Colour as colour units (3) pH (4) Temperature as °C (5) Turbidity as N.T.U., (6) Specific Conductivity  $\mu$ S/cm

<sup>+</sup> Median value Data Source: Ministry of Environment and Parks

124 TABLE 11(a)-Continued

<sup>\*</sup> All values are as mg/L except:

<sup>(1)</sup> pH (2) Temperature as °C (3) Specific conductivity as  $\mu$ S/cm (4) Colour as colour units (5) Turbidity as NTU (6) % Saturation as %

<sup>+</sup> Median value

TABLE 11(a) (Continued)

	Site 1100035 Serpentine River at Hwy 10				Hwy 10	
Constituent	Period of Record	No. of Values	la contraction of the contractio			
	Record	varues	Maximum	Minimum	Mean	
Alkalinity: total	1974-1979	18	69	14	44° A	
Arsenic: diss.	1974	1	< 0.005		0.24	
Boron:dissolved	1974-1979	22	1.8	< 0.1	13.8	
Carbon: organic	1974-1981	25	42 26	5 19	21.7	
inorganic	1981	3	9200	6.1	580	
Chloride	1974-1981	31 8	>24000	330	490 +	
Coliforms: fecal	1981	17	150	30	65.3	
Colour:True	1976-1981	14	111	26.8	62	
Hardness:total	1976-1979		200	6.6	23.3	
calcium	1974-1979	35	608	2.1	38.6	
magnesium	1974-1979	35	000	۲۰۱	50.0	
Metals:(total)	4.07 k. 4.004	1=	< 0.01	< 0.0005	_	
Cadmium	1974-1981 1974	15 2	< 0.005	< 0.005	< 0.005	
Chromium	1974-1981	8	0.014	0.001	0.006	
Copper	1974-1901	25	6.5	0.8	1.96	
Iron		32	< 0.1	< 0.001	0.009	
Lead	1974-1981	16	0.00006	< 0.00005	<0.00005	
Mercury	1974-1975	21	0.13	< 0.005	0.02	
Zinc	1974-1981	23	0.13	0.005	0.02	
Nitrogen:	1976-1981	16	0.735	0.014	0.245	
Ammonia	1981	10	0.79	0.02	0.53	
Nitrate	1981	4	0.044	< 0.005	0.021	
Nitrite Nitrate/Nitrite	1974-1981	40	3.3	< 0.02	0.68	
Kjeldahl	1974-1981	38	3	0.59	1.40	
Organie	1978-1981	8	3	0.43	1.33	
Total	1974-1979	12	14	1.02	2.37	
Oil & Grease	1,981	1	1.5	<b>-</b>	-	
Oxygen: BOD <sub>5</sub>	1981	3	55	< 10	25.	
Dissolved	1974-1979	27	15.8	4.1	9.9	
COD	1981	5	133	20.2	61.1	
% saturation	1974-1979	27	180.5	43	102.2	
Н	1974-1981	41	8.3	6	7.0 +	
Phenol	1981	1	0.007			
Phosphorus:ortho diss.	1974-1981	33	1.01	0.01	0.099	
total diss.	1974-1981	36	0.68	0.026	0.095	
total	1974-1981	40	1.56	0.075	0.221	
Potassium: dissolved	1974-1981	33	191	1.4	14	
Silica	1981	3	19.2	14.5	17.2	
Sodium:dissolved	1974-1981	30	5100	6	336 43	
Solids:Suspended	1974-1981	8	82	17	231	
dissolved	1979-1982	5	282	123 160	295	
total	1974-1981	6	396	99	1648	
Specific Conductivity	1974-1981	43 28	36300 1180	9.7	110	
Sulphate	1974-1981	5	4.4	1.9	2.8	
Tannin/Lignin	1981 1974-1979	31	22	2.7	14.6	
Temperature Turbidity	1974-1979	36	76	4.7	17.4	
14 544103			<u> </u>	ļ		

<sup>\*</sup> All values are as mg/L except:

<sup>(1)</sup> Coliform as MPN/100 mL (2) Colour as true colour units (3) pH (4) Temperature as °C (5) Turbidity as NTU, (6) Specific Conductivity as  $\mu$ S/cm

<sup>(7) %</sup> saturation as %

<sup>+</sup> Median value

TABLE 11(a) Continued

		Site 1	100049 Serpen	tine River at	64 Ave
Constituent	Period of Record	No. of Values		Values*	
	ne cor q	Varues	Maximum	Minimum	Mean
Alkalinity: total Boron:dissolved Chloride Colour: true Hardness:total calcium magnesium Metals: (total) Cadmium	1974-1979 1974-1979 1974-1979 1976-1979 1976-1979 1974-1979	22 18 24 14 18 32 32	113 0.3 556 400 568 85.5 86.5	<ul> <li>0.5</li> <li>0.1</li> <li>9.5</li> <li>5</li> <li>60.4</li> <li>7.5</li> <li>3.8</li> </ul>	25.7 0.17 98.9 86 268 40.5 38.6
Chromium Copper Iron Lead Mercury Zinc	1974-1975 1974 1974 1974-1977 1974-1975 1974-1975	13 2 3 21 13 13	0.007 < 0.005 0.03 33.4 0.006 < 0.00005 0.24	< 0.0005 < 0.005 0.003 0.6 < 0.001 < 0.00005 < 0.005	0.0008 <sup>+</sup> < 0.005 0.004 <sup>+</sup> 5.3 0.002 < 0.00005 0.09 <sup>+</sup>
Nitrogen:  Ammonia Nitrate/nitrite Kjeldahl Organic Total Oxygen: dissolved % saturation pH Phosphorus:ortho diss. total diss. total Potassium:Dissolved Sodium:Dissolved Sodium:Dissolved Solids:Suspended Dissolved total Specific Conductivity Sulphate Temperature	1976-1979 1974-1979 1974-1979 1978-1979 1976-1979 1974-1979 1974-1979 1974-1979 1974-1979 1974-1979 1974-1979 1976-1979 1979 1979 1979 1979 1974-1979	14 32 32 6 11 28 27 32 28 27 32 23 23 1 1 34 22 31	0.572 10.1 3 2 11 17.6 180.4 7.7 0.123 0.153 0.22 14.8 314 63 210 214 2060 664 26	0.023 < 0.02 0.37 0.49 0.84 2.4 26.6 3.8 < 0.003 0.003 0.0011 2.3 7.4 4 127 13.7 2.6	0.219 1.29 1.33 1.17 3.22 8.6 84.8 5.1 0.019 0.033 0.073 5.8 83 29 - 910 280 15.1

<sup>\*</sup> All values are as mg/L except:

<sup>(1)</sup> pH (2) Temperature as °C (3) Turbidity as NTU (4) Specific Conductivity as  $\mu$ S/cm (5) % saturation as %

<sup>+</sup> Median value

TABLE 11(a) Continued

Site 1100036 Serpentine River at 80 Ave					80 Ave	
Constituent	Period of	No. of	Values*			
	Record	Values	Maximum	Minimum	Mean	
Alkalinity: total Boron:Dissolved Carbon: organic inorganic Chloride	1974-1979 1974-1979 1974-1981 1981 1974-1981	22 21 23 2 30	81.8 0.1 20 15 196	12.5 < 0.1 3 15 4.2	46.6 < 0.1+ 10.7 15 29	
Coliforms: fecal Colour:True Hardness:total calcium magnesium	1981 1976-1982 1976-1979 1974-1979	1 15 17 38 38	790 150 62.1 15.8 15.9	- 10 15.1 8.9 1.3	- 57.7 45.4 10.2 4.6	
Metals: (total) Iron Nitrogen:	1976-1977	9	4.6	1.0	2.0	
Ammonia Nitrate Nitrite Nitrate/nitrite Kjeldahl Organic Total Oxygen: BOD <sub>5</sub> Dissolved COD % saturation Oil & Grease pH Phosphorus:ortho-diss. total Potassium:dissolved Silica Sodium:dissolved	1976-1981 1981 1981 1974-1981 1974-1979 1978-1979 1981 1974-1979 1981 1974-1981 1974-1981 1974-1981 1974-1981 1974-1981 1974-1981 1974-1981 1981 1981 1981	14 2 40 39 41 11 29 29 40 32 35 41 31 2	0.277 1.01 0.019 3.32 3 1 3 < 10 14.2 17 164.9 1.5 8.4 0.132 0.159 0.432 6.1 16 122	0.017 1.01 0.019 0.11 0.13 0.72 0.31 < 10 5.0 17 55.5 - 6.0 0.022 0.041 0.033 0.8 16 4.3	0.168 1.01 0.019 0.744 0.93 0.89 1.60 <10 8.9 17 87.9 - 7.2+ 0.074 0.091 0.168 2.4 16 22.4	
Solids:Suspended Dissolved total Specific Conductivity Sulphate Tannin/Lignin Temperature Turbidity	1974-1981 1979-1982 1974-1981 1974-1981 1974-1981 1981 1974-1979 1974-1979	8 2 5 42 26 2 34 39	38 145 194 860 34.2 0.4 21.5	100 104 58 <5 0.4 2.9 1.7	21 122 140 205 11.9 0.4 13.3 15.5	

<sup>\*</sup> All values are as mg/L except:

<sup>(1)</sup> Coliform as MPN/100 mL (2) Colour as colour units (3) pH

<sup>(4)</sup> Temperature as °C (5) Turbidity as N.T.U. (6) Specific Conductivity µS/cm

<sup>(7) %</sup> saturation as %

<sup>+</sup> Median value

Data Source: Ministry of Environment and Parks

	1		<del></del>		
		Site 0	300059 Serpen	tine River 80	Ave
Constituent	Period of No. of Record Values				
		laraco	Maximum	Minimum	Mean
Alkalinity: total Boron: diss. Carbon: Organic Inorganic	1972-1983 1972 1981-1983 1981-1983	48 3 21 14	76.6 < 0.2 23 22	13 < 0.1 4 17	44 < 0.1 <sup>+</sup> 10.1 19.5
Chloride Coliforms: fecal Colour:True Hardness:total calcium magnesium Metals: (total)	1972-1983 1973-1983 1972-1982 1972-1982 1972-1982 1972-1982	42 55 16 33 33 33	235 9200 150 138 21.5 20.6	5.2 40 20 17.5 4.5 1.53	34.2 1300 + 67.8 51.8 11.1 5.97
Aluminum Cadmium Chromium Copper Iron Lead Manganese Molybdenum Zinc	1982-1983 1982-1983 1972-1983 1972-1983 1973-1982 1972-1983 1972-1983 1972-1983	6 21 37 36 39 37 7	1.39 < 0.01 < 0.01 0.03 6.5 < 0.1 0.6 < 0.02 0.05	0.27 < 0.01 < 0.005 < 0.001 0.6 < 0.001 0.04 0.0027 < 0.005	0.76 0.01 0.005+ 0.004+ 1.7 0.004+ 0.14 0.01+ 0.015
Nitrogen:  Ammonia Nitrate Nitrite Nitrite nitrate/nitrite Kjeldahl Organic Oil and Grease Oxygen: BODs dissolved COD Saturation Phenol pH Phosphorus:ortho-diss. total diss. total Potassium:Dissolved Silica Sodium:Dissolved Solids:Suspended Dissolved total Specific Conductivity Sulphate Temperature Tannin/Lignin	1974-1983 1976-1983 1972-1983 1972-1983 1972-1983 1981 1981 1981 1981 1981 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983	23130933888385067774498888 54645433774498888	0.582 2.3 0.048 2.32 4 3.95 4.6 < 10. 16.7 71 167.8 0.008 7.8 0.261 0.291 0.552 5.9 20 146 84 490 730 1170 110 21 3.2	0.006 0.14 0.005 0.17 0.38 0.45 1.1 < 10 1.1 13 9.7 0.002 6.3 0.02 0.036 0.074 1.6 8.4 6.8 7 98 68 63 7.2 1.4	0.215 0.82 0.021 0.84 1.22 1.01 2.6 < 10 8.2 33.1 74.1 0.005 7.0+ 0.099 0.127 0.196 3.0 15.6 26 26 173 218 261 22.5 10.2 2.2

<sup>\*</sup> All values are as mg/L except

<sup>(1)</sup> Coliforms as MPN/100 mL (2) Colour as colour units (3) pH  $\,$ 

<sup>(4)</sup> Temperature as °C (5) Specific Conductivity as  $\mu S/cm$ 

<sup>+</sup> Median Value

TABLE 11(a) (Continued)

	Site 1100041 Serpentine River at 88 Ave						
Constituent	Period of Record	No. of Values		Values*			
	Record	varues	Maximum	Minimum	Mean		
Carbon: - Organic	1981 1981 1981 1981 1981 1981 1981 1981	22242 222222222222222222222222222222222	17 22 47 490 80  0.288 0.76 0.025 2. 0.79 2 < 10 62 6.9 0.097 0.376 4 18.1 39 251 85 336 399 49.2 2.5	17 22 47 230 80 0.288 0.76 0.025 2 0.79 2 < 10 62 6.9 0.097 0.376 4 18.1 39 251 85 336 399 49.2 2.5	17 22 47 460 + 80 0.288 0.76 0.025 2 0.79 2 <10 62 6.9+ 0.097 0.376 4 18.1 39 251 85 336 399 49.2 2.5		

<sup>\*</sup> All values are as mg/L except:

<sup>(1)</sup> Coliform as MPN/100 mL (2) pH (3) Temperature as °C

<sup>(5)</sup> Turbidity as N.T.U. (6) Specifc Conductivity as  $\mu$ S/cm

<sup>+</sup> Median value

TABLE 11(a) Continued

	Site 1100037 Serpentine River at 168 St.				
Constituent	Period of Record	No. of Values		Values*	
		Varues	Maximum	Minimum	Mean
Alkalinity: total Arsenic:Total Boron Carbon: Organic Inorganic Chloride: dissolved Coliforms: fecal Colour True Fluoride Hardness:total calcium magnesium Metals: (Total) Cadmium Chromium	1975-1978 1974 1974-1979 1974-1981 1981 1974-1981 1979-1981 1976-1981 1976-1979 1974-1979 1974-1979 1974-1975 1974-1975	7 2 21 25 3 4 11 15 1 18 38 38	9.7 < 0.005 < 0.1 21 18 30.4 3500 120 < 0.1 51.4 14 4.2 0.009 < 0.005	0.8 < 0.005 < 0.1 < 1 11 3.3 130 5 - 9.17 2.9 0.6 < 0.0005 < 0.005	4.8 < 0.005 < 0.1 5.4 15 7.7 872 + 29 - 36.9 9.6 2.8 0.0005 < 0.005
Copper Iron Lead Manganese Mercury Nickel Zinc Nitrogen:Ammonia Nitrate/Nitrite Nitrate Nitrite Organic Kjeldahl Oxygen: BOD <sub>5</sub> COD dissolved % saturation	1974-1979 1979 1974-1979 1974-1975 1974-1975 1974 1974-1979	3126612662346044999	0.007 4.2 0.044 0.07 < 0.0005 < 0.01 0.11 0.234 3.25 0.71 0.022 1.03 1.26 < 10 53 14.7	<pre>&lt; 0.001 0.1 &lt; 0.001 &lt; 0.02 &lt; 0.0005 - &lt; 0.005 0.01 0.03 0.65 0.016 0.14 0.11 &lt; 10 37 8.1</pre>	0.003 0.65 0.007 0.031 < 0.0005 - 0.015 0.0575 0.527 0.68 0.021 0.501 0.323 < 10 43.8 11.6
Oil and Grease pH Phosphorus:ortho diss.	1974-1979 1981 1974-1981 1974-1981 1974-1981 1974-1981 1974-1981 1979-1981 1979-1981 1979-1981 1974-1977 1981 1974-1981 1974-1981 1974-1981 1974-1981	1 42 35 37 42 34 7 7 4 3 29 44 33 38	149.1 1.2 9.3 0.108 0.12 0.269 3.7 27.8 298 63 80 100 19.5 70 365 2.3 18.9 36	84.7 - 6.7 0.008 0.012 0.022 0.7 3.3 84 2 258 0 16.3 < 5 41 1.9 3.5 1	111.3 -7.5 + 0.026 0.035 0.064 1.58 7.71 169 20 150 25 17.6 14.6 120 2.1 11.7 5.5

<sup>\*</sup> All values are as mg/L except

<sup>(1)</sup> Coliforms as MPN/100 mL (2) Colour as colour units (3) pH

<sup>(4)</sup> Temperature as °C (5) Turbidity as N.T.U. (6) Specific Conductivity as  $\mu$ S/cm (7) % saturation as %

<sup>+</sup> Median value

TABLE 11 (b)
SEDIMENT CHEMISTRY DATA SUMMARY, SERPENTINE RIVER

Constituent		Site 0300057				
		Period of Record	No. of Values	Values		
				Maximum	Minimum	
	Arsenic*	1982	1	< 25	-	
Carbon:	Organic+	1982	2	8.1	6.0	
	Inorganic+	1982	1	17	-	
Metals:					{	
	Aluminum+	1982	1	6.89	-	
	Boron*	1982	1	< 1	-	
	Barium+	1982	1	0.029	-	
1	Beryllium*	1982	1	< 1	-	
	Cadmium+	1982	1	< 0.001	-	
	Calcium+	1982	1	3.25	-	
ĺ	Chromium+	1982	1	0.031	-	
	Cobalt+	1982	1	0.012	-	
	Copper+	1982	1	0.021	-	
1	Iron+	1982	1	18.1	-	
	Lead+	1982	1	0.058	-	
	Magnesium+	1982	1	5.61	-	
	Manganese+	1982	1	0.182	-	
	Molybdenum*	1982	1	7	-	
	Nickel+	1982	1	0.038	-	
-	Selenium*	1982	1	< 10		
	Strontium*	1982	1	29	-	
1	Tellurium*	1982	1	< 20	-	
	Thallium*	1982	1	< 20	-	
	Tin*	1982	1	< 5	-	
	Titanium+	1982	1	0.19	-	
	Vanadium*	1982	1	26	-	
	Zinc+	1982	1	0.089	-	
Nitroge	n: Kjeldahl+	1982	2	2	1.463	
Phospho	rus:total+	1982	2	0.801	0.463	
	anic Component+	1982	2	5	1.92	

<sup>+</sup> Sample Unit = mg/g dry weight

<sup>\*</sup> Sample Unit =  $\mu g/g$  dry weight

TABLE 11(b) (Continued)

	Site 1100034 Serpentine River at Hwy 99				
Constituent	Period of Record	No. of Values	Values		
			Maximum	Minimum	
Arsenic*	1979	2	10	6	
Metals:					
Aluminum+	1979	1	8.69	-	
Barium+	1979	1	0.023	-	
Cobalt+	1979	1	0.015	-	
Calcium+	1979	2	3.15	2.7	
Cadmium+	1979	2	< 0.001	< 0.001	
Chromium+	1979	2	0.032	0.024	
Copper+	1979	2	0.021	0.015	
Iron+	1979	2	19.4	19.3	
Lead+	1979	2	0.023	0.019	
Magnesium+	1979	2	6.31	6.27	
Manganese+	1979	2 2	0.201	0.161	
Molybdenum*	1979	2	2	< 1	
Nickel+	1979	2	0.025	0.025	
Strontium*	1979	[ 1	27	-	
Titanium+	1979	1	0.276	-	
Vanadium*	1979	1 1	24	_	
Zinc+	1979	2	0.048	0.042	
Nitrogen:	1				
Kjeldahl+	1979	2	1.09	1.09	
Phosphorus:total+					

Data Source: Ministry of Environment and Parks + Unit = mg/g dry weight \* Unit = µg/g dry weight

TABLE 11(b) Continued

		Site 1100	035 Seri	oentine Ri <b>v</b> er	at Hwy 10	
Constituent		Period of	No. of Values	Values		
a que constante de la constant		Record	values	Maximum	Minimum	
Metals:	Arsenic* dissolved	1979	2	14	< 5	
	Aluminum+	1979	1	9.11	-	
	Antimony+	1979	1	0.015	-	
	Barium+	1979	1	0.036	<b>-</b>	
Ì	Cadmium+	1979	2	< 0.001	< 0.001	
	Calcium+	1979	2 2	4.47	2.75	
1	Chromium+	1979	2	0.036	0.014	
	Cobalt+	1979	1	0.016	-	
	Copper+	1979	2	0.026	0.013	
	Iron+	1979	2	26.7	15.2	
(	Lead+	1979	2	0.053	0.04	
	Magnesium+	1979	2	7.15	4.52	
	Manganese+	1979	2	0.52	0.256	
	Molybdenum+	1979	2	2	< 1	
	Nickel+	1979	2	0.045	0.038	
	Strontium*	1979	1	13	_	
[	Titanium+	1979	1	0.482	-	
	Vanadium*	1979	1	25	_	
I	Zinc+	1979	2	0.095	0.095	
Nitroge	n: Kjeldahl+	1979	2 2	2.79	2.52	
Phospho	rus:total+	1979	2	1.04	0.303	

<sup>+</sup> Unit = mg/g dry weight \* Unit =  $\mu$ g/g dry weight

TABLE 11(b) Continued

	Site 1100	036 Ser	36 Serpentine River at 80 Ave		
Constituent	Period of Record	No. of Values	Values		
			Maximum	Minimum	
Arsenic*	1979	2	15	< 5	
Metals:			<u>-</u>		
Aluminum+	1979	1	10.9	_	
Antimony+	1979	1	0.013	_	
Barium+	1979	1	0.061	-	
Cadmium+	1979	2	< 0.001	< 0.001	
Calcium+	1979	2	4.26	2.78	
Chromium+	1979	2	0.043	0.043	
Cobalt+	1979	1 1	0.017	-	
Copper+	1979	2	0.03	0.015	
Iron+	1979	2	29.0	19.1	
Lead+	1979	2	0.041	0.037	
Magnesium+	1979	2	8.49	5.34	
Manganese+	1979	2 2 2 2	0.361	0.274	
Molybdenum*	1979		2	< 1	
Nickel+	1979	2 2	0.037	0.017	
Strontium*	1979	1	16	_	
Titanium+	1979	1	0.438	_	
Vanadium*	1979	2	31	31	
Zinc+	1979	2 2 2	0.08	0.049	
Nitrogen: Kjeldahl+	1979	2	1.79	0.82	
Phosphorus:total+	1979	2	0.774	0.467	

<sup>+</sup> Unit = mg/g dry weight
\* Unit = μg/g dry weight

TABLE 11(b) Continued

	and the second	Site 0300059 River at 80	Serpentine Ave	Site 1100044 River at 10	
Cons	stituent	Period of Record	Values**	Period of Record	Values**
Carbon:	Arsenic* Boron+ Organic+ Inorganic+	1982 1982 1982 1982	29 < 1 63 70	1979 - - -	12 - - -
	Aluminum+ Barium+ Bryllium* Cadmium+ Calcium+ Chromium+ Cobalt+ Copper+ Iron+ Lead+ Magnesium+ Manganese+ Molybdenum* Nickel+ Selenium* Strontium* Telurium* Tinallium* Tinallium Vanadium* Zinc+ en: Kjeldahl+	1982 1982 1982 1982 1982 1982 1982 1982	11.8 0.081 < 0.001 6.9 0.031 0.014 0.041 19.5 0.098 4.23 0.379 10. 0.034 12 45 < 20 < 20 < 5 0.229 33 0.139 5.6	- 1979 1979 1979 1979 1979 1979 1979 197	- <0.001 2.58 0.019 - 0.012 16.4 0.071 4.61 0.478 1 0.015 
	Zinc+ en: Kjeldahl+ orus: total+				

Data Source: Ministry of Environment

<sup>+</sup> Unit = mg/g dry weight \* Unit = μg/g dry weight

<sup>\*\*</sup> n=1

TABLE 11(c)

COMPARISON OF SERPENTINE RIVER SEDIMENT DATA+ WITH OTHER B. C. SITES

Constituent	Serpentine River (Table 11b)	Fraser River <sup>21</sup> (Main Arm)	Sediment Cores From B.C. Lakes <sup>2</sup>	Accumulated Stormwater Sediments - Urban <sup>2</sup> *
Cadmium	<0.001	0.0065- 0.007	0.001- 0.011	<0.001- 0.019
Cobalt	0.012- 0.017	0.016 - 0.026	0.010- 0.063	
Copper	0.012- 0.041	0.018 - 0.026	0.004- 0.468	0.022- 0.029
Iron	15.2 -29.0	12.8 -16.7	2.30 -121.00	11.0 -13.4
Lead	0.019- 0.098	0.004 - 0.008	0.010- 0.560	0.160- 0.209
Manganese	0.161- 0.520	0.308 - 0.328	0.016- 9.96	0.195- 0.210
Nickel	0.015- 0.045	0.035 - 0.050	0.005- 0.134	
Zinc	0.039- 0.139	0.028 - 0.047	0.009- 1.06	0,101- 0,183

+ Unit = mg/g dry weight

D. Stancil. Fraser River Estuary Study, Water Quality. Aquatic Biota and Sediments. December

t 10

C. McKean. Ministry of Environment and Parks (unpublished)
L. Swain. Stormwater Monitoring of a Residential Catchment Area: Vancouver, B. C. Water
Management Branch, Victoria, B. C. Internal Report. June 1983.

TABLE 12 WATER QUALITY DATA SUMMARY HYLAND CREEK (SITE 9)

Constituent	Period	No. of		Values*	
	of Record	Values	Maximum	Minimum	Mean
CARBON: Organic Inorganic Coliforms: fecal Colour:True Chloride: dissolved Metals: (total) Copper Iron Lead Zinc Nitrogen:Ammonia Nitrate Nitrite Oil and Grease Oxygen: BOD <sub>5</sub> COD dissolved pH Phosphorus:dissolved Ortho-Diss. total	1981 1981 1981 1981 1981 1981 1981 1981	Values  2  4  1  1  1  1  4  3  3  1  4  4  3  2	Maximum  7 18 3500 40 15.9  0.002 1 0.006 0.005 0.058 0.79 0.011 2.1 < 10 22 10.8 8 0.045 0.034 0.071 1.9	Minimum  6 18 2800 20 0.016 0.57 0.007 - 12.2 8.5 7.4 0.038 0.029 0.069	Mean  6.5 18 3500 + 30 0.031 0.69 0.009 - 17.8 9.4 7.6+ 0.043 0.03 0.070
Potassium:dissolved Sodium:dissolved Solids:total Suspended Silica:Reactive Sulphate Dissolved Specific Conductivity Tannin/Lignin Temperature Turbidity	1981 1981 1981 1981 1981 1981 1981 1981	1 1 3 3 1 2 4 4 4	13.2 156 18 14.9 6.7 193 1 15.9	- 132 3 - 6.7 181 0.7	- 144 9.3 - 6.7 188 0.8 14.5

<sup>\*</sup> All values are as mg/L except:

<sup>(1)</sup> Coliforms as MPN/100 mL (2) Colour as colour units (3) pH

<sup>(4)</sup> Temperature as °C (5) Specific conductivity as  $\mu$ S/cm (6) Turbidity as NTU + Median value

Data Source: Ministry of Environment and Parks

TABLE 13
WATER QUALITY DATA SUMMARY
MAHOOD (BEAR) CREEK

Chromium			<del></del>	Site 11	00056	and the second section of the s
Alkalinity: total	Constituent	-	1	l	Values*	
Arsenic   1974-1979   2   0.006   <0.005   Color   Carbon: Organic   1974-1979   21   18   1   6.1   6.1   Chloride: Dissolved   1974-1979   29   13.2   3.7   6.7   Coliforms: total   1979   1   920   -   -   Colour:True   1976-1979   18   64.7   23.7   48.9   calcium   1974-1979   38   16.9   6.2   12.2   3.7   48.9   calcium   1974-1979   38   16.9   6.2   12.2   3.7   48.9   calcium   1974-1979   38   5.7   1.4   3.7   3.7   48.9   calcium   1974-1979   32   0.006   <0.005   <0.0005   cadmium   1974-1979   32   0.006   <0.005   cadmium   1974-1979   32   0.006   cadmium   1974-1979   32   0.007   cadmium   1974-1979   32   0.55   1.47   cadmium   1974-1979   32   0.55   0.5   1.47   cadmium   1974-1979   32   0.12   cadmium   1974-1979   32   0.12   cadmium   1974-1979   33   0.12   cadmium   1974-1979   34   0.117   0.009   0.051   cadmium   1974-1979   23   0.12   cadmium   1974-1979   39   0.12   cadmium   1974-1979   39   0.12   cadmium   1974-1979   39   1.35   0.08   0.42   cadmium   1974-1979   39   1.35   0.08   0.09   0.014   cadmium   1974-1979   39   0.273   0.024   0.064   cadmium   1974-1979   39   0.273   0.024   0.064   cadmium   1974-1979   29   10.9   3.5   6.43   cadmium   1974-1979   20   10				1	Minimum	Mean
Carbon: Organic Chloride: Dissolved   1974-1979   21	Arsenic	1974-1979	2	0.006		43.3 -
Chloride: Dissolved   1974-1979   29	1		,	3	<u>ł</u>	7
Coliforms: total   1979	-		1	1		1
Colour:True   1976-1979   15	f .	1	1		3.1	0./
Hardness:total 1976-1979 18 64.7 23.7 48.9 12.2 magnesium 1974-1979 38 5.7 1.4 3.7 Metals: (total) Cadmium 1974-1979 16 <0.0005			1		10	21
calcium magnesium         1974-1979 1974-1979         38         16.9 5.7         6.2 3.7           Metals: (total)         1974-1979 16         <0.0005	I .			l .	1	
Metals: (total)         1974-1979         38         5.7         1.4         3.7           Metals: (total)         1974-1979         16         <0.0005	Í					
Metals: (total)         Cadmium         1974-1979         16         <0.0005         <0.0005         <0.0005           Chromium         1974         2         0.006         <0.005	1					r .
Chromium 1974 2 0.006	· · · · · · · · · · · · · · · · · · ·					J. 1
Chromium	Cadmium	1974-1979	16	<0.0005	<0.0005	<0.0005
Iron	Chromium	1974	2	0.006	ł	_
Lead Manganese 1974,1979 15 0.18 0.001 0.006 Manganese 1974-1979 15 0.18 0.002 0.059 Mercury 1974-1979 17 0.00006 0.00005 0.000 0.000 0.014 0.005 0.015 0.014 0.005 0.01	Copper		9	0.007	<0.001	0.003
Manganese         1974-1979         15         0.18         <0.02         0.059           Mercury         1974-1979         17         0.00006         <0.00005	Iron	1974-1979	32	9.5	0.5	1.47
Mercury         1974-1979         17         0.00006         <0.00005         0.000           Zinc         1974-1979         23         0.12         <0.005	Lead	1974,1979	23		<0.001	0.006
Zinc	Manganese	1974-1979		0.18	<0.02	0.059
Nitrogen: Ammonia	1 - 1			0.00006	<0.00005	0.00005
Nitrate/Nitrite 1974-1979 39 3.45 0.51 0.58 Organic 1978-1979 5 0.98 0.25 0.58 Organic Njeldahl 1974-1979 39 1.35 0.08 0.42 Oxygen: dissolved 1974-1979 29 13.4 7.5 10.3 97.4 Phosphorus:ortho 1974-1979 39 8.1 6.9 7.6+ Phosphorus:ortho 1974-1979 39 0.273 0.024 0.064 Oxygen: dissolved 1974-1979 39 0.273 0.024 0.064 Oxygen: dissolved 1974-1979 39 0.273 0.024 0.064 Oxygen: dissolved 1974-1979 29 10.9 3.5 6.43 Oxygen: dissolved 1974-1979 29 10.9 3.5 6.43 Oxygen: dissolved 1974-1979 29 10.9 3.5 0.43 Oxygen: dissolved 1974-1979 29 76 71 73.5	!				<0.005	0.014
Organic	-				0.009	0.051
Kjeldahl       1974-1979       39       1.35       0.08       0.42         Total       1976-1979       10       4.32       1.13       1.82         Oxygen: dissolved dissolved pH       1974-1979       29       13.4       7.5       10.3         pH       1974-1979       39       8.1       6.9       7.6+         Phosphorus:ortho total dissolved total       1974-1979       31       0.36       0.009       0.019         total total       1974-1979       32       0.049       0.014       0.027         total potassium dissolved       1974-1979       29       3.3       0.9       1.71         Sodium dissolved Solids: Total Dissolved       1974-1979       29       10.9       3.5       6.43         Dissolved       1974-1979       2       76       71       73.5	•				į	
Total 1976-1979 10 4.32 1.13 1.82   Oxygen: dissolved 1974-1979 29 13.4 7.5 10.3   F saturation 1974-1979 29 135.2 76.3 97.4   Phosphorus:ortho 1974-1979 31 0.36 0.009 0.019   total dissolved 1974-1979 32 0.049 0.014 0.027   total 1974-1979 39 0.273 0.024 0.064   Potassium dissolved 1974-1979 29 3.3 0.9 1.71   Sodium dissolved 1974-1979 29 10.9 3.5 6.43   Solids: Total 1974-1979 2 98 84 91   Dissolved 1974-1979 2 76 71 73.5						
Oxygen: dissolved 1974-1979 29 13.4 7.5 10.3 97.4 PH 1974-1979 39 8.1 6.9 7.6+ Phosphorus:ortho total dissolved 1974-1979 39 0.36 0.009 0.019 total 1974-1979 39 0.273 0.024 0.064 Potassium dissolved 1974-1979 29 3.3 0.9 1.71 Sodium dissolved 1974-1979 29 10.9 3.5 6.43 91 Dissolved 1974-1979 2 76 71 73.5	-				7	
# saturation 1974-1979 29 135.2 76.3 97.4 pH 1974-1979 39 8.1 6.9 7.6+ Phosphorus:ortho total dissolved 1974-1979 31 0.36 0.009 0.019 total dissolved 1974-1979 32 0.049 0.014 0.027 total 1974-1979 39 0.273 0.024 0.064 Potassium dissolved 1974-1979 29 3.3 0.9 1.71 Sodium dissolved 1974-1979 29 10.9 3.5 6.43 Solids: Total 1974-1979 2 98 84 91 Dissolved 1974-1979 2 76 71 73.5	•					
pH       1974-1979       39       8.1       6.9       7.6+         Phosphorus:ortho total dissolved total       1974-1979       31       0.36       0.009       0.019         total       1974-1979       32       0.049       0.014       0.027         total       1974-1979       39       0.273       0.024       0.064         Potassium dissolved       1974-1979       29       3.3       0.9       1.71         Sodium dissolved       1974-1979       29       10.9       3.5       6.43         Solids: Total       1974-1979       2       98       84       91         Dissolved       1974-1979       2       76       71       73.5						
Phosphorus:ortho total dissolved total dissolved total dissolved total prize dissolved total 1974-1979 1974 1979 1974-1979 197						
total dissolved 1974-1979 32 0.049 0.014 0.027 total 1974-1979 39 0.273 0.024 0.064 Potassium dissolved 1974-1979 29 3.3 0.9 1.71 Sodium dissolved 1974-1979 29 10.9 3.5 6.43 Solids: Total 1974-1979 2 98 84 91 Dissolved 1974-1979 2 76 71 73.5	l -				1	i l
total 1974-1979 39 0.273 0.024 0.064 Potassium dissolved 1974-1979 29 3.3 0.9 1.71 Sodium dissolved 1974-1979 29 10.9 3.5 6.43 Solids: Total 1974-1979 2 98 84 91 Dissolved 1974-1979 2 76 71 73.5	=				1	t i
Potassium dissolved       1974-1979       29       3.3       0.9       1.71         Sodium dissolved       1974-1979       29       10.9       3.5       6.43         Solids: Total       1974-1979       2       98       84       91         Dissolved       1974-1979       2       76       71       73.5	,				1	
Sodium dissolved       1974-1979       29       10.9       3.5       6.43         Solids: Total       1974-1979       2       98       84       91         Dissolved       1974-1979       2       76       71       73.5					(	
Solids: Total 1974-1979 2 98 84 91 73.5		•			•	
Dissolved 1974-1979 2 76 71 73.5						
	Suspended	1974-1979	3	22	14	
Suspended   1974-1979   3   22   4   13   13   13   14   13   13   14   13   13		•				
Temperature   1974-1979   35   17.5   3.7   11.5						
Turbidity 1974-1979 39 104 2.2 14.2	-					

<sup>\*</sup> All values are as mg/L except:

<sup>(1)</sup> Coliforms as MPN/100 mL (2) Colour as colour units (3) pH

<sup>(4)</sup> Temperature as °C (5) Specific Conductivity as  $\mu S/cm$  (6) Turbidity as NTU

<sup>(7) %</sup> saturation as %

<sup>+</sup> Median value

TABLE 13 Continued

		<del></del>	Site 0300	056	
Constituent	Sample	No. of		Values*	
	Period	Values	Maximum	Minimum	Mean
Alkalinity:total Arsenic Boron: dissolved Carbon: Organic Inorganic Coliforms:total fecal Chloride: dissolved Colour True Hardness:total calcium magnesium Metals: (total)	1972-1983 1982 1972 1981-1982 1981-1983 1972-1976 1973-1983 1972-1982 1972-1982 1972-1982 1972-1982	39 6 37 77 192 40 13 30 30 30	66.3 <0.25 <0.2 7 17 >2400 16000 151 40. 63.9 17.3 5.7	19.8 <0.005 <0.1 15 330 20 2.8 10 24.4 7.3	43.3 <0.1 <sup>+</sup> 3.1 <sup>4</sup> 16 2400 + 490 + 10.7 20.8 45.4 12.3 3.6
Aluminum Cadmium Chromium Copper Iron Lead Manganese Molybdenum Zinc Nitrogen: Ammonia Nitrate/Nitrite Nitrate Nitrite Organic Kjeldahl Oil and Grease Oxygen: BOD <sub>5</sub> COD Dissolved % Saturation PH Phosphorus: ortho-diss. Total dissolved Total Potassium: dissolved Phenol Sodium: dissolved Solids: total dissolved Suspended Silica: Reactive Sulphate Specific Conductivity Tannin/Lignin Temperature	1982-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1973-1983 1974-1983 1974-1983 1974-1983 1974-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983 1972-1983	55043636589963711066438164683757366 233333 3322433 13344233 3223 34 3	0.58 <0.01 0.013 0.02 111 <0.1 0.8 0.01 0.38 1.47 0.89 1.6 0.089 1.6 <10 16.7 13.4 131 8.0 0.056 0.183 3.4 0.006 89 190 338 129 17.9 12.6 609 190 338 129 17.9 12.6 609 190 318	0.06 <0.01 <0.005 <0.001 0.005 0.005 0.005 0.005 0.005 0.016 <0.005 0.15 <10.016 0.016 0.027 1.3 <0.02 71 68 1 8.9 <73 73 1.1	0.21 <0.01 <0.005+ 0.003+ 1.48 0.009+ 0.014 0.063 1.99 0.014 11.42 10.9 98.1 11.42 10.9 98.1 11.42 10.9 98.1 11.42 10.03 0.052 1.8 0.003 0.1 11.3 10.6 17 14.4 14.2 14.6 17 14.4 14.6 17 14.6 17 14.6 14.6 15 16 17 14.6 17 14.6 17 14.6 17 14.6 17 14.6 17 14.6 17 14.6 17 14.6 17 14.6 17 14.6 17 14.6 17 14.6 17 14.6 17 14.6 17 14.6 17 18 18 18 18 18 18 18 18 18 18

<sup>\*</sup> All values are as mg/L except

<sup>(1)</sup> Coliforms as MPN/100 mL (2) Colour as colour units (3) pH (4) Temperature as °C (5) Specific Conductivity as  $\mu S/cm$  (6) Turbidity as N.T.U.

<sup>(7) %</sup> saturation as %

<sup>+</sup> Median value

Data Source: Ministry of Environment and Parks

TABLE 14 WATER QUALITY DATA SUMMARY LATIMER CREEK SITE 8

Constituent	Sample Period	No. of Values	1	Values*	
	rentod	values	Maximum	Minimum	Mean
Carbon: Organic	1981	3	8	6	6.7
Inorganic	1981	3	14	13	13.7
Chloride: dissolved	1981	2	51.2	35.4	43.3
Coliforms: fecal	1981	5	5400	790	3500 +
Colour:True	1981	5	60	40	36
Metals: (total)					
Copper	1981	1	0.002	-	_
Iron	1981	1	1.2		-
Lead	1981	1	0.003	<del>-</del>	-
Zinc	1981	1	<0.005	-	-
Nitrogen: Ammonia	1981	5	0.161	0.06	0.087
Unionized Ammonia	1981	5	0.00069	0.00024	0.00049
Nitrate	1981	4	1.99	1.6	1.77
Nitrite	1981	4	0.027	0.016	0.02
Oil and Grease	1981	2	1.7	1.4	1.55
Oxygen: BOD <sub>5</sub>	1981	2	<10	<10.	<10
COD	1981	5	28	12.7	18.6
dissolved	1981	5	9.5	7.8	8.4
рH	1981	5	7.6	7.1	7.3 +
Phosphorus:Ortho-Diss.	1981	4	0.038	0.021	0.029
total dissolved	1981	4	0.051	0.037	0.043
total	1981	3	0.089	0.069	0.082
Potassium:dissolved	1981	2	2.4	2.2	2.3
Sodium:dissolved	1981	1 1	26.8	-	-
Solids:total	1981	4	200	160	178
Suspended	1981	] 4 [	14	5	9
Dissolved	1981	4	177	153	169
Silica: Reactive	1981	2	12.3	11.6	11.95
Specific Conductivity	1981	5	295	225	248
Sulphate dissolved	1981	3	5.8	5	5.27
Tannin/Lignin	1981	5	1.7	1.2	1.32
Temperature	1981	5	22.5	12.5	15.8
Turbidity	1981	1	7.3	-	

<sup>\*</sup> All values are as mg/L except:

<sup>(1)</sup> Coliforms as MPN/100 mL (2) pH (3) Colour as colour units

<sup>(4)</sup> Turbidity as N.T.U. (5) Specific Conductivity as  $\mu S/cm$ 

<sup>+</sup> Median value

	SITE+							
CONSTITUENT*	¥	æ	ပ	Q	Ŀì	E4	U	ж
Feca1	480 80	2400 1300	230	790	230	140	1300	330
Nitrogen:	0.045 0.59	0.061 0.018	0.023	0.023	0.64	0.009		90.0
Un-lonized Ammonia	0.0001 0.0021	0.0002 0.0002	0.0001	6000.0	0.011	0.0001		0.0001
Nitrate							·· ·	1.49 0.049
Nitrite	0 62 3	0.33 0.87		0.16	0.61	0.17	0.47	0.011
X toldahl	0.66	0.39 0.89	0.31	0.18	1.25	0.18	0.48	
Oxvgen:Dissolved	6.8 5.6	8.6 8.8		7.4	9.5	9.6	8.9	7.6
Ha	7.1 6.9	7.1 7.5		8.1	7.7	7.7	-T-	6.7
Phosphorus: Ortho-Diss	0.072 0.031	0.026 0.072		0.167	0.054	0.132	0.059	0.015
Specific Conductivity	181 429	191 214		210	172	172	124	162
Tannin & Lignin	1.8 2.2	1.1 0.9		0.4	0.7	9.0	6.0	6.0
Temperature	17.8 21.5	15.9 21.5		18.4	18.2	15.4	16.1	15.
Turbidity	5.5	5.1		5.5	<u>س</u>	8	6.4	5.3

\* All values are as mg/L except:
(1) pH (2) Temperature as °C (3) Specific conductivity as uS/cm
(4) Coliforms as MPN/100 mL (5) Turbidity as N.T.U.

Data Source: Ministry of Environment and Parks
+ Samples collected on 23 July 1981 appear in the first column for each site while samples collected on 11 August 1981 appear in the first column for sates A and B.

TABLE 15
ESCAPEMENT VALUES FOR THE NICOMEKL RIVER

YEAR	соно	STEELHEAD
1947	1500	
1948	1500	
1949	1500	
1950	1500	
1951	1500	75
1952	3500	
1953	3500	
1954	3500	
1955	1500	
1956	1500	75
1957	750	
1958	750	
1959	750	
1960	750	
1961	3500	
1962	200	
1963	200	
1964	7500	
1965	200	
1966	400	
1967	400	
1968	200	
1969	75	
1970	1500	
1971	3500	
1972	1500	
1973	3500	
1974	2500	
1975	3500	
1976	3500	
1977	3400	
1978	1500	
1979	1500	Ì
1980	1750	
1981	1000	
1982	1200	
1983	1000	

TIMING OF SPAWNING:

ARRIVE	Sep	tember
START	Early	November
PEAK	Late	November
END	Late	December

Data From: Hancock, M.J. and D.E. Marshall. 1985. Catalogue of Salmon Streams and Spawning Escapements of Statistical Area 29. Canadian Data Report of Fisheries and Aquatic Science No. 495. Department of Fisheries and Oceans.

TABLE 16 EFFLUENT DATA SUMMARY AMBASSADOR INDUSTRIES LTD. LANGLEY PE 5269

CONSTITUENT	PERIOD OF RECORD	NO. OF		UES* MINIMUM
		4	0.06	
Nitrogen:ammonia	1980	1	0.86	<del>-</del>
nitrate/nitrite	1980	1	5.9	-
nitrate	1980	1	5.89	-
nitrite	1980	1	0.005	-
Oxygen : BOD <sub>5</sub>	1980,1982	2	469	299
dissolved	1980	1	7.7	-
На	1980,1982	2	7.2	6.7+
Phosphorus:ortho-dissolved	1980	1	2.02	-
total	1980	1	2,10	-
   Solids:Total	1980-1982	2	826	650
Specific Conductivity	1980,1982	2	535	255
Temperature	1980	1	20.5	<b>-</b>

<sup>\*</sup> All values are as mg/L except:

<sup>(1)</sup> pH (2) Specific Conductivity as  $\mu S/cm$  (3) Temperature as °C

TABLE 17 EFFLUENT DATA SUMMARY WESTERN PAVING LIMITED (PE 2367)

Constituent	Period of Record	No. of Values	Maximum	Values* Minimum	Mean
		**************************************			
Coliform:fecal	1974	1	5	-	_
total	1974	1	21	-	
Flow	1974-1978	45	1 145.6	110.47	830.32
рН	1974-1978	8	8.3	6.7	7.55+
Specific					
Conductivity	1974-1978	9	168	90	132
Solids			١	-	3-
Total	1974-1978	13	64 183 "	3	11 599
Dissolved	1978	, J 1	90	J 	11 233
	1974-1978	10	-	7 h	-
Suspended	19/4-19/0	13	64 268	74	12 190

<sup>\*</sup> All values are as mg/L except (1) Coliform as MPN/100 mL

<sup>(2)</sup> pH (3) Specific Conductivity as  $\mu$ S/cm, (4) Flow as m³/d.

<sup>+</sup> Median

# TABLE 18 EFFLUENT DATA SUMMARY SURREY CO-OPERATIVE ASSOCIATION SITE PE 4262

					Values*	
Constituent		Sample Period	No. of Values	Maximum	Minimum	Mean
Flow Oil and Grease	1. 2. 3.	1983 1976-1978 1976-1979 1983-1984 1984	,	0.0065 168 14 90 26.6	- 2.9 1 0	49.3 6.3 50.3
pH	1. 3. 4.	1977 1983 1984	2 1 1	7.1 7 7.18	7 - -	7.05 - -
Suspended Solids	1. 3. 4.	1977 1983 1984	2 1 1	67 8 27	45	56 - -
Specific Conductivity	1.	1977	2	287	202	245

<sup>\*</sup> All values are as mg/L except:

(1) Flow in  $m^3/d$  (2) pH (3) Specific Conductivity as  $\mu S/cm$ 

#### Analyses by:

- 1. Ministry of Environment and Parks
- 2. Chemex Labs Ltd.
- 3. Preventive Maintenance Methods Ltd.
- 4. Analyses by Central Technical Services Ltd.

# TABLE 19 WATER QUALITY DATA SUMMARY NICOMEKL RIVER

SITE 1100002: NICOMEKL RIVER AT DAM

Constituent	Period of Record	No. of Values	Maximum	Values* Minimum	Mean
Alkalinity: total	1974-1979	39	104	22	76.6
Boron: dissolved	1974-1979	41	3.7	0.2	1.9
Chloride	1974-1979	53	14 800	208	9144
Colour:True	1976-1979	22	100	5	33
Hardness: total	1976-1979	29	4 280	60	2180
calcium	1974-1979	71	312	17	185.5
magnesium	1974-1979	69	991	9	592
Nitrogen:					
Ammonia	1976-1979	20	0.13	<0.01	<0.01+
Unionized Ammonia	1976-1977	7	0.0023	0.004	0.0012
Organic	1978-1979	9	2.0	0.3	0.48
Total	1974-1979	20	5	0.35	1.25
Nitrate/Nitrite	1974-1979	71	3.2	<0.02	0.476
Kjeldahl	1974-1979	71	2	<0.01	0.52
Oxygen: dissolved	1972-1979	52	22.5	2.5	9.6
% Saturation	1972-1979	52	253.1	28.5	99.6
рН	1974-1979	71	8.9	6.6	7.6+
, -		56	0.164	<0.003	0.054
total diss.	1974-1979	62	0.215	0.01	0.072
Potassium	1974-1979	58	304	8.5	187
Salinity	1974-1979	43	23 100	400	15 666
Sodium	1974-1979	49	8 100	123	5198
Solids:	_				
Suspended	1976-1979	5	27	9	15 -
dissolved	1976-1979	4	15 985	2859 1	4 330
Total	1976-1979	4	23 600	2886 1	5 846
Specific Conductivity	1974-1979	75	40 800	452 2	3 725
Sulphate	1974-1979	48	1 975		1 296
Temperature	1974-1979	59	25	3.9	15.1
Turbidity	1974-1979	71	39	1.9	7.5

<sup>+</sup> Median

<sup>\*</sup> All values are as mg/L except (1) colour as relative units ,

<sup>(2)</sup> pH (3) Termperature as °C,

<sup>(4)</sup> Specific Conductivity  $\mu$ S/cm, (5) Turbidity as N.T.U.,

<sup>(6)</sup> Salinity 0/00 (7) % Saturation as %

TABLE 19' Continued SITE 0300060: NICOMEKL RIVER KING GEORGE HIGHWAY

Constituent	Period of Record	No. of Values	V Maximum	alues* Minimum	Mean
	necor a	varues	TRAINGH		
Alkalinity:total	1972-1983	47	81.7	18	50.6
Arsenic	1975-1983	9	<0.25	<0.005	
Carbon:organic	1982-1983	12	18	2	7
inorganic	1982	8	21	12	18.5
Chloride	1972-1983	41	9800	10.3	762
Coliform fecal	1973-1983	38	9200	<2	95+
Colour:True	1972-1982	11	200	20	65.5
Cobalt	1982_	6	<0.1	<0.1	<0.1
Hardness:total	1972-1983	38	2342	32.1	251
calcium	1972-1982	32	155	7.9	27.9
magnesium	1972-1983	42	475	3.0	44.5
Metals:total		^	d 31.11	0.40	0 1111
Aluminum	1977-1983	8	1.44	0.18	0.44
Cadmium	1982-1983	7	<0.01	<0.01	<0.01
Chromium	1972-1983	31	0.016	<0.005	<0.005+
Copper	1972-1983	34	0.03	<0.001	0.006
Iron	1973-1983	36	3.4	0.14	1.27
Lead	1972-1983	35	<0.1	<0.001	0.003+
Manganese	1979-1983	12	0.16	0.02	0.08
Molybdenum	1982-1983	76	0.01	<0.01	<0.01
Nickel	1972-1983	27	<0.05	<0.01	<0.01+
Zine	1972-1983	38	0.04	<0.005	0.016
Nitrogen	1974-1983	31	4.85	0.24	1.54
nitrate/nitrite Kjeldahl	1974-1983	31 41	3	0.34	1.18
Ammonia	1974-1983	38	0.428	<0.005	0.127
Un-ionized Ammonia		38	0.0226	0.0002	
Nitrate	1976-1983	31	4.82	0.22	1.51
Nitrite	1972-1983	48	0.127	<0.005	0.027
Organic	1973-1983	39	2	0.33	0.94
Oxygen					
dissolved	1972-1983	46	17.9	1.3	10.2
COD	1982-1983	13	49	<10	26.8
% saturation	1972-1983	46	188.5	12.7	98.8
рН	1972-1983	44	9.3	6.5	7.2 +
Phenol	1982	4	0.002	<0.002	0.002
Phosphorus					
ortho diss.	1974-1983	45	0.324	<0.003	0.073
total diss.	1976-1983	31	0.249	0.018	0.085
Potassium:Dissolved	1972-1983	37	47	1.9	7.43
Silica	1982	2	0.154	0.052	-
Sodium:Dissolved	1972-1983	37	5500	8.7	276
Solids:					_
Suspended	1974-1983	39	60	3	18
Dissolved	1979-1982	12	688	98	268
Total	1972-1983	29	10700	118	1032
Specific Conductivity	1972-1983	45	21250	117	1373
Sulphate	1972-1983	36	323	10.5	61.5
Temperature	1972-1983	46	25	1.9	12.3
Turbidity	1972-1975	6	35	3.4	13.5
20. 22.42.7					

<sup>+</sup> Median

<sup>\*</sup> All values are as mg/L except (1) Coliform as MPN/100 mL (2) Colour as Colour units, (3) pH (4) Temperature as °C (5) Turbidity as N.T.U. (6) Specific Conductivity as  $\mu S/cm$ 

<sup>(7) %</sup> Saturation as %

TABLE 19 Continued

SITE 1100003 NICOMEKL RIVER AT HIGHWAY 99

Constituent	Period of Record	No. of Values	Maximum	Values* Minimum	Mean
Alkalinity:total	1974-1979	43	88.2	17.5	54.3
Boron	1974-1979	43	2.4	<0.1	0.75:0.3+
Chloride	1974-1979	54	10300	16	2375
Coliform: total	1979-1983	2	240	140	-
Colour:True	1976-1979	22	150	15	51
Hardness:					
total	1976-1979	29	3780	28.4	364
calcium	1974-1979	69	238	6.6	61.1
magnesium	1974-1979	69	775	2.9	167
Metals: (Total)				•	
Cadmium	1974-1975	27	0.0014	<0.0005	<0.0005+
Chromium	1974	6	0.009	<0.005	0.006
Copper	1974-1979	16	0.02	0.001	0.006
Iron	1974-1979	54	4.5	0.2	1.05
Lead	1974-1979	37	0.012	0.001	0.003
Manganese	1974-1976 1974-1975	28 28	0.17 0.0001	0.03	0.07 <0.00005
Mercury Zinc	1974-1979	20 38	0.0001	0.005	0.014
Nitrogen:	1717 1717	30	0.11	0.005	0.017
Ammonia	1976-1979	13	0.22	0.012	0.113
Un-ionized Ammonia		11	0.074	0.00007	
Organic	1978-1979	8	2	0.63	1.02
Nitrate/nitrite	1974-1979	70	3.9	<0.02	0.77
Kjeldahl Oxygen: dissolved	1974-1979 1974-1979	70 56	2.0 22.5	<0.01 0.8	0.88 11.0
% saturation	1974-1979	56 56	263.8		115.4
pH	1974-1979	70	9.2	6.5	7.75
Phosphorus	1214 1313	10	9.2	0.5	(•12
ortho-diss.	1974-1979	59	0.322	<0.003	0.055
total diss.	1974-1979	62	0.346	0.011	0.076
Potassium: Dissolved	1974-1979	56	213	2.1	53.7
Solids:	1717 1717	50	213	<b>4.</b> • 1	73+1
Suspended	1974-1979	31	68	3	16
Dissolved	1974-1979	16	7367		1420
Total	1974-1970	29	11000		1448
Sodium: Dissolved	1974-1979	51	5720		1407
Specific Conductivity		74	30000	_	7885
=	1974-1979	49			
Sulphate		_	1472		335.7
Temperature	1974-1979	60	24 n.6	3	15.6
Turbidity	1974-1979	68	46	1.2	9.5

<sup>+</sup> Median

<sup>\*</sup> All values are as mg/L except (1) Coliform as MPN/100 mL, (2) Colour as colour units, (3) pH (4) Specific Conductivity as µS/cm, (5) Temperature as °C, (6) Turbidity as N.T.U. (7) % saturation as % Data Source: Ministry of Environment and Parks.

TABLE 19 Continued

SITE 1100004: NICOMEKL RIVER AT 40 AVE

Constituent F	Period of Record	No. of Values	Maximum	Values* Minimum	Mean
Alkalinity	1980	5	70.7	43.9	57.6
Colour:True	1980	5	100	40	54
Hardness:total	1980	5	91.5	84.7	89
calcium	1980	5	20.5	18.6	19.5
magnesium	1980	5	10.0	9.3	9.8
Nitrogen:					
Ammonia	1980	5	0.31	0.044	0.14
Un-ionized Ammonia	a 1980	4	0.004	0.0008	
Organic	1980	5	2	0.95	1.25
Total	1980	5	3	1.86	
nitrate/nitrite	1980	5	1.42	0.58	1.03
Kjeldahl	1980	5	2	1.06	1.33
pH	1980	12	9.4	6.8	7.4+
Phosphorus:					
ortho-Diss.	1980	5	0.105	0.013	0.044
total	1980	5	0.31	0.092	0.196
Solids:					_
Suspended	1980	12	29	7	16
Dissolved	1980	12	274		216
Total	1980	12	300		232
Specific Conductivity	y 1980	12	408		334
Sulphate	1980	5	55.1	38.2	45.9
Temperature	1980	5	19	3.8	9.2

<sup>+</sup> Median

<sup>\*</sup> All values are as mg/L except (1) Colour as colour units, (2) pH

<sup>(3)</sup> Specific Conductivity as  $\mu$ S/cm, (4) Temperature as °C

TABLE 19 Continued SITE 1100005: NICOMEKL RIVER AT 168 ST

Constituents	Period of Record	No. of Values	V Maximum	alues* Minimum	Mean
Alkalinity:Total	1974-1980	6	72.5	52.3	62.5
Boron: dissolved	1974-1979	21	0.9	<0.1	0.2
Chloride	1974-1979	29	4850	10.1	412
Colour:True	1976-1980	17	150	20	47
Hardness:		·	-		·
total	1976-1980	20	108	38.1	67.6
calcium	1974-1980	41	118	7.4	19.8
magnesium	1974-1980	41	321	3.2	24.1
Metals (total):				_	
Cadmium	1974-1975	15	<0.0005	<0.0005	<0.0005
Chromium	1974	2	<0.005	<0.005	<0.005
Copper	1974	5	0.012	0.001	0.005
Iron	1974-1977	25	5	0.5	1.5
Lead	1974-1975	17	0.007	<0.001	<0.001+
Mercury	1974-1975	17	0.00003	<0.00005	<0.00005
Zine	1974-1975	17	0.03	<0.005	0.011
Nitrogen:					
Ammonia	1976-1980	16	0.352	0.027	0.141
Un-ionized Ammonia		16	0.0024	0.0002	0.0011
nitrate/nitrite	1974-1980	41	7.0	0.41	1.28
Kjeldahl	1974-1980	41	3.0	0.26	0.94
Oxygen:dissolved	1974-1979	28	16.5	5.4	10.3
% saturation	1974-1979	28	185.8	69.8	106.1
рH	1974-1980	49	8.4	6.4	7.3+
Phosphorus:					
ortho diss.	1974-1980	33	0.394	0.006	0.077
total diss.	1974-1980	36	0.437	0.019	0.096
Potassium:Dissolved	1974-1979	30	101	1.6	10
Sodium:Dissolved Solids:	1974-1979	27	2680	8.8	252
Suspended	1974-1980	28	81	4	15
Dissolved	1980	10	302	132	204
Total	1974-1980	26	9796	146	655
Specific Conductivity		51	14800	118	996
Sulphate	1974-1980	28	661	9.9	78.1
Temperature	1974-1980	44	23	3.3	14.3
Turbidity	1974-1980	42	56	2	12.3

<sup>\*</sup> All values are as mg/L except (1) Colour as colour units, (2) pH (3) Specific Conductivity as  $\mu S/cm$ , (4) Temperature as °C, (5) Turbidity as N.T.U. (6) % saturation as %.

TABLE 19 Continued

SITE 1100006: NICOMEKL RIVER AT 192 ST.

Constituent	Period of	No. of	Va	alues*	
	Record	Values	Maximum	Minimum	Mean
Alkalinity:Total	1974-1979	23	69.3	18.6	50.9
Boron: dissolved	1974-1979	22	<0.1	<0.1	<0.1
Chloride	1974-1979	30	32.3	6.3	20.2
Coliform-fecal	1979	1	130	-	-
-total	1979	1	540	-	-
Colour	1976-1979	15	150	15	40.7
Hardness:					
total	1976-1979	19	63.3	18.8	47
calcium	1974-1979	40	16	4.4	11.4
magnesium	1974-1979	40	6.1	1.9	4.3
Metals:Total	151				
Cadmium	1972	1	<0.0005	-	-
Copper	1974-1979	5	0.018	0.002	0.008
Iron	1974-1979	17	<b>4.</b> 4	0.6	1.5
Lead	1974-1979	7	0.01	<0.001	0.004
ł.	1974	2	0.00008		
Mercury Zinc	1974-1979	7	0.13	<0.005	0.025
	1314-1313	ł	<b>0.</b> 1.5	(0.003	
Nitrogen:	1976-1979	14	0.282	0.018	0.082
Ammonia		14	0.001	0.000	
Un-ionized Ammonia	1978-1979	5	1.0	0.29	0.60
Organic	1974-1979	12	4.0	1.1	1.84
Total		41	2.5	0.5	1.14
nitrate/nitrite	1974-1979	41	2.0	0.16	0.59
Kjeldahl	1974-1979		15.0	8.1	10.7
Oxygen: dissolved	1974-1979	31	134.0	74.0	101.1
% saturation		31 41	8.1	6.7	7.4+
pН	1974-1979	41	0.1	0.7	{ • ¬
Phosphorus:	40mh 40m0	22	0.204	0.029	0.063
ortho diss.	1974-1979	32	0.228	0.046	-
total	1974-1979	34	3	1.5	1.9
Potassium: Dissolved	1974-1979	31		6.3	19.2
Sodium:Dissolved	1974-1979	29	33.3	0.5	1344
Solids:	4.081/ 4.080	05	<b>-7</b>	2	13
Suspended	1974-1979	27	57	3	147
Dissolved	1978-1979	4	168	1 32 98	149
Total	1974-1979	27	176		188
Specific Conductivi		43	278	76	
Sulphate	1974-1979	26	14.7	<5 2 )	9.8
Temperature	1974-1979	35	16.5	3.4	11.5
Turbidity	1974-1979	41	46	1.5	8.1

<sup>+</sup> Median

<sup>\*</sup> All values are as mg/L except (1) Coliform as MPN/100 mL, (2) Colour as colour units, (3) pH (4) Specific Conductivity as  $\mu$ S/cm,

<sup>(5)</sup> Temperature as °C, (6) Turbidity as N.T.U. (7) % saturation as %.

TABLE 19 Continued SITE 0300061: NICOMEKL RIVER AT 192 STREET

Constituent	Period of Record	No. of Values	V Maximum	alues* Minimum	Mean
		, ar acc	TIGATING		
Alkalinity:Total	1972-1983	46	77.1	21.5	53.7
Arsenic	1975 <b>-</b> 1983	10	<0.25	<0.005	-
Carbon:organic	1982-1983	10	7	<1	1.8
inorganic	1982-1983	9	21	19	20
Chloride	1972-1983	43	50	8.2	23.2
Coliform:fecal	1973-1983	45	9200	50	790+
Colour:True	1972-1982	12	50	10	29
Hardness:total	1972-1983	41	80.5	22	50.5
calcium	1972-1983	41	21.5	5.1	12.4
magnesium	1972-1983	45	6.7	2.25	4.8
Metals:Total		•		_	
Aluminum	1982-1983	8	0.51	0.06	0.19
Cadmium	1982	7	<0.01	<0.01	<0.01
Chromium	1972-1983	25	0.012	<0.005	0.007
Cobalt	1982	7	<0.1	<0.1	<0.1
Copper	1972-1983	38	0.02	<0.001	0.003+
Iron	1973-1983	39	2.7	0.38	1.06
Lead	1972-1983	31	<0.1	<0.001	<0.001+
Manganese	1979-1983	13	0.27	0.05	0.09
Molybdenum	1982	7	<0.01	<0.01	<0.01
Nickel	1972-1983	30	<0.05	<0.01	<0.01+
Zinc	1972-1983	42	0.025	0.005	0.009
Nitrogen:					
Un-ionized Ammonia		37	0.0012		38 0.0004
nitrate/nitrite		30	2.55	0.97	1.56
Kjeldahl Ammonia	1972-1983 1974-1983	40 37	2.0 0.434	0.19 <0.005	0.64 0.090
Nitrate	1976-1983	30	2.53	0.95	1.55
Nitrite	1972-1983	47	0.04	<0.005	0.014
Organic	1973-1983	38	2.0	0.17	0.57
Oxygen:					
COD	1982-1983	10	22	<10	13.6
dissolved	1972-1983	43	13.3	7.5	10.5
% saturation	1972-1983	43	158.7	78.7	92.1
рН	1972-1983	43	8.4	7.0	7.4+
Phenol	1982	4	<0.002	<0.002	<0.002
Phosphorus:					
ortho diss.	1973-1983	43	0.099	0.016	0.050
total diss.	1976-1983	30	0.120	0.019	0.060
Potassium:Dissolved	1972-1983	38	3.7	1.5	2.3
Silica	1982	2	17.9	17.6	17.8
Sodium: Dissolved	1972-1983	35	38.2	11.1	21.5
Solids:		<del>-</del> -	<u>.</u> -		-
Suspended	1973-1983	35	54	3	1 4
Dissolved	1972-1983	24	180	$7\overset{\sim}{4}$	140
Total	1972-1983	28	190	102	152
Specific Conductivity		44	290	88	215
Sulphate	1972-1983	37	20	5.6	10.7
Temperature	1972-1982	13	23	3.5	9.9
Turbidity	1972-1975	7	17	3.7	10.6
Tur orarey	1714-1717	· · · · · · · · · · · · · · · · · · ·	11	J• I	10.0

<sup>+</sup> Median

<sup>\*</sup> All values are as mg/L except (1) Coliform as MPN/100 mL, (2) Colour as colour units, (3) Specific Conductivity as  $\mu$ S/cm, (4) Temperature as °C (5) Turbidity as N.T.U. (6) pH (7) % saturation as %

TABLE 19 Continued
SITE 0300062: NICOMEKL RIVER AT 64 AVENUE

Constituent	Period of Record	No. of Values	V Maximum	alues* Minimum	Mean
Alkalinity	1972-1983	50	65.8	22	47.1
Arsenic	1975-1983	10	<0.25	<0.005	-
Carbon:organic	1982-1983	10	7	2	3.9
inorganic	1982	. 5	15	1 4	14.2
Chloride	1972-1983	45	22.9	3	11.3
Coliform:fecal	1973-1983	47	>2400	207	920+
Colour:True	1972-1982	12	55	15	34.2
Hardness:total	1972-1982	43	67.4	24.9	46.3
calcium	1972-1982	43	15.6	5.9	10.9
magnesium	1972-1982	43	8.5	2.4	4.6
Metals:(Total)	_	_		0.41	0 00
Aluminum	1982	7	0.69	0.14	0.38 <0.01
Cadmium	1982-1983	8 26	<0.01 0.016	<0.01 <0.005	<0.01 <0.005 <sup>+</sup>
Chromium Cobalt	1972-1983 1982	20 7	<0.01	<0.005	<0.01
Copper	1972-1983	30	0.02	<0.001	0.002+
Iron	1973-1983	39	5.5	0.25	1.2
Lead	1972-1982	39	<0.1	<0.001	0.002+
Manganese	1972-1983	37	0.13	0.01	0.04
Molybdenum	1975-1983	9	0.02	0.0005	<0.01 <sup>+</sup> <0.01 <sup>+</sup>
Nickel	1972-1983	29	<0.05	<0.01 <0.005	<0.005
Zine	1972-1983	34	0.02	(0.005	(0.00)
Nitrogen: ammonia	1974-1983	40	0.525	<0.005	0.092
un-ionized ammonia		30	0.0017	0.00003	
nitrate/nitrite	1976-1983	33	3.1	1.29	2.26
nitrite	1972-1983	50	0.034	0.005	0.013
Kjeldahl	1972-1983	43	2.01	0.19	0.71
organic	1973-1983	4	2.0	0.14	0.62
Oxygen:	1002-1002	11	21	<10	16.3
COD dissolved	1982-1983 1972-1983	49	14.1	7.8	10.8
% Saturation	1972-1983	49	128.1	71.4	96.9
Phenol	1982	4	<0.002	<0.002	<0.002
pH	1972-1983	46	8.1	6.9	7.5+
Phosphorus:					0.000
ortho diss.	1973-1983	46	0.138	0.026	0.058
total diss.	1976-1983	32	0.175	0.044 0.9	0.082 1.8
Potassium	1972-1983	39 38	3.9 21	3.6	13.3
Sodium	1972-1983	38	۷ ا	J. U	ر • ر ،
Solids: Suspended	1974-1983	41	87	3	19
Dissolved	1972-1978	23	160	70	116
Total	1972-1983	32	226	76	138
Specific Conductivity		46	214	69	165
Sulphate	1972-1983	39	17.6	<5.0	7.5
Temperature	1972-1983	49	19	0.0	9.6
Turbidity	1972-1977	8	56	2.4	12.5
Turbidity	17(5 17(1				

<sup>+</sup> Median

<sup>\*</sup> All values are as mg/L except (1) Coliform as MPN/100 mL, (2) Colour as colour units, (3) pH, (4) Specific Conductivity as  $\mu$ S/cm,

<sup>(5)</sup> Temperature as °C, (6) % Saturation as % Data Source: Ministry of Environment and Parks

TABLE 19 Continued Site 1100007: NICOMEKL RIVER AT 228 STREET

Constituent	Period of	No. of		Values*	
	Record	Values	Maximum	Minimum	Mean
Alkalinity:Total	1977-1979	9	51.9	28.2	46.7
Boron:Dissolved	1978-1979	3	<0.1	<0.1	<0.1
Chloride	1978 1979	7	6.4	4.7	5.3
Coliform:fecal	1979	1	240		-
total	1979	11	920	-	-
Colour:True	1977-1979	5	80.0	20	49.5
Hardness:					
total	1977-1979	11	60	36.5	47.9
calcium	1977-1979	11	13.3	8.7	11.2
magnesium	1977-1979	11	6.5	3.6	4.8
Metals:Total				_	
Copper	1979	5	0.016	0.001	0.005
Iron	1979	5	1.80	0.4	0.88
Lead	1979	5	0.003	<0.001	0.0016
Zinc	1979	5	0.11	<0.005	0.027
Nitrogen:		_			•
Ammonia	1977-1979	7	0.185	0.018	0.068
Un-ionized Ammonia	1977-1979	6	0.0006	0.0002	
Organic	1978-1979	6	1.82	0.038	0.865
Total	1978-1979	10	4.38	1.51	2,58
nitrate/nitrite	1977-1979	11	2.44	0.91	1.92
Kjeldahl	1977-1979	8	2	0.35	0.74
Oxygen: dissolved	1977-1979	8	13	8.7	10.8
% saturation	1977-1979	8	129.7	91.2	105.5
Н	1977-1979	11	7.9	7.2	7.7+
Phosphorus:			,	• -	,
ortho diss.	1977-1979	7	0.141	0.041	0.061
total	1977-1979	7	0.191	0.047	0.082
Potassium	1978-1979	6	3.6	1.1	1.9
Sodium	1978-1979	7	8.3	6	7.4
Solids:		·	3	-	1 • •
Suspended	1978-1979	2	28	16	22
dissolved	1978-1979	2	110	86	98
Total	1978-1979	2	126	114	120
Specific Conductivity		8	145	112	135
Sulphate	1978-1979	6	8.3	<5	6.4
Temperature	1977-1979	10	18	5	13.4
Turbidity	1977-1979	11	19	2.2	7.2

<sup>+</sup> Median

<sup>\*</sup> All values are as mg/L except (1) Coliform as MPN/100 mL, (2) Colour as colour units, (3) pH, (4) Specific Conductivity as  $\mu S/cm$ , (5) Temperature as °C, (6) Turbidity as NTU, (7) % Saturation as %

TABLE 20
DATA SUMMARY SEDIMENT CHEMISTRY
NICOMEKL RIVER

		Λ	VALUES (mg/g	/g dry-weight	ght)			
Constituent	Site 1100007	Site 0300061	Site 1100006	Site 11 Maximum	1100005 <sup>+</sup> Minimum	Site 0300060	Site 1 Maximum	1100003 <sup>+</sup> Minimum
Argenio*	16	<25	16	17	15	43	10	<5
Boron*	. 1	<b>□</b>	ŧ	İ	1	\ \	ı	ţ
Calcium	2.82	2.39	2.94	4.23	2.46	9.4	4.24	2.74
Carbon-organic	1	<0.3	ţ	1	1	26	1	1
-inorganic	ł	1.7	!	1	ı	10	1	t
METALS:								
Aluminum	!	6.42	1	10.9	ı	12.9	7.58	ı
Barium	1	0.034	1	0.043	ı	0.043	0.038	ı
Beryllium*	1	₽	1	!	ı	▽	ı	1
Cadmium	<0.001	<0.001	<0.001	<0.001	<0.001	~	<0.001	<0.001
Chromium	0.023	0.018	0.023	0.36	0.022	0.052	0.035	0.021
Cobalt	ı	<0.01	1	0.022	İ		0.014	1
Copper	0.008	0.011	600.0	0.022	0.02		0.012	0.011
Iron	14	13.1	14.8	56.6	19.2		20.5	14.6
Lead	0.017	0.012	0.016	0.071	0.034	0.156	0.03	0.02
Magneslum	h9.t	3.52	5.09	7.22	5.17		6.18	5.11
Manganese	0.341	0.545	0.438	0.372	0.297		0.214	0.200
Molybdenum*	2	7	2	2	<b>~</b>	13	\ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \	\ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \
Nickel	0.016	0.02	0.019	0.028	0.027	0.057	0.022	0.021
Selenium*	•	<10	1	1 -	ı	را د :	! (	1
Strontium*	ı	10	1	<del> </del>   -	ı	. 44	<b>2</b> 3	1
Tellurium*	ı	<20	ł	1	1	025	1	!
Thallium*	1	<20	ţ	!	1	020	ī	ı
Tin*	1	<b>&lt;</b> 2	1	1	1	. بح	1 4	1
Titanium	ſ	0.334	ı	0.427	ı	0.223	0.311	ı
Vanadium*	ŧ	56	ı	28	1	39	12	
Zinc	0.045	0.035	0.041	0.057	0.055	0.133	0.048	0.042
Nitrogen:			,	<u>c</u>	30.1	c	1 21	0 83
- Kjeldahl	0.43	0.14	0.33	0.614	97.0	7 -	0.531	0.406
Phosphorus-total	0.63	300.0		•			}	
organic material	1	<b>▽</b>	- 1	i	1	10	1	ı
)	_							

Data Collected in 1979 \* Units = µg/g dry welght + n=2

Data Source: Ministry of Environment and Parks

#### TABLE 21 WATER QUALITY DATA SUMMARY ANDERSON CREEK

		Si	te 0300063 An	derson Creek		
Constituent	Sample Period	No. of Values	Values*			
	101100	Values	Maximum	Minimum	Mean	
Alkalinity:total	1972-1977	24	56	18.5	40.3	
Arsenic	1975	2	<0.005	<0.005	<0.005	
Boron: dissolved	1972	3	<0.2	<0.1	_	
Chloride: dissolved	1972-1977	24	6.3	2.8	4.9	
Coliforms:total	1973-1977	20	>2400	80	240 +	
fecal	1972-1977	20	1600	14	175 +	
Colour True	1972-1977	7	60	<5	29.3	
Hardness:total	1972-1977	24	58.5	24.2	47.1	
Calcium	1972-1977	23	15.5	6	12.4	
Magnesium	1972-1977	23	4.8	2.25	4.0	
Metals: total			•			
Chromium	1972-1977	12	0.007	0.005	<0.005+	
Copper	1972-1977	19	0.02	<0.001	<0.001+	
Iron	1974-1977	16	1.4	<0.1	0.44	
Lead	1972-1977	19	<0.003	<0.001	<0.001+	
Manganese	1972-1977	18	0.07	<0.02	0.05	
Molybdenum	1975	1 1	<0.0005	-	-	
Nickel	1972-1977	18	0.01	<0.01	0.01	
Zine	1972-1977	19	0.011	<0.005	<0.005+	
Nitrogen:Ammonia	1973-1974	6	0.11	<0.01	0.032	
Unionized Ammonia	1973-1977	17	0.0003	0.00005	0.0002	
Nitrate/Nitrite	1976-1977	7	2	1.46	1.68	
Nitrate	1972-1976	7	1.98	1.45	1.67	
Nitrite	1972-1978	24	0.015	<0.005	0.006	
Organic	1973-1976	14	1.1	0.05	0.34	
Kjeldahl	1972-1977	13	2.0	0.03	0.42	
total	1972	3	3.11	1.19	1.84	
Oxygen: dissolved	1972-1977	23	13.5	10.2	11.39	
% saturation	1972-1977	23	130.3	89.4	99.7	
рH	1972-1977	20	8	7	7.6 +	
Phosphorus: ortho-diss.	1976-1977	19	0.049	0.011	0.019	
total dissolved	1973-1977	7	0.063	0.012	0.026	
total	1972-1977	24	0.112	0.015	0.04	
Potassium:Dissolved	1972-1977	19	2.7	1.1	1.36	
Silica Reactive	1972	3	19.1	9.3	14.5	
Sodium:Dissolved	1972-1977	17	6.1	3.3	5.16	
Solids:total	1972-1977	8	132	80	108	
Suspended	1974-1977	16	42	1	7	
Dissolved	1972-1977	22	116	62	96.5	
Specific Conductivity	1972-1977	22	156	68	126	
Sulphate	1972-1977	19	14.1	5.2	8.6	
Tannin/Lignin	1972	2	0.5	0.5	0.5	
Temperature	1972-1977	24	12.3	3.5	8.2	
Turbidity	1976-1977	7	11	0.6	3.5	
					ر ٠ ر	

<sup>\*</sup> All values are as mg/L except

<sup>(1)</sup> Coliforms as MPN/100 mL (2) Colour as colour units (3) pH

<sup>(4)</sup> Temperature as °C (5) Specific Conductivity as  $\mu$ S/cm (6) Turbidity as N.T.U. (7) % Saturation as %

<sup>+</sup> Median value

TABLE 21 Continued

		Sit	e 1100017 And	erson Creek		
Constituent	Sample	No. of Values		Values*		
	Period	values	Maximum	Minimum	Mean	
Alkalinity:total Boron: dissolved Chloride Coliforms:fecal total Colour:True Hardness:total Calcium Magnesium Metals: total Copper Iron Lead Zinc Nitrogen:Ammonia Unionized Ammonia Nitrate/Nitrite Organic Kjeldahl total Oxygen: dissolved % saturation PH Phosphorus:ortho-Diss. total dissolved total Potassium:Dissolved Sodium:Dissolved Solids:total Suspended	1977-1979 1978-1979 1978-1979 1979 1977-1979	9 3 7 1 12 12 12 12 12 14 16 16 16 16 17 3 4	Maximum  56.9  <0.1 6.9 49 79 50 67 17.1 5.9  0.006 1.2 0.005 0.11 0.113 0.0003 3.7 1 1 5 17 159.3 8.1 0.049 0.067 0.093 2.7 7.1 134 15	Minimum  26.5 <0.1 5.3 - <5 43.6 11.2 3.8 <0.001 0.2 <0.001 <0.005 0.001 1.74 0.09 0.1 2.16 8.25 76.4 7.3 0.014 0.017 0.022 1.1 5.9 104 1	49.9 <0.1 6.01 - 15.42 56.1 14.9 4.6 0.002 0.41 0.002 0.019 0.025 0.00018 2.15 0.41 0.34 2.64 11.13 105.2 7.8 + 0.019 0.024 0.032 1.42 6.23 116.67 5.25	
dissolved Salinity Specific Conductivity Sulphate Tannin/Lignin Temperature Turbidity	1979 1977 1977-1979 1978-1979 1977 1977-1979 1977-1979	3 2 12 6 1 11 12	119 0 166 14.5 0.2 14 6.9	102 0 131 7 - 5.5 0.8	111 0 150.25 8.9 - 11.25 1.72	

<sup>\*</sup> All values are as mg/L except

<sup>(1)</sup> Coliforms as MPN/100 mL (2) Colour as colour units (3) pH

<sup>(4)</sup> Temperature as °C (5) Specific Conductivity as  $\mu$ S/cm (6) Turbidity as N.T.U.

<sup>(7) %</sup> Saturation as %

<sup>+</sup> Median value

#### 158 TABLE 22 WATER QUALITY DATA SUMMARY MURRAY CREEK

		Sit	ce 0300064 Mur	ray Creek	
Constituent	Sample Period	No. of Values			
	rentod	varues	Maximum	Minimum	Mean
Alkalinity:total	1972-1977	24	53.5	17	37.7
Arsenic	1975	2	0.005	<0.005	-
Boron : dissolved	1972	3	<0.2	<0.1	<del>-</del>
Chloride: dissolved	1972-1977	24	8.7	2	4.3
Coliforms: fecal	1973-1979	21	>2400	79	350 +
total	1972-1977	19	3500	110	1300 +
Colour True	1972-1977	7	50	20	33
Hardness:total	1972-1977	24	51	18	37
Calcium	1972-1977	24	12	4.2	8.7
Magnesium	1971-1977	23	5.1	1.8	3.7
Metals: total	4000 4000	1 ,,	0.013	<0.005	0.006
Chromium	1972-1977	11	0.013	<0.005	0.003
Copper	1972-1977	19 16	0.02 1.6	0.30	0.65
Iron	1974-1977	19	0.005	<0.001	0.002
Lead	1972-1977	18	0.005	0.01	0.002
Manganese	1972-1977 1975	1 1	0.007	0.007	0.052
Molybdenum Nickel	1972-1977	1 18	0.007	<0.01	<0.01
Zinc	1972-1977	19	0.009	<0.005	0.006
Nitrogen: Ammonia	1974-1977	16	0.161	0.014	0.055
Unionized Ammonia	1974-1977	14	0.0007	0.000083	0.0003
Nitrate/Nitrite	1972-1977	7	1.84	0.78	1.22
Nitrate	1972-1977	17	4.1	0.36	1.22
Nitrite	1972-1977	24	0.036	<0.005	0.009
Organic	1973-1977	14	1.01	0.15	0.43
Kjeldahl	1972-1977	18	1.117	0.12	0.44
total	1972	3	1.72	0.95	1.42
Oxygen: dissolved	1972-1977	23	13.3	9.9	11.68
% saturation	1972-1977	23	123.5	88.8	102.4
pH	1972-1973	20	8.3	7	7.5 +
Phosphorus:ortho-Diss.	1973-1977	19	0.118	0.015	0.037
total dissolved	1976-1977	7	0.152	0.027	0.059
total	1972-1977	24	0.201	0.026	0.068
Potassium:Dissolved	1972-1977	19	3.8	1.5	2.13
Silica:Reactive	1972	3	16.4	5.7	10,9
Sodium:Dissolved	1972-1977	17	9.1	3.2	6.14
Solids:total	1972-1977	8	110	64	93
Suspended	1974-1977	16	55	2	10
dissolved	1972-1977	22	124	46	85
Specific Conductivity	1972-1977	21	163	56	108
Sulphate	1974-1977	18	12.8	<5	5.8
Tannin/Lignin	1974	2	2.7	0.7	1.7
Temperature	1972-1977	24	14.5	1	8.3
Turbidity	1972-1975	7	12	2.6	7

<sup>\*</sup> All values are as mg/L except

<sup>(1)</sup> Coliforms as MPN/100 mL (2) Colour as colour units (3) pH

<sup>(4)</sup> Temperature as °C (5) Specific Conductivity as  $\mu$ S/cm (6) Turbidity as N.T.U.

<sup>(7) %</sup> Saturation as %

<sup>+</sup> Median value

TABLE 22 Continued

	11	00022 MURRA	Y CREEK AT	48 AVENUE	
	Period VALUES				
Constituent	of Record	NO. OF VALUES	MAXIMUM	MINIMUM	MEAN
Alkalinity:Total	1974-1979	23	54.5	16.5	41.4
Boron: dissolved	1974-1979	21	<0.1	<0.1	<0.1
Carbon:organic	1974-1979	22	23	<1	6.3
inorganic					
Chloride	1974 1979	30	19	2.3	4.4
Coliform:fecal	1979	1	70	-	-
total	1979	1	220		-
Colour (true)	1974-1979	15	150	15	38
Fluoride	1979	5	50	20	28
Hardness:					
Total	1976-1979	19	51.3	22.8	38
Calcium	1974-1979	39	11.8	4	9.0
Magnesium	1979	5	5.3	3.1	4.1
Metals:Total					
Cadmium	1974-1976	15	<0.0005	<0.0005	<0.0005
Copper	1974-1979		0.016	<0.001	0.003
Iron	1974-1979	,	8.6	0.3	1.3
Lead	1974-1979		0.004	<0.001	0.001
Zine	1974-1979	23	0.12	<0.005	0.012
Nitrogen:	11314 1313				
Ammonia	1972-1979	14	0.207	0.012	0.067
Un-ionized Ammonia	1976-1979	1	0.0007	0.0001	0.0003
	1974-1979	1	3	0.11	0.48
Kjeldahl Nitrate/nitrite	1974-1979		2.53	0.43	1.06
	1978-1979	ł .	1.0	0.21	0.45
Organic	1974-1979		3	0.87	1.48
Total	1974-1979	4	14.2	6.0	10.4
Oxygen: dissolved			120.8	62.8	96.6
: % Saturation	1974-1979		8.0	6.9	7.5+
pH	1974-1979	40	)	0.7	1 1.0
Phosphorus:	1071: 1070	21	0 227	0.015	0.038
ortho-diss.	1974-1979		0.237	0.015	0.036
total	1974-1979	<b>†</b>	0.572	4	1.8
Potassium:Dissolved	1979	3 5 3 4	2	1.7	123.6
Sodium:Dissolved	1979	2 2	134	105 88	
Solids:Total	1976-1979	] 3	106		97
Suspended	1976-1979		18	3	10
Sulphate	1974-1979		6.8	<5	<5 <sup>+</sup>
Temperature	1974-1979	•	16.5	3.2	11
Turbidity	1974-1979	40	88	1.1	7.5

<sup>\*</sup> All values are as mg/L except (1) Coliform as MPN/100 mL, (2) Colour as colour units, (3) pH, (4) Temperature as °C, (5) Turbidity as N.T.U. (6) % Saturation as %

TABLE 23

A COMPARISON OF SEDIMENT CONSTITUENTS FROM DIFFERENT LOCATIONS WITHIN B.C.

Constituent	Nicomeki River	Serpentine River	Fraser River <sup>21</sup> (Main Arm)	Sediment Cores	Accumulated Stormwater
					Setiments - Orban Catchment <sup>2</sup> *
Aluminum <sup>+</sup>	7.58 - 10.9				
Antimony <sup>+</sup>	0.012 - 0.017				
Arsenic*	<5.0 - 17.0				
Barium <sup>+</sup>	0.38 - 0.43				
Cadmium <sup>+</sup>	<0.001	<0.001	0.0065 - 0.007	0 001	
Calcium <sup>+</sup>	2.46 - 4.24				<0.001 - 0.019
Chromium <sup>+</sup>	0.021 - 0.36				
Cobalt+	0.014 - 0.022	0.012 - 0.017	0.016 - 0.026	0.00 - 0.063	
Copper <sup>+</sup>	0.008 - 0.022		0.018 - 0.026	600.0	
Iron <sup>+</sup>	14.00 - 26.6	1	12.8 - 16.7	00 101   00 0	0.022 - 0.029
Lead <sup>+</sup>	0.016 - 0.071	0.019 - 0.098	0.004 - 0.008	1	11 - 13.4
Magnesium <sup>+</sup>	4.64 - 7.22			000:0 - 010:0	0.160 - 0.209
Manganese <sup>+</sup>	0.209 - 0.438	0.161 - 0.520	0.308 - 0.328	0 016 - 0 06	L
Molybdenum*	<1.0 - 2.00			06.6	0.135 - 0.210
Nickel+	0.016 - 0.028	0.015 - 0.045	0.035 - 0.050	0 006 :: 0 13]	
Titanium +	0.311 - 0.427			0.003 = 0.134	
Stront1um*	14.00 - 23.0	· · · · · · ·			
Vanadium*	21.00 - 28.0				
Zinc <sup>+</sup>	0.041 - 0.057	0.039 - 0.139	0.028 - 0.047	0.009 - 1.06	0.101 - 0.183

+Unit = mg/g dry weight except as noted by \* \*Unit =  $\mu g/g$  dry weight

TABLE 24
WATER QUALITY DATA SUMMARY,
BOUNDARY BAY

		Site 03	300070:BOUNDA	RY BAY WEST		
Constituent	Period of Record	No. of Values	Values*			
	Record	varues	Maximum	Minimum	Mean	
Alkalinity: total	1980,1982	3	94.8	85.6	89.2	
Calcium	1982	1	260	-	-	
Chloride	1980,1982	3	13400	11500	12366	
Metals:(total)			:			
cadmium	1982	1 1	0.001	-	<b>-</b>	
copper	1980,1982	2	<0.001	<0.001	<0.001	
iron	1980,1982	2 2 2	0.017	<0.005	-	
lead	1980-1982		<0.001	<0.001	<0.001	
manganese	1980	1	0.002	-	· <del>-</del>	
mercury	1980,1982	3	0.00068	<0.00005	<0.00005	
nickel	1980,1982	3 2	<0.01	<0.01	<0.01	
zinc	1980,1982	2	0.005	<0.005	<del>-</del>	
Nitrogen:				·	/	
Ammonia	1982	2	0.03	0.013	0.021	
Unionized Ammonia	1974-1980	13	0.0036	0.000092	0.00072	
Nitrate	1980,1982	3	0.02	<0.02	<0.02	
Nitrite	1980,1982	3	0.006	<0.005	<0.005	
Organic	1980,1982	3 3 3 3 3 3 3	0.18	0.05	0.13	
Nitrate/Nitrite	1980,1982	3	0.03	<0.02	<0.02	
Kjeldahl	1980,1982	3	0.21	0.08	0.16	
Нq	1980,1982	3	8.3	7.9	8.2 +	
Phosphorus: ortho diss	1980,1982	3	0.034	0.006	0.017	
total diss.	1980,1982		0.038	0.014	0.024	
Sulphate:dissolved	1980	1 1	1562	-	-	
Solids: dissolved	1980,1982	2	22900	22200	22550	
suspended	1980,1982	2	5	5	5	
Specific Conductivity	1980,1982	3	39900	33300	35700	
Silica: Dissolved	1982	2	2.1	0.9	1.5	
Turbidity	1982	2	1.4	0.5	0.95	

<sup>\*</sup> All values are as mg/L except:

<sup>(1)</sup> pH (2) Specific Conductivity as  $\mu$ S/cm (3) Turbidity as N.T.U.

<sup>+</sup> Median value

TABLE 24 Continued

		Site 0	300071:BOUNDA	RY BAY EAST		
Constituent	Period of Record	No. of Values				
			Maximum	Minimum	Mean	
Alkalinity:Total	1980,1982	3	806	85.8	332.3	
Chloride	1980,1982	3	16400	10400	12830	
Coliform:fecal	1980,1982	2	5	<2	-	
Hardness: Calcium	1982	1	240	_	_	
Magnesium	1982	1	690	_	-	
Metals: (total)						
cadium	1982	1	0.0015		-	
chromium	1980,1982	3	0.016	0.005	0.009	
copper	1980,1982	3 3 2	0.002	<0.001	<0.001	
iron	1980,1982	3	0.076	0.018	0.048	
lead	1980,1982	2	0.001	<0.001	~	
mercury	1982	1	<0.00005	_	-	
nickel	1980,1982	3	<0.01	<0.01	<0.01	
zine	1980	1	0.01	-	-	
Nitrogen:		Ì				
Ammonia	1980,1982	3	0.042	<0.01	0.014	
Un-ionized Ammonia	1974-1980	13	0.0026	0.0002	0.00096	
Nitrate	1980.1982	3	0.17	<0.02	0.06	
Nitrite	1980,1982	3 3 3 3	0.079	<0.005	0.03	
Organic	1980,1982	3	0.3	0.03	0.2	
Nitrate/Nitrite	1980,1982	3	0.18	<0.02	0.06	
Oxygen:Dissolved	1980	1	10.4	-	-	
рH	1980,1982	3	8.3	8.0	8.0+	
Phosphorus: ortho diss	1980,1982	3	0.056	0.005	0.028	
total diss.	1980,1982	3	0.065	0.013	0.037	
Solids: Suspended	1980,1982	თ თ თ თ თ	9	5	3	
Specific Conductivity	1980,1982	3	44900	32100	36400	
Sulphate: dissolved	1980	1	1538	_		
Tannin & Lignin	1982	1	<0.01	~	_	
Temperature	1980	1	10.4	-	-	
Turbidity	1980,1982	2	2.3	0.7	1.5	

<sup>\*</sup> All values are as mg/L except:

<sup>(1)</sup> Coliform as MPN/100 mL (2) pH (3) Temperature as °C

<sup>(4)</sup> Turbidity as N.T.U.

<sup>+</sup> Median value

TABLE 25
BOUNDARY BAY CRAB FISHERY

YEAR	QUANTITY OF CATCH*	NUMBER OF FISHING DAYS
1980 1981 1982 1983	78 000 60 000 30 800 47 600	922 1288 627 491
1984	54 578	464

<sup>\*</sup> measured in kg

DATA SOURCE: Annual Statistical Summary of the Crab Fishery in Area 29 (Boundary Bay)

Department of Fisheries and Oceans,

Nanaimo, B.C.

TABLE 26
BOUNDARY BAY HERRING FISHERY

YEAR	NUMBER OF SPAWNS PER YEAR	AVERAGE SPAWN LENGTH *	AVERAGE DAYS OF SPAWN
1979	0	0	0
1980	1	914	59
1981	0	0	0
1982	0	0	0
1983	0	0	0
1984	1	914	59
1985	0	0	0

### \* MEASURED IN METRES

DATA SOURCE: FEDERAL FISHERIES: Annual Statistical Summary of Herring Spawning Data in Area 29 (Boundary Bay) Department of Fisheries and Oceans, Nanaimo, B. C.

TABLE 27
BACTERIOLOGICAL DATA SUMMARY:
BOUNDARY BAY

			JUNI	Ε	JULY		AUGUST		90th PERCENTILE
SITE	NO*	PERIOD	RANGE	MEDIAN	RANGE	MEDIAN	RANGE	MEDIAN	r ERCENTIDE
1		1982-1985	<3-240	43	7-1100	240	<3 <b>-</b> 93	15	438
1A		1979-1981	<3-460	23	<3->2400	33	9 <b>-</b> 1100	43	716
2		1982-1985	4-460	68	9->2400	33	<3-460	4	240
2A		1979-1981	9->2400	33	<3->2400	195	<3-1100	43	1100
3		1982-1985	4-240	43	9-460	93	4-240	23	240
3A		1979-1981	4-1100	33	<3-1100	98	<3-240	23	972
4		1982-1985	<3-1100	23	21->2400	122	<3->2400	7	1880
4 A		1979-1981	15-240	43	<3-1100	43	4->2400	10	460
5		1982-1985	4-93	23	7-460	33	7-93	23	93
5A		1979 <b>-</b> 1981	<3-460	43	<3-460	15	<3-240	23	240
6		1982-1985	<3-1100	43	15-240	23	<3-75	15	240
6A		1979-1981	<3->2400	23	4-240	23	<3-1100	26	350
7		1982-1985	<3-43	9	4-1100	43	7-240	7	43
7A		1979-1981	4-43	9	<3-1100	9	3-1100	11	82
8		1982-1985	<3 <b>-</b> 150	6	<3-460	15	4-1100	33	360
8A		1979-1982	<3 <b>-</b> 23	6	4-460	43	4-23	9	212
9		1982-1985	<3-1100	23	4-150	18	9-43	18	177
9A		1979-1981	<3 <b>-</b> 240	23	<3-460	9	<3 <b>-</b> 43	23	240
10		1982-1979	<3 <b>-</b> >2400	23	7->2400	93	4->2400	93	>2400
10A		1979-1981	<3 <b>-</b> 460	13	<3->2400	23	<3-93	15	460
11		1982-1985	4-150	7	9-93	23	<3-93	4	93
11A		1979-1981	7 <b>-</b> 93	23	<3 <b>-</b> 1100	43	4->2400	23	972

<sup>\*</sup> Sites identified by "A" are in close proximity to those with the same number identifier.

Data Source: Ministry of Health; Boundary Health Unit

TABLE 28

ESTIMATED NUTRIENT LOADINGS FROM DIFFUSE AGRICULTURAL SOURCES

NITROGEN VALUES kg/year

DISCHARGE TO	FINISHING CATTLE	COWS	CALVES	YEARLINGS	TOTAL
Sam Hill Cr.	3 400			11 970	15 370
Little Campbell R.	9 044	1 972	49		11 064
Nicomekl R.	5 100	5 304	141	160	10 704
Serpentine R.	345 100	5 644	163	42 972	393 879

## PHOSPHORUS VALUES kg/year (P)

DISCHARGE TO	FINISHING CATTLE	COWS	CALVES	YEARLINGS	TOTAL
Sam Hill Cr.	235			966	1 201
Little Campell R.	625	230	6		861
Nicomekl R.	352	618	17	13	1 000
Serpentine R.	23 842	657	20	3 471	28 001

TABLE 29

MAXIMUM CONCENTRATION OF TOTAL AMMONIA NITROGEN FOR PROTECTION OF AQUATIC LIFE (mg/L-N)

рН	Temp.	1.0	2.0	3.0	4.0	5.0	6.0	7.0	8.0	9.0	10.0
6.566.6.901.2.3.4.566.7.8.901.2.3.4.566.7.8.9.0.1.2.3.4.5.6.7.8.9.0.1.2.3.4.5.6.7.8.9.0.0.1.2.3.4.5.0.0.0.0.0.0.0.0.0.0.0.0.0.0.0.0.0.0	27.7 27.9 26.8 25.8 21.6 19.1 21.4 21.6 21.6 21.6 21.6 21.6 21.6 21.6 21.6	28.3 27.5 26.5 25.2 21.3 19.8 17.8 16.0 14.1 10.7 11.0 14.1 10.7 11.0 14.1 10.7 11.0 14.1 10.7 11.0 11.0 11.0 11.0 11.0 11.0 11	27.9 27.2 26.2 25.1 23.9 22.5 20.9 17.5 15.7 14.0 12.2 10.5 8.98 7.60 1.65 1.65 1.06 0.856 0.692	27.58 25.98 25.98 22.70 17.35 12.48 12.09 17.35 10.48 10.48 10.99 11.53 10.88 10.63 10.63 10.68	27.2 26.4 25.5 23.3 21.4 23.3 21.4 23.3 21.4 20.8 17.1 10.3 7.4 22.5 11.3 11.3 11.3 11.3 11.3 11.3 11.3 11	26.8 26.1 25.2 23.6 20.6 20.6 20.6 20.6 16.9 15.4 11.7 10.1 8.67 3.14 4.90 3.14 9.15 1.67 3.14 9.15 1.69 1.69 1.69 1.69 1.69 1.69 1.69 1.69	26.5 25.8 24.9 23.7 21.4 19.3 16.0 13.3 10.0 13.3 10.0 10.0 10.0 10.0 10	26.2 25.56.66.51.7 11.5.8 11.5.9 11.5	26.0 25.4 44.2 29.5 9.5 17.0 14.0 19.5 19.5 19.5 19.5 19.5 19.5 19.5 19.5	25.7 25.0 25.0 25.0 25.0 25.0 25.0 25.0 25.0	25.57 23.99 21.85 11.60 11.20 11.60 11.20 11.60
	11.0	12.0	13.0	14.0	15.0	16.0	17.0	18.0	19.0	20.0	
6.5 6.6 6.6 6.7 7.7 7.3 4.5 6.7 7.7 7.7 7.7 7.8 8.8 8.8 8.8 8.8 9.0	25.2 24.5 23.7 21.6 3.9 17.4 15.9 21.6 11.1 9.14.6 11.1 9.14.6 11.1 9.17.8 65.8 11.9 11.9 11.9 11.9 11.9 11.9 11.9 11	25.0 24.3 23.5 22.4 20.8 17.7 14.5 15.7 14.5 10.5 11.0 98.18 17.3 11.0 98.18 17.3 11.0 98.18 17.3 11.0 98.18 17.3 11.0 11.0 11.0 11.0 11.0 11.0 11.0 11	24.1 23.3 22.3 21.3 22.3 21.3 21.3 21.3 21	24.6 23.1 23.1 22.1 19.5 17.1 15.9 18.3 17.1 10.8 17.1 10.8 17.1 10.8 17.1 10.8 17.1 10.8 17.1 10.8 17.1 10.8 17.1 10.8 17.1 10.8 10.8 10.8 10.8 10.8 10.8 10.8 10	24.5 23.0 23.0 21.0 19.7 16.9 15.4 16.9 15.9 17.6 15.9 17.6 17.6 17.6 17.6 17.6 17.6 17.6 17.6	24.68 24.68 21.98 19.6 15.8 15.8 12.2 10.7 26.7 26.7 26.7 26.7 26.7 26.7 26.7 26	24.2 23.7 21.8 22.7 21.8 20.7 21.8 20.7 21.8 20.7 21.8 21.7 21.7 2.9 20.6 4.5 21.9 21.9 21.9 21.9 21.9 21.9 21.9 21.9	24.0 23.36 22.6 21.7 20.6 19.4 18.1 16.7 15.2 13.6 12.1 10.6 12.1 10.6 12.1 10.6 12.3 10.5 10.8 17.0 10.8 17.0 10.8 17.0 17.0 17.0 17.0 17.0 17.0 17.0 17.0	23.9 23.565316.6 19.3 18.6 15.1 10.6 15.1 10.6 15.4 10.6 15.4 10.6 15.4 10.6 15.4 10.6 15.4 10.6 10.8 10.8 10.8 10.8 10.8 10.8 10.8 10.8	23.8 23.4 21.5 22.4 21.5 20.4 21.5 20.5 17.9 16.5 15.1 12.0 10.5 10.5 10.5 10.5 10.5 10.5 10.5 10	

TABLE 30 AVERAGE 30-DAY CONCENTRATION OF TOTAL AMMONIA NITROGEN FOR PROTECTION OF AQUATIC LIFE (mg/L-N)

рН	Temp.	1.0	2.0	3.0	4.0	5.0	6.0	7.0	8.0	9.0	10.0
56789012345678901234567890 666667777777777788888888888888	2.08 2.08 2.08 2.08 2.08 2.08 2.08 2.08	2.055 2.	2.02 2.02 2.02 2.02 2.02 2.02 2.02 2.02	1.99 1.99 1.99 1.99 1.99 1.99 1.99 1.99	1.97 1.97 1.97 1.97 1.97 1.97 1.97 1.97	1.94 1.94 1.94 1.94 1.994 1.995 1.995 1.995 1.995 1.995 1.957 1.48 200.759 1.48 200.334 0.240 0.161 0.131	1.92 1.92 1.92 1.92 1.92 1.92 1.92 1.93 1.93 1.93 1.93 1.93 1.93 1.93 1.93	1.90 1.90 1.90 1.90 1.90 1.90 1.90 1.90	1.88 1.88 1.88 1.88 1.88 1.88 1.88 1.88	1.86 1.86 1.86 1.86 1.86 1.86 1.87 1.87 1.87 1.87 1.60 1.35 1.90 1.37 1.90 1.37 1.90 1.37 1.90 1.37 1.90 1.37 1.90 1.37 1.90 1.90 1.90 1.90 1.90 1.90 1.90 1.90	1.84 1.84 1.884 1.884 1.885 1.855 1.855 1.855 1.855 1.857 1.870 0.775 1.322 1.322 1.322 1.322 1.322 1.332 1.
	11.0	12.0	13.0	14.0	15.0	16.0	17.0	18.0	19.0	20.0	
6.6.78.9012345678901234567890 1.234567890123456.7890	1.82 1.83 1.83 1.83 1.83 1.83 1.83 1.83 1.83	1.81 1.81 1.81 1.81 1.81 1.81 1.81 1.82 1.82	1.80 1.80 1.80 1.80 1.80 1.80 1.80 1.80	1.78 1.78 1.78 1.78 1.79 1.79 1.79 1.79 1.80 1.80 1.80 1.87 0.764 0.241 0.241 0.134	1.77 1.77 1.77 1.77 1.77 1.77 1.77 1.78 1.78	1.64 1.64 1.64 1.65 1.65 1.65 1.65 1.65 1.66 1.66 1.42 1.02 0.65 1.02 0.42 1.02 0.42 1.02 0.42 0.42 0.15 0.15 0.15	1.52 1.52 1.52 1.53 1.53 1.53 1.53 1.53 1.53 1.54 1.54 1.54 1.54 1.54 1.54 1.54 1.54	1.41 1.41 1.42 1.42 1.42 1.42 1.42 1.42	1.31 1.31 1.32 1.32 1.32 1.32 1.32 1.32	1.22 1.22 1.22 1.22 1.22 1.23 1.23 1.23	

<sup>the average of the measured values must be less than the average of the corresponding individual values in Table 30.
each measured value is compared to the corresponding individual values in Table 30.</sup>